

1 **Mitigating life cycle GHG emissions of roads to be built through**  
2 **2030: case study of a Chinese province**

3  
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31 **Abstract**

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33 The expansion of road networks in emerging economies such as China causes  
34 significant greenhouse gas (GHG) emissions. This development is conflicting with  
35 China's commitment to achieve carbon neutrality. Thus, there is a need to better  
36 understand life cycle emissions of road infrastructure and opportunities to mitigate  
37 these emissions. Existing impact studies of roads in developing countries do not address  
38 recycled materials, improved pavement maintenance, or pavement-vehicle interaction  
39 and electric vehicle (EV) adoption. Combining firsthand information from Chinese  
40 road construction engineers with publicly available data, this paper estimates a  
41 comprehensive account of GHG emissions of the road pavement network to be  
42 constructed in the next ten years in the Shandong province in Northern China. Further,  
43 we estimate the potential of GHG emission reductions achievable under three scenario  
44 sets: maintenance optimization, alternative pavement material replacement, and EV  
45 adoption. Results show that the life cycle GHG emissions of highways and Class 1-4  
46 roads to be constructed in the next 10 years amount to 147 Mt CO<sub>2</sub>-eq. Considering the  
47 use phase in our model reveals that it is the dominant stage in terms of emissions,  
48 largely due to pavement-vehicle interaction. Vehicle electrification can only

49 moderately mitigate these emissions. Other stages, such as materials production and  
50 road maintenance and rehabilitation, contribute substantially to GHG emissions as well,  
51 highlighting the importance of optimizing the management of these stages. Surprisingly,  
52 longer, not shorter maintenance intervals, yield significant emission reductions.  
53 Another counter-intuitive finding is that thicker and more material-intensive pavement  
54 surfaces cause lower emissions overall. Taken together, optimal maintenance and  
55 rehabilitation schedules, alternative material use, and vehicle electrification provide  
56 GHG reduction potentials of 11%, 4%-16% and 2%-6%, respectively.

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## 58 1. Introduction

59 Roads are one of the dominant forms of transportation infrastructure globally, along  
60 with rail lines, shipping ports, and airports. Roads can accelerate economic  
61 development and improve quality of life by providing connections between regions,  
62 and thus enabling the circulation of products, technologies, and knowledge (Banister,  
63 2012; Yu and Lu, 2012; Yu et al., 2012). However, road construction leads to  
64 greenhouse gas (GHG) emissions both because pavement construction causes high  
65 emissions and because the increased availability of roads leads to induced traffic which  
66 is predominantly carried out with polluting fossil-fuel powered vehicles (Yu and Lu,  
67 2012). Road infrastructure is still unevenly distributed at a global level with limited  
68 road access in low-income countries (Meijer et al., 2018; Weiss et al., 2018). An  
69 estimated 800,000 km of roads will be needed to provide the vast majority of the global  
70 population with quasi-universal road access (Wenz et al., 2020). In order to achieve  
71 this, an estimated 2 Gt of CO<sub>2</sub> would be emitted, representing around 0.25% of the  
72 carbon budget still available for compliance with the 2°C target (Wenz et al., 2020).

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74 China's scale of infrastructure construction is unprecedented among all developing  
75 countries. The expected pavement construction, upgrade and expansion in the coming  
76 decades will make it more difficult for China to meet its goal of carbon neutrality.  
77 According to forecasts by the China Highway Construction Industry Association and  
78 the Ministry of Transport, the expected total investment for road construction in 2021-  
79 2030 is around \$2.94 trillion. This encompasses 3,365,000 km of pavement of all  
80 pavement levels (Yicai, 2018). Further, maintenance and rehabilitation (M&R) of the  
81 existing short-lived network of pavements adds to the impact of new construction  
82 (Wang, 2013). To achieve the recently proposed target of carbon neutrality (Wang and  
83 Zhang, 2020), an environmentally preferable system of road construction and  
84 management is needed.

85

86 China is currently piloting and promoting alternative materials and engineering  
87 methods in order to improve the quality and energy efficiency of pavement construction.  
88 In addition to commonly promoted materials such as recycled asphalt and pavement,  
89 China is also experimenting with the use of recycled materials for in-situ hot or cold  
90 recycling, in-plant cold recycling, and full-depth cold recycling (Ministry of Transport,  
91 2019). However, these materials and technologies are not currently implemented at a  
92 large scale in the country, and their life cycle environmental impact and the  
93 corresponding potential for GHG mitigation remain unclear. Moreover, current road  
94 construction and M&R decisions rarely consider the full life cycle impact of roads  
95 reasonably and comprehensively.

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97 Established pavement LCA models for high-income countries are inadequate for China  
98 or other developing countries. Pavement characteristics depend on countries' individual

99 design and construction standards and practices, which change with economic  
100 development (Liu et al., 2020). For example, China's pavement has been found to have  
101 a notably shorter design life compared to developed countries, which is caused by a  
102 combination of engineering and social factors such as overloaded traffic (Gao and Cao,  
103 2001; Huang, 2000), insufficient structures such as thin pavement layers (Liu et al.,  
104 2020), and poor construction quality.

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106 This paper thus estimates the GHG emissions of the pavements to be built in the next  
107 ten years 2021-2030 in a selected province of China: Shandong province. Firsthand  
108 data from Chinese road engineers and publicly available data were combined for the  
109 research.

### 111 1.1 Pavement life cycle assessment

112 Life cycle environmental assessment of pavement quantifying the impacts of (1)  
113 material production, (2) transportation, (3) construction, (4) use, (5) M&R, and (6) end-  
114 of-life, is well established in the literature (Jiang and Wu, 2019). The product system  
115 'pavement' is defined in such a way that the use-phase addresses additional energy use  
116 and emissions of vehicles caused by sub-optimal road conditions, e.g. roughness and  
117 deflection on the road surface. There are only a limited number of LCA studies  
118 addressing pavement in China and these studies are limited in scope. For example, Chen  
119 et al. (2017) find that emissions of new road construction vary widely between 10.6 and  
120 823 t CO<sub>2</sub>-eq/km<sup>2</sup> among China's provinces but limit their analysis to the construction  
121 phase only. Guo et al. (2014) also include the maintenance stage using a simplified  
122 approach. They conclude that the production and M&R stages emit 1.85 kt CO<sub>2</sub>-eq/km  
123 and 1.76 kt CO<sub>2</sub>-eq/km, respectively. Guo et al. (2017) find that the production stage  
124 contributes most to the life cycle environmental impacts but do not consider M&R and  
125 use stages appropriately. In addition, none of the mentioned studies use primary data  
126 from Chinese road construction engineers.

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128 The use phase can significantly change the life cycle emissions of pavements,  
129 especially due to pavement-vehicle interaction (PVI). PVI focuses on the excess energy  
130 consumption of vehicles caused by road roughness, texture, and deflection, and can  
131 account for a significant percentage of the total vehicle energy consumption (Beuving,  
132 2004). Two of the major models of PVI are roughness-induced PVI and deflection-  
133 induced PVI (Louhghalam et al., 2017; Louhghalam et al., 2014; Louhghalam et al.,  
134 2015; Pouget et al., 2012). Deflection-induced PVI is primarily influenced by pavement  
135 structure with stiffer pavements having less deflection, while roughness-induced PVI  
136 can significantly increase with time due to damages to the road surface. In existing  
137 studies, the PVI emissions often accounted for more than 50% of the total life cycle  
138 GHG emissions of pavements (Araújo et al., 2014; Noshadravan et al., 2013; Xu et al.,  
139 2019; Yu and Lu, 2012; Zhang et al., 2010). However, the majority of previous studies  
140 on PVI did not consider the mitigation potential from vehicle electrification.

### 142 1.2 GHG emission reduction strategies

143 Previous research has investigated improvement strategies across all life cycle stages.  
144 Here we adopt three of these. M&R equipment and schedules depend on pavement  
145 structure (Yu and Lu, 2012; Yu et al., 2013), traffic volume (Wang et al., 2012), and  
146 soil conditions (Thenoux et al., 2007) among others and significantly impact the  
147 magnitude and variability of pavement life cycle emissions. Optimization of M&R can  
148 therefore contribute to significant life cycle emission reductions (alongside cost

149 reductions) (Chan et al., 2008; Labi and Sinha Kumares, 2005; Mandapaka et al., 2012;  
150 Praticò et al., 2011; Zhang et al., 2010).

151

152 The use of novel materials for road engineering shows potential for reduction of  
153 environmental impacts (and cost), with RAP, fly ash and polymer identified as the most  
154 popular new materials (Balaguera et al., 2018). Some of the novel materials, such as  
155 RAP and recycled concrete aggregates (RCA), have a significantly lower  
156 environmental impact compared to hot mix asphalt (HMA) (Chiu et al., 2008; Kucukvar  
157 and Tatari, 2012; Vidal et al., 2013) in addition to a cost advantage (Ponte, 2016). Other  
158 materials, such as glass (Chiu et al., 2008; Huang et al., 2009), ash from the incineration  
159 of municipal solid waste (Olsson et al., 2006), and industrial waste (Schwab et al., 2014)  
160 do not show a significant advantage over traditional HMA pavement due to leachate  
161 and a reduced life cycle performance of the resulting roads. The potentially reduced  
162 performance of pavements, which can lead to more use-phase vehicle emissions, have  
163 been studied for multiple materials (Araújo et al., 2014) and production methods such  
164 as warm mix asphalt (WMA) (Anthonissen et al., 2015). However, the available studies  
165 on new paving materials focus on roads in North America and Europe. The  
166 environmental, social, and economic implications discussed in these studies are  
167 therefore not applicable to China or other developing countries (Balaguera et al., 2018).

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169 In summary, existing studies have demonstrated that pavement GHG emissions can be  
170 quantified for specific contexts. However, the complete life cycle pavement GHG  
171 emissions including the stages of use or M&R have not been quantified for China. In  
172 addition, the life cycle mitigation potential of alternative materials, optimized M&R  
173 schedule, and EV adoption have not been captured in this context. Finally, we are not  
174 aware of existing studies using primary data from Chinese road construction engineers.

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### 176 1.3 Study objective

177 Hence, the objectives of the current study are to develop a representative LCA model  
178 for pavements in one of China's Northern provinces utilizing firsthand data, and to  
179 explore the GHG emission reduction possibilities for these pavements. The sets of  
180 scenarios explored are 1) M&R schedule, 2) alternative low carbon materials, 3) and  
181 EV adoption. The knowledge gained can be used to provide environmental information  
182 and insights for policy makers and practitioners in decisions regarding pavement  
183 structure and M&R.

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## 185 2. Methodology

### 186 2.1. Goal and scope

187 The goal of the study is to assess the life cycle global warming potential of five classes  
188 of asphalt pavement in the Shandong Province as shown in Table 1. The Shandong  
189 Province is selected to represent the typical pavement situation in Northern China. The  
190 functional unit is one lane-km (1 km) over the entire pavement lifetime. The scope of  
191 this study includes the stages of material production, transportation, construction, use,  
192 and M&R. The projected length of the newly constructed road network pavement is  
193 reported in Table S1, while the system boundary of this analysis is shown in Figure 1.

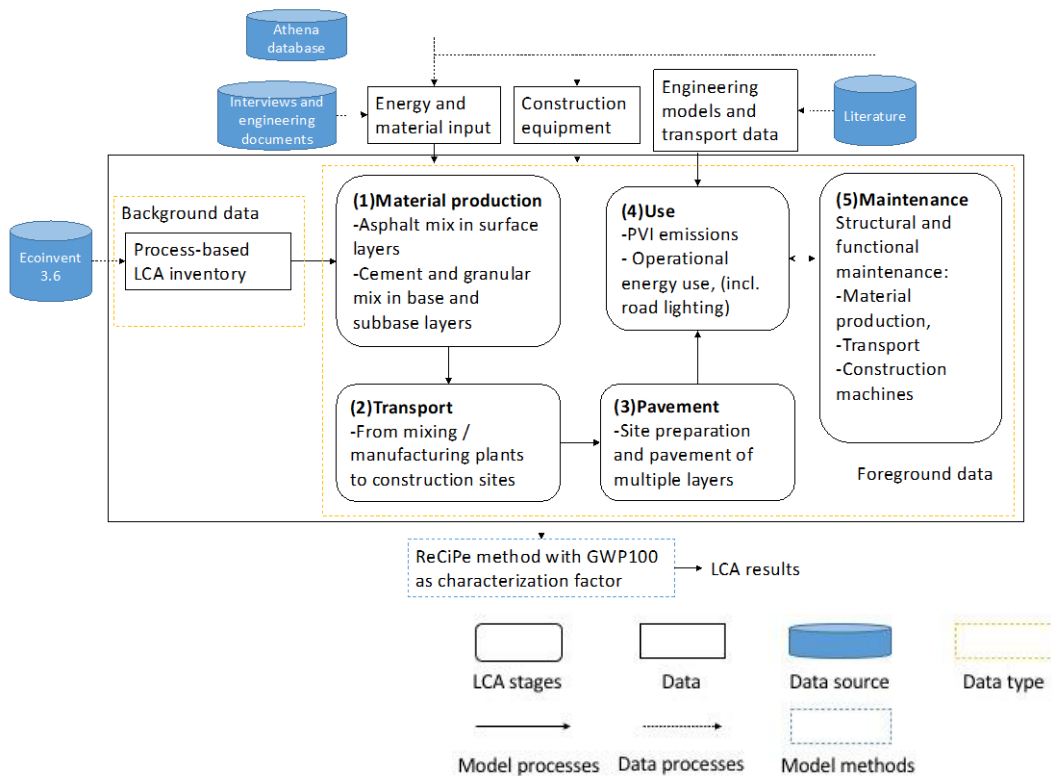
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196 *Table 1 The five technical classes of asphalt pavement analyzed in this study*

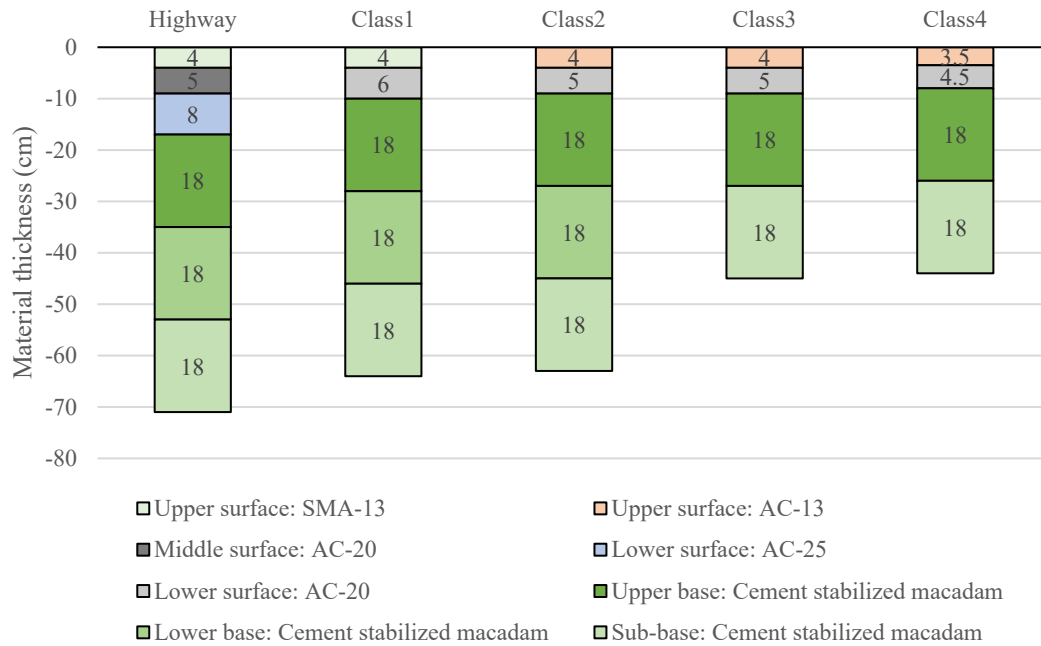
	Highway	Class I	Class II	Class III	Class IV
Speed (km/h)	80-120	60-100	60-80	30-40	20-30
Average daily traffic (PCU, passenger car unit)	>15,000	>15,000	5,000-15,000	2,000-6,000	<2,000
Number of lanes	≥4	≥4	2	2	2
Life span assumed (years)	20	20	15	15	15
Percentage of heavy trucks	14%	14%	7%	7%	7%

197 Source: National standards JTG F10-2006 and JTG B01-2014  
 198



202 *Figure 1. Data input and system boundary of pavement LCA.*

203 Data for pavement structure and materials is collected from national standards (Li and  
 204 Li, 2006; Ministry of Transport, 2004, 2006a, b, 2008, 2014, 2015; Yao, 2006),  
 205 engineering papers, and interviews with Chinese engineers. LCA calculations are  
 206 conducted with OpenLCA utilizing background data from the ecoinvent database 3.6.  
 207 We assume that all production and use stages are typical for the context of the Shandong  
 208 Province. For details on additional methods and data please see section 1 and 5 in the  
 209 SI. Emission factors of materials and data quality are listed in Table S15 and S16.



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Figure 2: Material thickness of 5 classes of asphalt pavement, where 0 represents the surface. SMA: stone-matrix asphalt. AC: asphalt concrete. Number (e.g. 13 in SMA-13) refers to nominal maximum aggregate size (mm). See Table S2.

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## 2.2 Scenarios sets and Sensitivity

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This study explores the uncertainty of both pavement-related GHG emissions as well as the potential for their mitigation. Three sets of scenarios are summarized in Table 3. They consider the key factors defined for this study including: M&R schedule, alternative materials, and vehicle electrification. The study also varies parameters of binder and moisture content, energy consumption in manufacturing and construction, transport distance, and traffic growth rate covering the range of reported values in various reports and peer-reviewed papers (Tables S7, S8).

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### Scenario 1: Maintenance and Rehabilitation schedule

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M&R schedules can have a significant impact on the full life cycle GHG emissions of pavement by consuming materials and energy while influencing pavement performance in the use phase. Studies have been conducted on optimizing the total cost of M&R considering the user cost and agency cost throughout the life cycle (Chan et al., 2008; Guo et al., 2019; Guo et al., 2020; Mandapaka et al., 2012; Pratico et al., 2011) while the discussion on carbon optimization is relatively limited (Pellicer et al., 2016; Renard et al., 2021; Torres-Machi et al., 2014). The central question to be answered in this optimization problem is which M&R option to choose at each stage to achieve a minimization of GHG emissions over the entire life cycle. The level of roughness and deflection were restricted in the optimization model in a way that met the requirements of pavement service performance and standards (Ministry of Transport, P., 2018). The detailed description of the optimization model is summarized in the SI (Table S9 and Eq S5-S19).

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### Scenario 2: Alternative low carbon material

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Materials can directly and indirectly contribute to a significant percentage of life cycle GHG emissions of pavement. Among many recycled materials discussed, RAP and

242 RCA show considerable advantages in availability and relatively satisfactory road  
243 performance. However, it is still challenging to use 100% of these recycled materials  
244 as they might not provide the same function as traditional materials. Therefore, the  
245 percentage of recycled materials should be decided on in tandem with the production  
246 method and the use of additives (e.g. sasobit), considering the required pavement  
247 quality and energy input. The change in road features can have important impacts on  
248 use-phase emissions. See Table S8 in the SI for detailed assumptions and discussions  
249 on material content, energy consumption, density, and stiffness changes.  
250

### 251 *Scenario 3: Adoption of EVs*

252 The promotion of EVs can potentially reduce GHG emissions of the use phase. This  
253 section discusses the use-phase GHG emissions under different scenarios with EV  
254 adoption rate, EV on-board energy consumption, and carbon intensity of electric  
255 charging. For simplicity and considering the low average age (3.3 years) of passenger  
256 cars in China (Launch Tech and DataEye, 2015), we use projected EV sales percentages  
257 in 2030 as a replacement percentage. Detailed assumptions on EV adoption rates, future  
258 EV energy use improvements, and carbon intensity of electricity can be found in Table  
259 S11, Table S12, and Table S13. The most optimistic scenario, for example, assumes  
260 very high adoption rates and energy use improvements of EVs, and an electricity mix  
261 consistent with the international 2°C target. While the High EV scenario considers a  
262 substantial uptake of EVs, no change in the pavement design code is reported by the  
263 respective authorities as a result of fleet electrification (Ministry of Transport). While  
264 the increasing trend in the use of EVs can substantially reduce the use-phase impact of  
265 the pavement life cycle, their effect on pavement design and future code adoption  
266 remains an open question.  
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Scenario Set	Scenarios
<b>M&amp;R schedule</b>	<ul style="list-style-type: none"> <li>- Fixed schedule</li> <li>- Maintenance optimization</li> </ul>
<b>Alternative materials</b>  <i>Key parameters varied:</i> <ul style="list-style-type: none"> <li>- Material composition</li> <li>- Energy consumption</li> <li>- Stiffness</li> <li>- Density</li> <li>- Cement content</li> <li>- Moisture content</li> </ul>	For surface asphalt (A) layer: <ul style="list-style-type: none"> <li>- A1: Typical materials and method (HMA)</li> <li>- A2: 20% RAP, with 30% of asphalt binder in RAP function as binder</li> <li>- A3: 40% RAP + 3% Sasobit, with 50% of asphalt binder in RAP function as binder</li> <li>- A4: WMA (3% Sasobit)</li> </ul> For base granular (G) layers: <ul style="list-style-type: none"> <li>- G1: Traditional materials and method with 4.8% cement</li> <li>- G2: 40% RAP with 4.5% cement</li> <li>- G3: 42% RCA with 4.3% cement</li> <li>- G4: 40% RAP + 60% milled granular layer with 4.5% cement</li> <li>- G5: Traditional materials for production + Cold-in-place recycle for structural M&amp;R with 4.5 cement</li> <li>- G6: 40% RAP + 60% milled granular layer for production + Cold-in-place recycle for structural M&amp;R with 4.5 cement</li> </ul>
<b>EV adoption</b>  <i>Key parameters varied:</i> <ul style="list-style-type: none"> <li>- Fuel economy</li> <li>- Electricity GHG emission factor</li> <li>- EV adoption rate</li> </ul>	EV adoption rate: <ul style="list-style-type: none"> <li>- None</li> <li>- Low (L): 8.5-37% passenger, 0.5% truck</li> <li>- Medium (M): 11.7-44.5% passenger, 1.3% truck</li> <li>- High (H): 14.8-52% passenger, 2.0% long-haul truck</li> </ul> On-board energy consumption: <ul style="list-style-type: none"> <li>- Current</li> <li>- 2030 prediction</li> </ul> Carbon intensity of the electricity mix: <ul style="list-style-type: none"> <li>- 2018 electricity</li> <li>- 2030 BAU</li> <li>- 2030 New policy</li> <li>- 2030 450 ppm policy (2°C target)</li> </ul>

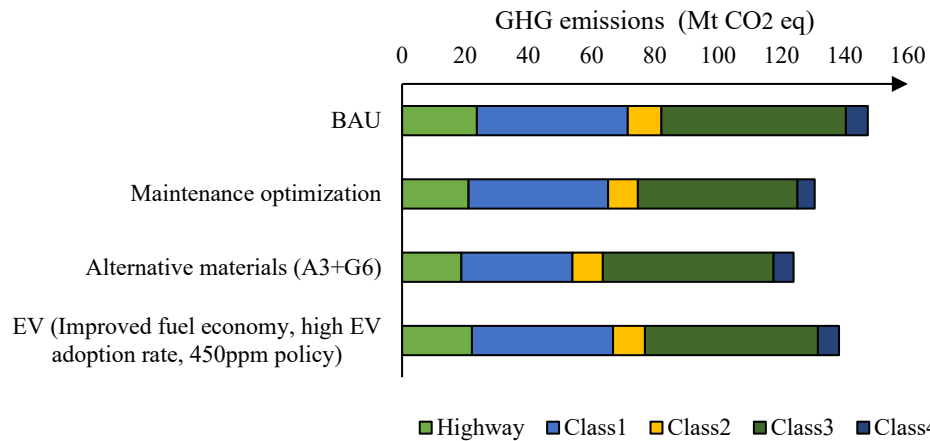
## 270 3. Results

## 271 3.1 LCA results at provincial level

272 Figure 3 shows the life cycle GHG results of the newly constructed road network in the  
 273 province of Shandong through 2030. The life cycle GHG emissions of highway and  
 274 Class 1-4 roads to be constructed in the next 10 years are about 24, 48, 11, 58, and 7  
 275 Mt CO<sub>2</sub>-eq, respectively, with a total of 147 Mt CO<sub>2</sub>-eq under the business as usual  
 276 (BAU) scenario. Estimated impacts under various alternative scenarios are shown in  
 277 Figure 3. Material replacement has a major influence on the results, with a potential  
 278 reduction of up to 16% reduction or 24 Mt CO<sub>2</sub>-eq. Similarly, M&R optimization and  
 279 EV adoption could reduce life cycle GHG emissions by about 17 Mt (11%) and by up  
 280 to 9 Mt (6%), respectively.

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282  
 283 *Figure 3. Life cycle GHG emissions of newly built roads in the Shandong Province through*  
 284 *2030 and across selected scenarios. The expansion of the road network is by 2287, 5317,*  
 285 *4560, 22998, and 6145 km for highways and class 1-4 roads, respectively (Table S1).*

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287 **3.2 Life cycle GHG emissions of pavement per functional unit**

288 Pavement design and assumed traffic volume have a large influence on life cycle GHG  
 289 emissions. The life cycle GHG emissions (see Table S5) for Highway, Class 1, Class  
 290 2, Class 3, Class 4 are on average 1,720, 1,500, 1,170, 860 and 570 tons CO<sub>2</sub>-eq/lane-  
 291 km as shown in Figure 4. Pavements experiencing higher traffic volumes require more  
 292 material and energy inputs during the manufacturing, construction, and transport stages.  
 293 The results for each class of pavement generally vary from around -15% to +25%,  
 294 depending on parameter selection, as shown by the error bars in Figure 4(a).

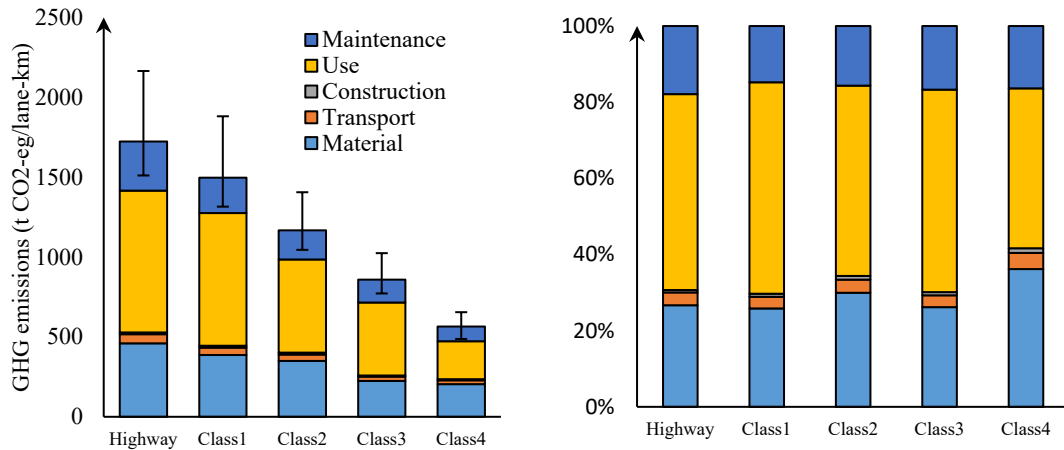
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296 A large share of impacts are due to the PVI effect in the use phase. As shown in Figure  
 297 4(b), the use phase emissions of all roads (except for Class 4) account for about 50-55%  
 298 of per-unit GHG emissions, while production emissions represent around 25-30% of  
 299 life cycle impacts per lane-km. The M&R phase, which includes the production and  
 300 transport of materials, and fuel consumption for on-site construction, contributes about  
 301 15-18% to specific GHG emissions. For the lower-traffic Class 4 roads, M&R has a  
 302 significantly higher impact share. Transport of materials and operation of on-site  
 303 machinery jointly accounts for less than 5% of emissions. Deflection-induced PVI,  
 304 which primarily impacts truck, dominates use-phase emissions (over 80%, Figure 4(c)).

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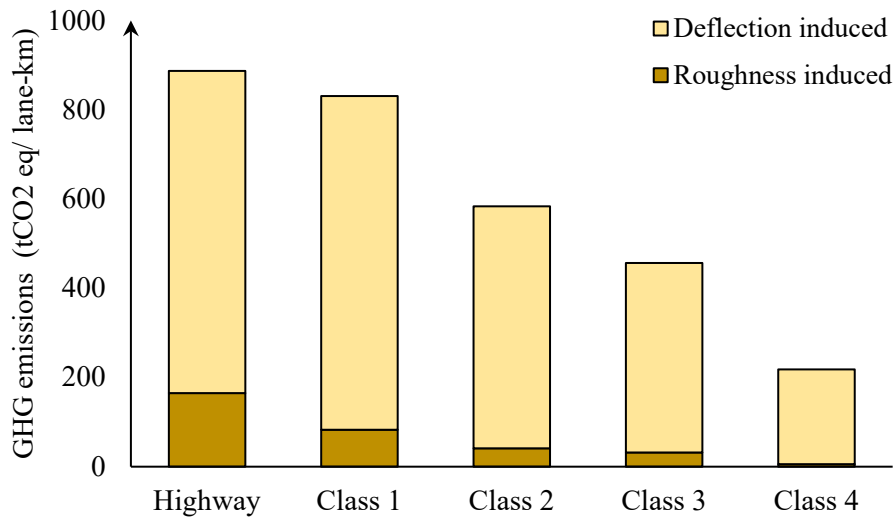
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(a) Life-cycle emissions with error bars illustrating the sensitivity of results.

(b) Relative contribution of each life cycle stage.

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(c) GHG emissions of the use stage.

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312 *Figure 4 GHG emissions by pavement class and life cycle stage.*

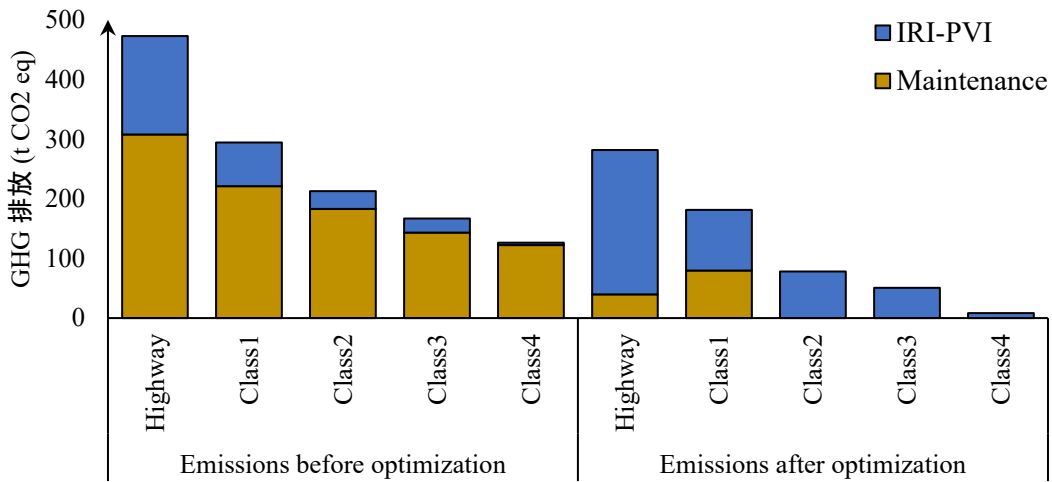
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### 314 3.3 Scenarios Analysis

#### 315 3.3.1 Life cycle GHG emissions under M&R schedule

316 Optimization of the M&R schedule can achieve reductions in GHG emissions, as  
 317 shown in Figure 5. Surprisingly, the optimization procedure identified longer  
 318 maintenance intervals than current practice to reap maximum emissions reductions.  
 319 Taking highways as an example, an optimal maintenance schedule would foresee a  
 320 functional maintenance in year 15, whereas current practice is to perform a functional  
 321 maintenance in year 7 and a structural maintenance in year 14, reducing M&R-related  
 322 emissions by 87%, while roughness-induced emissions increase by 46%. Overall, a  
 323 combined reduction in emissions of approximately 40% is achieved for highways and  
 324 Class 1 pavements. Reductions for Class 2-4 pavements are even higher, at 63%, 70%,  
 325 and 93%. Corresponding absolute emission reductions are 191 t, 113 t, 135 t, 116 t and  
 326 118 t CO<sub>2</sub>-eq per lane-km. In relative terms, the reduction in emissions for class 2-4  
 327 is around 10%.

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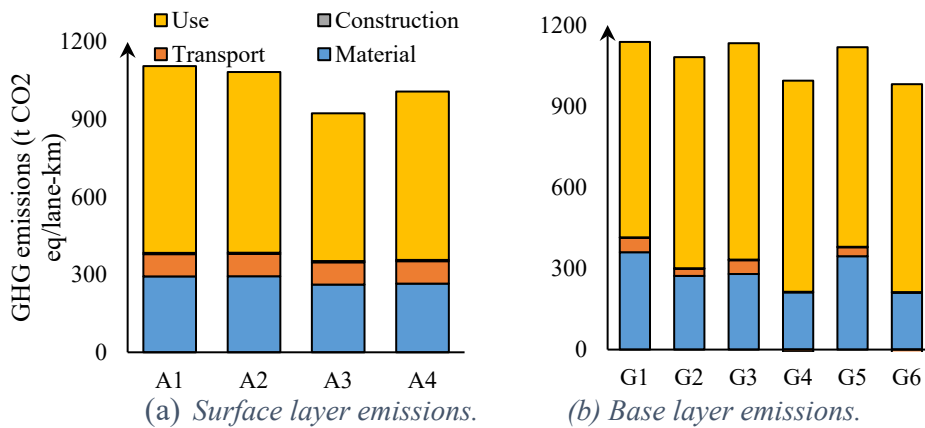
Figure 5. Maintenance and roughness-induced PVI emissions under non-optimized and optimized conditions.

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### 3.3.2 Life cycle GHG emissions under scenarios of alternative low carbon materials

335 Low-carbon surface and base materials can realize a combined reduction of specific  
336 life cycle emissions of up to 16% for highway pavements. The life cycle impacts of  
337 replacing surface materials and base materials for highways are shown separately in  
338 Figure 6, where the categories “material” and “transport” indicate the total production  
339 and transport of materials for both the construction and maintenance stages. For  
340 highways, scenarios A2, A3 and A4 can reduce emissions by 2.1%, 16.5%, and 8.9%.  
341 Material input, energy demand and stiffness changes are the underlying factors of  
342 overall GHG emissions. For example, A3 shows an 11% emissions reduction in the  
343 material production stage, whereas A2 achieves only a 0.3% reduction. This is due to  
344 the fact that A2 requires lower material inputs but increased energy demand for heating  
345 RAP. By contrast, the significantly reduced energy consumption of WMA combined  
346 with reduced heating energy requirements for fewer raw materials is responsible for the  
347 fact that the largest mitigation potential is achieved in scenario A3. Stiffness change, as  
348 shown in table S10, leads to use-phase GHG reductions of 3%, 21% and 10% in the  
349 scenarios A2, A3 and A4 respectively. In addition, emissions due to hauling RAP from  
350 the demolition site to the plant are offset by its avoided transport emissions from the  
351 demolition site to the landfill.

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(a) Surface layer emissions. (b) Base layer emissions.  
 Figure 6. Highway pavement emissions by life-cycle stage with alternative low carbon materials. A1: Traditional materials and method (HMA); A2: 20% RAP, with 30% of asphalt binder in RAP function as binder; A3: 40% RAP + 3% Sasobit, with 50% of asphalt binder in RAP function as binder; A4: WMA (3% Sasobit). G1: Traditional materials and method with 4.8% cement; G2: 40% RAP with 4.5% cement; G3: 42% RCA with 4.3% cement; G4: 40% RAP + 60% milled granular layer with 4.5% cement; G5: Traditional materials for production + Cold-in-place recycle for structural M&R with 4.5% cement; G6: 40% RAP + 60% milled granular layer for production + Cold-in-place recycle for structural M&R with 4.5% cement.

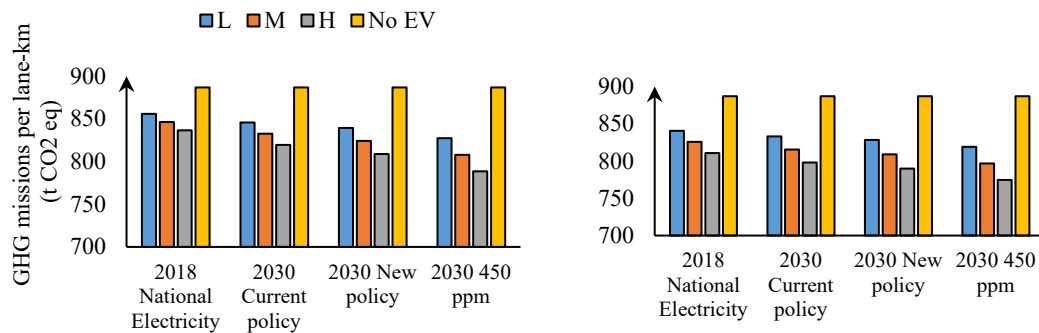
365 For highways, the total life cycle GHG reduction under the scenarios G2, G3, G4, G5  
 366 and G6 in all stages are 4.9%, 0.4%, 13.4%, 1.7%, and 15.2%. The decline in GHG  
 367 emissions in base layers mainly stems from the reduction in cement content. Cement  
 368 production accounts for 71% of the GHG emissions in the original base material  
 369 production stage. With RAP and RCA, the cement content is reduced from 4.8% to 4.5%  
 370 and 4.3%, respectively, resulting in a notable decline in GHG emissions. In addition,  
 371 production of crushed aggregates and gravel represents 14% and 7% of specific GHG  
 372 emissions, providing carbon mitigation potential for the replacement with RAP or RCA.  
 373 On the contrary, reduced base layer stiffness increases GHG emissions during the  
 374 use phase. The other road classes exhibit the same pattern as discussed for highways, see  
 375 Figure S1 and Figure S2. For Class 1-4 pavements, the G6 scenario achieves the largest  
 376 reductions in all stages combined, on the order of 19%, 23%, 15%, and 25%,  
 377 respectively.

378  
379

3.3.3 Life cycle GHG emission under scenarios of EV adoption

380 EV adoption only has a modest effect on emission reductions. Figure 7 illustrates the  
 381 use phase GHG emissions of each highway class with varying assumptions on EV  
 382 adoption rate, on-board energy use, and carbon intensity of electricity, as specified in  
 383 Table S11, S12 and S13. Assuming the current electricity mix and average EV energy  
 384 consumption, use-phase GHG emissions could be reduced by 31, 41 and 50 t CO2-  
 385 eq/lane-km under scenarios of low, medium and high EV adoption compared to the  
 386 100% ICEV scenario, which is labeled as ‘No EV’. The ideal scenario combination of  
 387 high EV adoption rate, low carbon electricity and reduced energy consumption would  
 388 yield a 13% use phase emissions reduction. Note that even in the high EV adoption  
 389 scenario, only 2% of long-haul trucks are assumed to be EV, which partly explains  
 390 the modest results. While EV replacement would contribute 6%, energy efficiency  
 391 and electricity de-carbonization would contribute 3% and 4%. For road classes 1-4,

392 use-phase emission reductions are on the order of 11% under the ideal scenario  
 393 (Figure S3).



394  
 395

396 *Figure 7. Use-phase emissions under various scenarios of vehicle electrification and*  
 397 *electricity-decarbonization, assuming (left) current EV on-board energy use and (right)*  
 398 *improved EV energy efficiency. L: low adoption rate of EVs; M: medium adoption rate*  
 399 *of EVs; H: high adoption of EVs; No EV: No adoption of EVs.*

#### 400 4. Discussion

401 This work identified a potential for considerable GHG reductions in the life cycle of  
 402 pavements in China. The total emissions from new pavement to be constructed in the  
 403 Shandong province is roughly equivalent to the combustion emissions of 6.5 million of  
 404 China's most popular vehicle, the VW Jetta, traveling for 15,000 km each year for 10  
 405 years (assuming an emissions factor of 151g CO<sub>2</sub>/km (DOE, 2020)). Considering that  
 406 Shandong is already one of the leading provinces in terms of transport infrastructure  
 407 maturity, there is an urgent need to mitigate these emissions. We find that not only the  
 408 construction of new pavements is carbon-intensive but also the rehabilitation of the  
 409 existing pavement stock. In our calculations and projections, the use-phase and M&R  
 410 stages, which have been commonly neglected in previous research, have the largest  
 411 contribution to life-cycle emissions per lane-km (more than 60%). Managing these  
 412 phases in an optimal way can therefore provide a considerable carbon reduction  
 413 opportunity for the transport sector. The use of alternative materials has a large potential  
 414 for emission reductions as well, albeit to a lesser degree.

415

416 Our results indicate that pavement construction in China is more carbon-intensive per  
 417 lane-km compared to developed countries. For example, despite the fact that our  
 418 assumptions on lifespan are much shorter than those of Xu et al., 2019, who study life  
 419 cycle emissions of pavement construction in the US, we find comparable life cycle  
 420 emission results for pavements with similar assumptions (i.e. Road 4 in Xu's study vs  
 421 our Class 1, Road 5 vs. our Class 4), as shown in Figure S5 and Table S14.

422

423 The thinner surface layer used for Shandong's pavements and the resulting high  
 424 deflection-induced PVI indicate an important opportunity for low-carbon pavement  
 425 design. Assuming a 30% increase in surface layer height, which would be in accordance  
 426 with the design code and practice of developed countries, would lead to a 23%  
 427 reduction of deflection-induced PVI emissions. After taking into account the increased  
 428 material and construction emissions, the net reduction in life cycle pavement emissions  
 429 are still in the range of around 5 -8%. The findings of this paper support  
 430 recommendations for further improvements in road quality in China from the  
 431 perspective of life cycle carbon management (cf. Liu et al., 2020; Wang, 2013).

432

433 The substantial carbon reduction potential of the M&R optimization scenario supports  
434 the need for further quality assessment of pavements and M&R planning. However,  
435 there are still factors that deserve further consideration. One possible criticism of our  
436 work is the simplified assumption of constant layer height, which allows for unchanged  
437 deflection-induced PVI in our model. In reality, popular M&R methods like overlays  
438 can influence deflection-induced PVI with altered layer height. Another potential  
439 drawback is the insufficient consideration of the cost of road users, i.e. drivers. Given  
440 that user cost is an important factor affecting the economics of road M&R schedule  
441 (Chan et al., 2008; Mandapaka et al., 2012; Praticò et al., 2011), the current model may  
442 overestimate the potential cost effectiveness. In addition, M&R decisions in China  
443 require comprehensive consideration of several indicators (Ministry of Transport, 2018)  
444 in addition to life cycle carbon emissions (Figure S6). Finally, although we carefully  
445 selected data appropriate for the context of the Shandong province, whenever possible,  
446 further context-specific data may be needed in the future, such as emission factors for  
447 asphalt materials, sales scenarios for electric vehicles, as well as pavement deterioration  
448 curves.

449

450 Specific emissions of each stage calculated in this study are generally consistent with  
451 existing research, except for a remarkably large gap between deflection-induced PVI  
452 and roughness-induced PVI (Akbarian, 2012; Akbarian, 2015; Louhghalam et al., 2017;  
453 Xu et al., 2019). The disproportionately high emissions from deflection-induced PVI  
454 can be explained by two main factors: (1) the notably thin pavement surface, and (2)  
455 the high truck traffic in China. The slow growth of IRI as calculated by the selected IRI  
456 model further contributed to the disproportionately high use-phase emissions. However,  
457 the academic discussion of IRI prediction models for roads in China is scarce (Wu and  
458 Zhou, 2009; Wang and Han, 2007; Zhao et al., 2018).

459

460 In practice, pavement design and M&R frequency largely depend on various factors,  
461 including local geography, traffic volume, and available construction funds. In this  
462 work, simplified assumptions for these factors were necessary due to limited data.  
463 Changes in pavement stiffness due to the use of alternative materials should ideally be  
464 based on real-world experimental data, which was not available for this work. However,  
465 for this study input data was selected with great care based on a synthesis of a large  
466 number of mutually independent studies, interviews with engineers, and official  
467 government documents. Future research should include albedo effects, which can  
468 further influence use phase emissions.

469

## 470 5. Conclusion

471 The goal of this study was to provide a systematic environmental management  
472 perspective on road construction in a Chinese province that can be replicated and  
473 widely applied. Pavements cause significant GHG emissions throughout their life cycle.  
474 While previous work omitted the use phase, this work demonstrates that it contributes  
475 the most to total emissions, followed by material production, and M&R. Optimized  
476 pavement structure designs and M&R schedules can substantially reduce life cycle  
477 GHG emissions by up to 16% and 11% respectively. Vehicle electrification has a  
478 negligible impact on the reduction of use-phase emissions (up to 6%). Carbon  
479 management should be incorporated into decision-making for new road planning and  
480 M&R.

481

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