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RESEARCH ARTICLE

Genomic estimation of quantitative genetic parameters in wild admixed populations

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Abstract

- 1. Heritable genetic variation among free-living animals or plants is essential for populations to respond to selection and adapt. It is therefore important to be able to estimate additive genetic variance V_A , which can be obtained using a GLMM known as the *animal model*. An underlying assumption of the standard animal model is that the study population is genetically unstructured, which is often unrealistic. In fact, admixture might be the norm rather than the exception in the wild, like in geographically structured populations, in the presence of (im)migration or in reintroduction and conservation contexts. Unfortunately, animal model sthat account for genetically differentiated subpopulations have recently become popular, but methodology is currently only available for cases where relatedness among individuals can be estimated from pedigrees.
- 2. To ensure that genetic group animal models with heterogeneous V_A remain applicable to populations with genomic data but no pedigrees, there is a clear need to generalize these models to the case when exclusively genomic data are available. We therefore introduce such methodology for wild admixed systems by extending methods that were recently suggested in the context of plant breeding. Our extension relaxes the limiting assumptions that currently restrict their use to artificial breeding set-ups.
- 3. We illustrate the usefulness of the extended genomic genetic group animal model on a wild admixed population of house sparrows resident in an island system in Northern Norway, where genome-wide data on more than 180,000 single nucleotide polymorphisms (SNPs) are available to derive genomic relatedness. We compare our estimates of quantitative genetic parameters to those derived from a corresponding pedigree-based genetic group animal model. The satisfactory agreement indicates that the new method works as expected.
- 4. Our extension of the very popular animal model ensures that the upcoming challenges with increasing availability of genomic data for quantitative genetic studies of wild admixed populations can be handled. To make the method widely

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available to the scientific community, we offer guidance in the form of a tutorial including step-by-step instructions to facilitate implementation.

KEYWORDS

additive genetic variance, admixed populations, animal model, breeding value, genetic groups, genomic relatedness matrix, house sparrows, local ancestry

1 | INTRODUCTION

A major goal in quantitative genetics is to disentangle environmental and genetic contributions to a phenotype within a study population (Charmantier et al., 2014; Falconer & Mackay, 1996; Lynch & Walsh, 1998). The partitioning of the phenotypic variance in a population into additive genetic and environmental components is of particular interest, as the additive genetic variance (V_A) is a crucial determinant of the rate by which phenotypes may respond to selection across generations. Simply speaking, the larger the V_A of a focal trait, the faster a population is able to respond to a given strength of selection, leading to a higher rate of adaptive evolution (Walsh & Lynch, 2018). The magnitude of V_A can thus be a determinant for how quickly and well wild populations may adapt to changing environmental conditions. Insight into such mechanisms is particularly important in conservation and wildlife management (Carlson et al., 2014; Frankham et al., 2010), and also highly relevant in breeding programs (Nyquist & Baker, 1991).

One well-established statistical tool to estimate genetic parameters like V_A is the linear mixed effects model known as the animal model (Henderson, 1984; Kruuk, 2004; Wilson et al., 2010). The animal model relies on knowledge about genetic relatedness between every pair of individuals (stored in a genetic relatedness matrix), which should reflect how similar they are at causal loci within their genomes (Speed & Balding, 2015; Weir et al., 2006). We denote the additive genetic impact on an individual's phenotype as its genetic value (or breeding value). Importantly, the V_A estimated by the animal model is relative to the so-called base population, while the definition of the base population itself depends on how one chooses to measure genetic relatedness (Legarra, 2016). The basic animal model relies on the assumption that there is a single, homogeneous base population without genetic substructures, and thus the model may produce biased estimates of V_A when base populations consist of individuals with systematically different genetic parameters (Wolak & Reid, 2017).

Genetic substructures are present in populations where breeding occurs between individuals from genetically divergent populations (e.g. due to historical isolation). The resulting gene flow is known as *admixture* (Tang et al., 2005), and is frequently encountered in the contexts of cross-breeding (Toosi et al., 2010) and hybridization (Grabenstein & Taylor, 2018). Instances of admixture in wild systems can arise in various ways, both naturally and humaninduced (Lenormand, 2002). Populations can, for example, receive immigrants from distant populations (e.g. Wolak & Reid, 2017), and metapopulations are subject to ongoing admixture between subpopulations through dispersal (e.g. Saatoglu et al., 2021). In conservation contexts, reintroduction schemes may involve translocating individuals from elsewhere to reinforce endangered populations (e.g. Ranke et al., 2020). Admixture is indeed expected to be relevant in most wild populations, because some level of dispersal and gene flow occurs into all populations except the few that are completely isolated (Bowler & Benton, 2005; Ronce, 2007). Consequently, the aforementioned homogeneity assumption of the basic animal model might frequently be violated in wild systems, potentially producing biased estimates. Thus, to accurately estimate the genetic parameters of wild populations, admixture should be taken into account.

Fortunately, animal models have been extended to account for admixture by partitioning the base population into genetic groups (Quaas, 1988; Wolak & Reid, 2017), where group-specific mean genetic values, and optionally group-specific additive genetic variances, are allowed. If the genetic groups only differ in the mean breeding value, but not in their V_A , then allowing for group-specific mean genetic values is sufficient to correct the aforementioned biases. However, in some applications it is more realistic to also allow for group-specific additive genetic variances, which comes at the cost of higher demands on both computation time and data (Muff et al., 2019).

In genetic group animal models (or simply genetic group models), we will differentiate between purebred individuals whose genomes belong to a single genetic group, and admixed individuals whose genomes are a mix of contributions from two or more genetic groups. As an example, when a local population receives immigrants, genetic group models may assign known natives as purebred in one genetic group, and known immigrants as purebred in a second genetic group, thereby explicitly incorporating the base population's genetic structure into the model (Wolak & Reid, 2017). The admixed set of individuals then contains all descendants from native/immigrant matings, and we track the respective admixture proportions in each individual in later generations, for example by following the pedigree. Some genetic group animal models also include segregation variance, an additional source of variance that emerges under admixture. These variances are mostly relevant in artificial breeding scenarios, or when the number of loci impacting the phenotype is very small (Slatkin & Lande, 1994). Fortunately, segregation variances are negligible under the assumption of the infinitesimal model (Bulmer, 1971), which is standard for complex traits in humans, animal and plant breeding, as well as natural populations (Hill, 2012; Hill & Kirkpatrick, 2010).

Genetic relatedness estimates used for basic animal models have traditionally been derived from pedigrees. Similarly, pedigree-based genetic group extensions of the animal model that account for admixture are well-established in the plant and animal breeding literature (García-Cortés & Toro, 2006; Lo et al., 1993; Schaeffer, 1991), and have recently found their way into applications for wild animal systems (Muff et al., 2019; Reid et al., 2021; Wolak & Reid, 2017). Pedigrees can produce estimates of relatedness and global ancestry (the overall proportion of a genome belonging to some genetic group) that are true on expectation (given a correct pedigree), by tracing all matings and applying the usual Mendelian rules of inheritance (Wright, 1922). Pedigree-based animal models have favourable computational properties due to the sparseness of the relatedness matrices and their inverses (Henderson, 1984), but these methods also have some inherent weaknesses. Realized genetic relatedness and global ancestry, for example, often differ greatly from the expected value derived from pedigrees (Hill & Weir, 2011), and pedigrees are often error-prone or incomplete (Keller et al., 2001; Ponzi et al., 2019).

Animal models that use single nucleotide polymorphisms (SNPs) to derive relatedness (Speed & Balding, 2015; Stanton-Geddes et al., 2013; Wang et al., 2017) have become increasingly popular due to improved genotyping technologies and decreased costs (Andrews et al., 2018). Provided that the number of genotyped loci is high enough, such *genomic* animal models, which rely on *genomic* relatedness matrices (GRMs), generally provide more accurate estimates of quantitative genetic parameters than pedigree-based animal models (Bérénos et al., 2014; Gienapp et al., 2017). On the other hand, the genomic approach is more computationally challenging compared to the pedigree-based version, as GRMs are dense.

A current drawback of genomic animal models is that it is still unclear how to handle genetic groups, in particular in the case of wild admixed populations with heterogeneous V_{Δ} . Recently, Rio, Moreau, et al. (2020) developed a heterogeneous V_A genetic group animal model with a genomic framework (denoted MAGBLUP-RI) for use in plant breeding. MAGBLUP-RI relies on knowledge about local ancestry, that is, the ancestry of each individual allele, indicating which genetic group that allele is descended from (Gravel, 2012). Through the use of local ancestry information, the model explicitly incorporates the fact that an admixed individual's genome is a mosaic of ancestries from different genetic groups. Local ancestry can be inferred from genotype data (Geza et al., 2019; Schubert et al., 2020), and thus MAGBLUP-RI is a genetic group model that solely relies on genomic data. The main drawback of the MAGBLUP-RI method is that it assumes homozygosity at all loci. While this restriction might be justified in a plant breeding set-up, it precludes use of the model on wild study systems, where heterozygous SNPs are common. Moreover, Rio, Moreau, et al. (2020) allow for only two genetic groups, which is likely to be insufficient in fragmented systems in the wild.

In this paper (based on a master's thesis by Aase, 2021), we overcome existing limitations of previous methodology and generalize *genomic* genetic group animal models such that they can be readily used to analyse wild admixed populations of diploid individuals with group-specific V_A . To this end, we have extended MAGBLUP-RI so that it handles SNP data with heterozygous (rather than only homozygous) loci and any number of genetic groups. Additionally, we assume that segregation variances are negligible, although we also provide full model derivations for segregation terms in Supporting Information S1. As a proof of concept, we have applied the extended genomic genetic group animal model to a metapopulation of house sparrows *Passer domesticus*, and compared the results to a pedigree-based, but otherwise equivalent, model proposed by Muff et al. (2019). All the steps of the implementation are detailed in a tutorial (Supporting Information S2).

2 | MATERIALS AND METHODS

2.1 | The animal model and genetic group models

The simplest version of the animal model for a continuous phenotype y_i of individual i in a population of size N is given by the mixed model

$$y_i = \mu + g_i + \varepsilon_i,$$

where the intercept μ is a fixed effect, and the genetic (or breeding) value g_i and the independent residual term $\varepsilon_i \sim N(0, \sigma_{\varepsilon}^2)$ of individual *i* are random effects (Kruuk, 2004; Wilson et al., 2010). Additional fixed and random effects can be included to account for other (non-genetic) sources of covariance between observations, such as sex, time of measurement or environmental effects (Kruuk & Hadfield, 2007). Fitting the model involves estimating the values of fixed effects and the parameters of the probability distributions of random effects. We give the vector of genetic values $\mathbf{g} = (\mathbf{g}_1 \dots \mathbf{g}_N)^T$ the distribution $N(\mathbf{0}, \sigma_G^2 \cdot \mathbf{G})$, where the random effect parameter σ_G^2 (to be estimated) is the V_A and \mathbf{G} is a matrix containing known estimates of all pairwise genetic relatednesses between individuals. The mean genetic value (of zero) and the additive genetic variance σ_G^2 are assumed to be homogeneous within the base population. The entries of \mathbf{G} can be estimated from a pedigree or from genomic data.

In the case where **G** is a GRM, that is, when its entries are derived from genomic (SNP) data, a widely used estimate was given by VanRaden (2008), where the genomic relatedness G_{ij} between individuals *i* and *j* is

$$G_{ij} = \frac{\sum_{m=1}^{M} (v_{im} - 2\hat{p}_m) (v_{jm} - 2\hat{p}_m)}{2\sum_{m=1}^{M} \hat{p}_m (1 - \hat{p}_m)},$$
(1)

with M being the number of genotyped loci, v_{im} and v_{jm} being the number of copies of the alternate allele (usually the minor allele) at the m^{th} locus in diploid individuals *i* and *j* respectively (known from genotyping) and \hat{p}_m being the estimated allele frequency of the alternate allele at locus *m*. A relatedness measure is thus obtained by comparing the alleles of two individuals at every SNP, while

weighting by the allele frequencies at each locus. Sharing rare alleles thus contributes more to relatedness estimates than sharing common alleles.

When the assumption that the genetic values in the entire base population are identically distributed is violated, the animal model should be extended accordingly. In the genetic group animal model, admixture is accounted for by assuming each genetic group r = 1, ..., R has a group-specific mean genetic value γ_r (rather than mean 0) and (if desired) a group-specific V_{Δ} (Muff et al., 2019; Wolak & Reid, 2017). To this end, we partition g into group-specific genetic value vectors $\mathbf{g}^{(r)} \sim N\left(\mathbf{0}, \sigma_{G_r}^2 \cdot \mathbf{G}_r\right)$, where $\sigma_{G_r}^2$ is the groupspecific V_{A} and \mathbf{G}_{r} is a group-specific relatedness matrix. Let the total genetic value U_i be the sum of group-specific genetic values whose probability distributions depend on the group r. If a purebred individual in group r has mean genetic value γ_n denoted as the genetic group effect of group r, then $U_i = \sum_{r=1}^{R} (\pi_{ir} \gamma_r + g_i^{(r)})$, where π_{ir} is the global ancestry of individual i with respect to group r. For identifiability reasons we add the constraint that one of the groups must have $\gamma_r = 0$, and we label this group as the 'reference group'. Thus, genetic group effects γ_r can be estimated in the animal model by including estimates of global ancestries π_{ir} as fixed effect covariates, while group-specific additive genetic variances $\sigma_{G_{c}}^{2}$ can be found by including the stochastic part of U_i , namely $g_i^{(r)}$, as random effects. The estimators for the global ancestries π_{ir} and group-specific genetic relatedness matrices \mathbf{G}_r depend on the genetic group method at hand (e.g. García-Cortés & Toro, 2006; Muff et al., 2019; Rio, Moreau, et al., 2020; Wolak & Reid, 2017).

2.2 | Genomic genetic groups model in an inbred system

Rio, Moreau, et al. (2020) previously included admixture in an animal model based on genomic data (which they label MAGBLUP-RI). They let the total genetic value U_i be the sum of numeric contributions to the phenotype from each genotyped locus (i.e. a sum of allele effects). The contribution from each locus depends on its genotype and, importantly, on its *local ancestry*. All loci are assumed to be homozygous, and thus each has only two possible genotypes (both chromosomes have either the reference or alternate allele), as is the case in highly inbred plant breeding systems. In other words, the organisms under study are treated as *de facto* haploid. Further, only two genetic groups are assumed to exist. While the contribution of a locus given its local ancestry and genotype is considered deterministic, the genotype and local ancestry of each locus are themselves considered random variables, and therefore U_i will also be a random variable.

Rio, Moreau, et al. (2020) use the statistical properties of the total genetic values U_i to derive the distribution of the group-specific genetic values $g_i^{(r)}$. To derive the entries of the group-specific GRMs, the covariances between total genetic values U_i must be considered. A central assumption here is that genotypes on different loci do not correlate, so no linkage disequilibrium (LD) is present. In addition to

the expected total genetic value of a purebred individual in group r, γ_r , Rio, Moreau, et al. (2020) define the parameters $\theta_{ij}^{(r)}$ and $\Gamma_{ij}^{(r)}$. We can interpret $\theta_{ij}^{(r)}$ as the proportion of the alleles of individuals i and j that have the same local ancestry, averaged across all loci, that is, the overlap of r-descended regions in the two genomes. $\Gamma_{ij}^{(r)}$ can be thought of as a genetic relatedness conditional on shared group membership in group r. The expected value and covariance for the total genetic values (when ignoring segregation variances) has then been shown to be

$$\mathsf{E}\left(U_{i}|\boldsymbol{\pi}_{i}\right) = \sum_{r=1}^{2} \pi_{ir} \gamma_{r}, \qquad (2)$$

$$\operatorname{Cov}\left(U_{i}, U_{j} | \boldsymbol{\pi}_{i}, \boldsymbol{\pi}_{j}, \boldsymbol{\theta}_{ij}, \boldsymbol{\Gamma}_{ij}\right) = \sum_{r=1}^{2} \theta_{ij}^{(r)} \Gamma_{ij}^{(r)} \sigma_{\mathsf{G}_{r}}^{2}, \tag{3}$$

where $\sigma_{G_r}^2$ is the group-specific V_A for group r (Rio, Moreau, et al., 2020). Thus, as in other genetic group models, the expected total genetic value of individual i is a weighted average of the mean expected value in the genetic groups, where the weights π_{ir} are i's global ancestry proportions in the groups r = 1, 2 (Equation 2). Furthermore, using Equation (3) and estimators for $\theta_{ij}^{(r)}$ and $\Gamma_{ij}^{(r)}$, the entries of the group-specific GRMs were estimated by Rio, Moreau, et al. (2020) as

$$(G_{r})_{ij} = \underbrace{\frac{1}{M} \sum_{m=1}^{M} \lambda_{imr} \lambda_{jmr}}_{=\hat{\theta}_{ij}^{(r)}} \cdot \underbrace{\frac{\sum_{m=1}^{M} \lambda_{imr} \left(w_{im} - \hat{p}_{mr}\right) \lambda_{jmr} \left(w_{jm} - \hat{p}_{mr}\right)}{\sum_{m=1}^{M} \lambda_{imr} \lambda_{jmr} \hat{p}_{mr} \left(1 - \hat{p}_{mr}\right)}}_{=\hat{\Gamma}_{ij}^{(r)}}, \quad (4)$$

where $w_{im} = 1$ indicates the presence of two copies of the alternate allele on individual *i*'s m^{th} locus, $w_{im} = 0$ indicates two reference alleles, $\lambda_{imr} = 1$ if these alleles are descended from group r (and 0 otherwise) and \widehat{p}_{mr} is the estimated alternate allele frequency on locus *m* within group *r*. Note that the estimator $\hat{\Gamma}_{ii}^{(r)}$ in expression (4) is a version of the VanRaden (2008) GRM in Equation (1), modified so that genotypes only contribute to the relatedness estimate if they share the same local ancestry (i.e. when $\lambda_{imr} = \lambda_{jmr} = 1$). Like this, admixture is explicitly incorporated into the relatedness estimator. This modified GRM is scaled by the estimated genome overlap $\hat{\theta}_{ii}^{(r)}$ so that the impact on the group-specific additive genetic variance from a pair of individuals only comes from the proportion of their genes that originate from the same group. To model group-specific V_A with MAGBLUP-RI, one can thus include the random effects $\mathbf{g}^{(r)} \sim N\left(\mathbf{0}, \mathbf{G}_r \cdot \sigma_{\mathbf{G}_r}^2\right)$ for each group r, where the entries of \mathbf{G}_r are estimated in Equation (4).

2.3 An extension to wild systems

In order to make the genomic genetic group model applicable to wild systems, we need to extend MAGBLUP-RI such that it allows for heterozygosity, as well as for an arbitrary number of genetic groups $R \ge 2$. These extensions will allow us to model admixture in most wild study systems.

To allow for heterozygosity, we consider the contributions (allele effects) of the two alleles at each locus separately. We split the genotype indicator w_{im} into two allele indicators $w_{im}^{(1)}$ and $w_{im}^{(2)}$, where the local ancestries are indicated by the binary variables $\lambda_{imr}^{(1)}$ and $\lambda_{imr}^{(2)}$, respectively ($\lambda_{imr}^{(\cdot)} = 1$ indicates local ancestry from group r). When we write $w_{im}^{(h)}$, $h \in \{1, 2\}$ thus denotes which of the two copies of the chromosomes the allele is located on (we only consider diploid organisms). We further assume the contributions to the total genetic value U_i from a given locus to be equally weighted between its two present alleles. In other words, we assume a lack of dominance effects within loci with respect to their contributions to the phenotype, as the contribution from a heterozygous locus will be the mean of the effects of the possible homozygotes at the locus. As in MAGBLUP-RI, the allele effects are still assumed to depend on the locus, on the allele variant (reference or alternate) and on the local ancestry of the allele (i.e. the effect of the allele depends on which group the allele was inherited from). We also retain the assumption that allele variant indicators on different loci (e.g. m and $m' \neq m$) are uncorrelated (which results in a lack of LD). While the two allele indicators on the same locus are given identical probability distributions, we also assume that they are uncorrelated.

Similar to Rio, Moreau, et al. (2020), we consider the expected value and covariance between total genetic values U_i of different individuals, but under an updated definition of U_i where we model single-allele effects and an arbitrary number of groups. We only present the results and assume segregation variances can be neglected, but the full model derivation (including segregation variances) is shown in Supporting Information S1. Our derivation shows that the statistical properties of U_i found by Rio, Moreau, et al. (2020) are retained under the new definition of U_i. In other words, we show that Equations (2) and (3) still hold in the extended model, if we sum from r = 1 up to R rather than 2 in both equations. Thus, our estimate of the GRM entries remains $(G_r)_{ii} = \hat{\theta}_{ij}^{(r)} \cdot \hat{\Gamma}_{ij}^{(r)}$. However, the parameters $\theta_{ii}^{(r)}$ and $\Gamma_{ii}^{(r)}$ have somewhat expanded meanings in the extended model, as they now refer to overlap of allele ancestry and relatedness from comparing alleles separately (rather than genotypes) respectively. We therefore update their respective estimators to fit our single-allele paradigm.

Let us first generalize the parameter estimators from Rio, Moreau, et al. (2020) to account for single alleles. Let the global ancestry proportion π_{ir} for individual *i* in group *r* be estimated by the observed group membership proportion

$$\hat{\pi}_{ir} = \frac{1}{2M} \sum_{m=1}^{M} \sum_{h=1}^{2} \lambda_{imr}^{(h)},$$
(5)

and the group allele frequency p_{mr} at locus m in group r by the observed group allele frequency

$$\widehat{p}_{mr} = \frac{\sum_{i=1}^{N} \sum_{i=1}^{2} \lambda_{imr}^{(h)} w_{im}^{(h)}}{\sum_{i=1}^{N} \sum_{j=1}^{2} \lambda_{imr}^{(h)} \lambda_{imr}^{(h)}}.$$
(6)

Since the base population induced by using GRMs corresponds to the population from which the allele frequency is derived (Hayes et al., 2009; Legarra, 2016), the base population of a genetic group will, due to Equation (6), consist of its purebred individuals along with the parts of the genomes of admixed individuals which have local ancestry from that group. Admixed individuals are thus partially members of multiple base populations.

When dealing with local ancestry indicators $\lambda_{imr}^{(h)}$ and allele indicators $w_{im}^{(h)}$ in different individuals, the assignment of a chromosome copy to be h = 1 or h = 2 is arbitrary and will not correspond between different individuals. These designations thus cannot be differentiated when comparing different individuals, so both alleles at locus *m* in one individual must be compared to both alleles on locus *m* in another individual. Thus, we estimate the genome overlap coefficient $\theta_{ii}^{(r)}$ by

$$\widehat{\theta}_{ij}^{(r)} = \frac{1}{4M} \sum_{m=1}^{M} \sum_{h=1}^{2} \sum_{h'=1}^{2} \lambda_{imr}^{(h)} \lambda_{jmr}^{(h')}.$$
(7)

To arrive at the estimate for the group-conditional relatedness $\Gamma_{ij}^{(r)}$, we again compare both alleles at locus *m* in individual *i* with both alleles at locus *m* in individual *j*:

$$\widehat{\Gamma}_{ij}^{(r)} = \frac{\sum_{m=1}^{M} \sum_{h=1}^{2} \sum_{h'=1}^{2} \lambda_{imr}^{(h)} \left(w_{im}^{(h)} - \widehat{p}_{mr} \right) \lambda_{jmr}^{(h')} \left(w_{jm}^{(h')} - \widehat{p}_{mr} \right)}{\frac{1}{2} \sum_{m=1}^{M} \sum_{h=1}^{2} \sum_{h'=1}^{2} \lambda_{imr}^{(h)} \lambda_{jmr}^{(h')} \widehat{p}_{mr} \left(1 - \widehat{p}_{mr} \right)}.$$
(8)

Our GRM with entries $\hat{\theta}_{ij}^{(r)} \cdot \hat{\Gamma}_{ij}^{(r)}$ is a generalization of the wellknown GRM **G** in Equation (1), and also incorporates the use of local ancestry from the estimator in Equation (4). It is easy to see that in the one group case (R = 1) the GRM entries $\hat{\theta}_{ij}^{(r)} \cdot \hat{\Gamma}_{ij}^{(r)}$ simplify to the entries of **G** (shown in Supporting Information S1).

2.4 | Application to a house sparrow metapopulation

We applied the extended genomic genetic group animal model to a metapopulation of house sparrows living on islands in the Helgeland region in Northern Norway. This metapopulation has been subject to a long-term study since 1993, and the data used here include in total 4,625 phenotypic records of wing length, body mass and tarsus length from 1,932 sparrows measured between 1993 and 2016 on eight islands that are known to have inter-island dispersal (Baalsrud et al., 2014; Muff et al., 2019; Saatoglu et al., 2021). Large-scale genotyping of house sparrows from the study system has resulted in high-quality genotyped individuals, Niskanen et al., 2020). Across the 3,032 × 181,354 genotypes in the genomic dataset, roughly one third are heterozygous, which justifies the need for our extension of the genomic genetic groups model.

Subpopulations of sparrows located on the eight different islands vary in environmental conditions such as habitat, buffering 6

	Inner	Outer	Other	Admixed	Total
Genotyped	224	97	85	2,626	3,032
Phenotyped	197	73	41	1 620	1 932

TABLE 1 The numbers of genotyped and phenotyped individuals in the purebred reference populations (Inner, Outer and Other), the admixed population (Admixed) and in total (Total). Note that all phenotyped individuals were genotyped.

against bad weather and population density, and can be broadly defined as belonging to either the set of inner or outer islands (Baalsrud et al., 2014; Muff et al., 2019; Niskanen et al., 2020; Saatoglu et al., 2021). To account for possible genetic differences between the subpopulations originating from these two sets of islands, we partitioned the study population into genetic groups, namely an inner genetic group (encoded as 1) and an outer genetic group (encoded as 2). Sparrows from other islands in the study system (that were not systematically SNP genotyped) are also present in the dataset due to dispersal, and we place these sparrows into a third genetic group other (encoded as 3).

Throughout, we will be comparing the results from the genomic genetic group animal model to results from an otherwise equivalent pedigree-based model (Muff et al., 2019). The two models are identical, except that the pedigree-based model uses pedigree-derived global ancestries denoted q_{ir} (see Wolak & Reid, 2017), rather than $\hat{\pi}_{ir}$, and pedigree-derived group-specific genetic relatedness matrices (see Muff et al., 2019) instead of the group-specific GRMs **G**_{*r*}.

In a natural study system like ours, every individual is probably admixed at least to some extent due to dispersal in the past. To perform the local ancestry inference, however, we had to assign some individuals as purebred in the genetic groups of interest. We did this by classifying any individual that has both parents missing in the pedigree as a purebred in one of the groups for the genomic model, as determined by information about their natal island (Saatoglu et al., 2021; Table 1 shows the respective sizes of the purebred and admixed subpopulations). Thus, the starting condition of the genomic genetic groups model is as similar as possible to the corresponding pedigree-based genetic groups model (Muff et al., 2019) and hence allows a valid comparison between pedigree-based and genomic genetic group models. Conversely, individuals with at least one known parent in the pedigree were considered admixed in the genomic model. A further description of the empirical data from the house sparrow study system can be found in Supporting Information S3.

As a prerequisite for our chosen local ancestry inference method (Dias-Alves et al., 2018), we performed gametic phasing of the SNP data, which involves identifying which of the two alleles at each locus was inherited from which parent. In terms of our mathematical notation, this step determines which alleles within an individual should be designated which values of $h \in \{1,2\}$ together. After using PLINK 1.9 (Chang et al., 2015) to convert the genomic data to the appropriate input format, we used Beagle 5.1 (Browning et al., 2018) with default settings to perform the gametic phasing. The phasing was done separately on each of the purebred populations and the

admixed population since they are assumed to be genetically distinct. In addition to performing the gametic phasing, Beagle imputed any missing genotypes in the genomic data. The local ancestry $\lambda_{imr}^{(h)}$ of every considered allele in the admixed population was then inferred via the command-line version of the Python package Loter (Dias-Alves et al., 2018).

With values for the allele variant indicators $w_{im}^{(h)}$ and local ancestry indicators $\lambda_{imr}^{(h)}$, we estimated global ancestries π_r , group-specific allele frequencies \hat{p}_{mr} and group-specific GRMs \mathbf{G}_r for $r \in \{1,2,3\}$ using Equations (5)–(8). Despite only 1,932 sparrows having phenotype data, all 3,032 genotyped sparrows were used in setting up the group-specific relatedness matrices \mathbf{G}_r to improve the accuracy of the relatedness estimates. We used the R package BGData (Grueneberg & de los Campos, 2019) to manage the large genomic datasets, and to calculate the group-specific GRMs. Finally, a value of 10^{-12} was added to the diagonals of the \mathbf{G}_r matrices to ensure positive definiteness (since purebreds induce zeros on the diagonals).

Given the global ancestries π_r and group-specific GRMs \mathbf{G}_r , we can formulate the full genomic genetic groups model

$$y_{ij} = \mu + \mathbf{x}_{ij}^{\mathsf{T}} \boldsymbol{\beta} + \sum_{r=2}^{3} \hat{\pi}_{ir} \gamma_r + \sum_{r=1}^{3} g_i^{(r)} + id_i + island_{ij} + year_{ij} + \varepsilon_{ij}, \qquad (9)$$

where y_{ii} is the jth phenotypic measurement for individual i, μ is an intercept, \mathbf{x}_{ii} is a vector storing the fixed covariates sex (0 male and 1 female), age (in years since birth), month (May through August as a continuous covariate) and inbreeding coefficient F_{GRM} (computed by Niskanen et al., 2020), β is a vector of the fixed effects, $\hat{\pi}_{ir}$ is i's estimated global ancestry proportion in group r and γ_r is the fixed genetic group effects in groups $r \in \{1,2,3\}$. We set $\gamma_1 = 0$ for identifiability reasons, thus the genetic group effects γ_2 and γ_3 denote the deviations in the respective group's mean total additive genetic effect from inner (Wolak & Reid, 2017). The random effects include group-specific genetic values $g_i^{(r)}$, which are entries in the random vector $\mathbf{g}^{(r)} \sim N\left(\mathbf{0}, \sigma_{G_r}^2 \mathbf{G}_r\right)$, the permanent environmental effects $id_i \sim N(0, \sigma_{ID}^2)$, the effect of island of measurement *island*_{ij} ~ $N(0, \sigma_{island}^2)$, the effect of year of measurement *year*_{ij} ~ $N(0, \sigma_{\text{vear}}^2)$ and the residual term $\epsilon_{ij} \sim N(0, \sigma_{\epsilon}^2)$. Inclusion of the island of measurement as a common environmental random effect is especially critical in this model, as it ensures the model can distinguish between environmental island effects and the genetic group effects. We implemented all genetic group animal models in a Bayesian framework with the R-INLA package (Rue et al., 2009, see tutorial in Supporting Information S2). Fieldwork in the house sparrow metapopulation was carried out in accordance with permits from

the Norwegian Food Safety Authority (permit 6475), the Norwegian Environment Agency and the Ringing Centre at Stavanger Museum, Norway (permits 619 and 820).

3 | RESULTS

3.1 | Global ancestry proportions

As an initial plausibility check of the local ancestry inference, we first compared the estimated genomics-based global ancestry proportions $\hat{\pi}_{ir}$ to their pedigree-derived counterparts q_{ir} . Recall that π_{ir} was not estimated for purebred individuals, but was instead assumed to be 0 or 1 for use in a reference panel. We therefore only compare $\hat{\pi}_{ir}$ and q_{ir} for admixed individuals that were also phenotyped. For most individuals in this subpopulation, the global ancestry proportions estimated with the two methods were relatively similar (Figure 1), with correlations of 0.82, 0.83 and 0.60 for inner, outer and other, respectively. Both the correlations and the scatter plots thus indicate that the methods are in relatively good agreement.

We observe that individuals with global ancestries equal to 0 or 1, or fractions like 0.5, 0.25 and 0.75, are common in the pedigreebased model, but not in the genomic model. These individuals are the offspring of two purebreds of the same group (and thus also considered purebred by the pedigree model), 'hybrid' offspring of purebreds from different groups and offspring of hybrids back-crossed with purebred individuals, respectively. The accumulation of these values reflects that the pedigree-based model operates with expected inheritance of ancestry (which is, e.g. exactly 0.5 in a hybrid), whereas the genomic model considers realized ancestries. Another noteworthy difference is that more points lie above the diagonal than below for the inner global ancestry proportions (when q_{i1} is below ca. 0.75), reflecting that the genomic method tends to assign larger inner-ancestries than the pedigree-based method, while the opposite is true for outer (more points lie under the diagonal, for $q_{i2} > 0.25$). Both methods rarely assign large (>0.5) admixture proportions to the other group, but the pedigree-based method considers more individuals purebred in this group. Moreover, the agreement between pedigree-derived and genomics-derived global

ancestries is lowest for the other group, which is not surprising given that the group is a heterogeneous collection of individuals from all remaining islands. Generally, the methods seem to disagree the most for individuals that are considered purebred by just the pedigree ($q_{ir} \in \{0, 1\}$), and mostly agree regarding individuals with intermediate global ancestry proportions.

3.2 | Comparison of pedigree-based and genomic results

We report posterior means and 95% highest posterior density credible intervals (HPD Cls) for all parameters in the models, and also posterior modes for the random effect variances since they potentially have skewed posterior distributions (Tables 2 and 3). Additionally, the full posterior distributions for the genetic group means γ_r and variances $\sigma_{G_r}^2$ are displayed graphically for each of the three groups $r \in \{1,2,3\}$ (Figures 2 and 3 respectively). Since inner serves as the baseline for the mean genetic value, γ_1 was fixed to zero and therefore has no posterior distribution.

Generally, the agreement between the pedigree-based and the genomic models regarding fixed effects is quite good. The results indicate that sparrows descended from the inner islands tend to have genes for longer wings and body mass, as visible by the group differences in the genetic group means γ_r (Table 2; Figure 2). These estimated genetic group effects are somewhat more pronounced in the genomic models than the pedigree-based models, but the model disagreements are small compared to the uncertainties in the estimates. For tarsus length, the impact of genetic group ancestry on the mean genetic value is small to non-existent. Interestingly, this implies that genetically, sparrows with outer ancestry have longer tarsi relative to their body size (i.e. wing length and body mass) compared to sparrows with inner ancestry. Among the remaining fixed effect parameters, we see that for wing length the pedigree-based model finds slightly stronger negative effects of inbreeding than the genomic model, but again the differences are small relative to the uncertainty (Table 2). The pedigree-based and genomic models are in close agreement regarding all remaining fixed effects.

For the group-specific additive genetic variances, the two model paradigms produce posteriors with very similar shapes and with



FIGURE 1 Scatter plots for global ancestry proportion derived from the pedigree (*x*-axes, q_{ir}) and genomic local ancestry inference (*y*-axes, $\hat{\pi}_{ij}$). Each point refers to one individual. The plots only contain points for phenotyped admixed individuals (N = 1,620). Points are partially transparent to show density patterns in areas with overlapping points. Diagonals are shown as dashed lines.

TABLE 2 Posterior statistics for the fixed effects of the genetic group animal models, as derived from pedigree-based and genomic genetic group models for the three investigated phenotypic traits. For each parameter, the posterior mean is reported in the first row, and the 95% HPD CI in the second row. The parameters denote the effects of being female compared to male (β_{sex}), inbreeding ($\beta_{F_{GRM}}$), month (β_{month}) and age (β_{age}). The genetic group effects γ_2 and γ_3 of outer and other, respectively, denote the effect of being purebred in these groups relative to inner.

	Wing length		Body mass		Tarsus length	
Parameter	Pedigree	Genomic	Pedigree	Genomic	Pedigree	Genomic
β_{sex}	-2.75	-2.76	0.47	0.47	-0.09	-0.08
	(-2.89, -2.62)	(-2.89, -2.63)	(0.29, 0.64)	(0.29, 0.65)	(-0.16, -0.02)	(-0.15, -0.01)
$\beta_{F_{GRM}}$	-2.14	-1.77	-0.98	-0.93	-0.65	-0.67
	(-3.68, -0.60)	(-3.32, -0.23)	(-3.06, 1.09)	(-3.02, 1.15)	(-1.46, 0.16)	(-1.48, 0.15)
$\beta_{\rm month}$	-0.19	-0.19	-0.30	-0.30	0.03	0.03
	(-0.23, -0.15)	(-0.23, -0.15)	(-0.36, -0.24)	(-0.36, -0.24)	(0.02, 0.03)	(0.02, 0.04)
$\beta_{\sf age}$	0.46	0.46	0.09	0.09	-0.00	-0.00
	(0.43, 0.50)	(0.43, 0.50)	(0.03, 0.15)	(0.03, 0.15)	(-0.01, 0.00)	(-0.01, 0.00)
γ ₂	-0.24	-0.41	-0.52	-0.90	-0.02	0.07
	(-0.53, 0.07)	(-0.93, 0.13)	(-0.88, -0.15)	(-1.54, -0.27)	(-0.13, 0.10)	(-0.12, 0.25)
γ ₃	-0.18	-0.51	-0.35	-0.49	0.02	-0.06
	(-0.51, 0.15)	(-1.14, 0.12)	(-0.81, 0.10)	(-1.12, 0.12)	(-0.16, 0.19)	(-0.39, 0.25)

similar posterior modes (Table 3; Figure 3). Either the posteriors overlap almost completely, or the genomic model gives a posterior shifted to lower values than the pedigree-based model—the latter is a common and expected pattern when comparing these types of models (discussed further in the next section). In all models we find among-group differences in V_A for each phenotypic trait. For wing length and body mass, the V_A in outer tends to be higher than in inner, which again tends to be higher than in other (Figure 3). When it comes to tarsus length, the outer group probably has a smaller V_A than inner and other. Posteriors for the V_A are usually narrowest in inner, which is expected because sample sizes are largest in this group. All these differences between the groups (including differences in the γ_r estimates) indicate that the use of genetic group models is justified, or even necessary.

4 | DISCUSSION

In this article, we have extended the genetic group animal model such that it can be used with genomic data to estimate group-specific V_A for wild admixed populations. In order to derive the group-specific GRMs, we incorporated local ancestry information, and we considered single alleles to allow heterozygous genotypes. This is a generalization of previous methodology that was developed within a plant breeding set-up, where it was assumed that only homozygote genotypes were present (Rio, Moreau, et al., 2020). Furthermore, we allow for any number of genetic groups, rather than only two. As a proof of concept, we applied the extended method to a metapopulation of house sparrows, and show that the results are in line with the results from a corresponding, but pedigree-based, genetic groups model.

In our illustrative example, there was relatively good agreement between the genomic model and the pedigree-based model results, though we often observe stronger genetic group effects and smaller group-specific V_A in the genomic model. The disagreement on genetic group effects γ_r might be explained in part by the discrepancies in global ancestry proportions (Figure 1). Notably, the genomic vs. pedigree differences in V_A (i.e. smaller V_A in the genomic model) correspond to differences observed in otherwise equivalent genetic group animal models with homogeneous additive genetic variances (Figure S1, Supporting Information S4.1). This result suggests that the genomic genetic group method itself is not introducing a downward bias in V_A , but rather the inherent differences in using pedigree-based vs. genomic methods (see e.g. Powell et al., 2010). The pattern of V_A being smaller when estimated from genomic data than when estimated from pedigree data is often observed and has been extensively discussed in the literature (Evans et al., 2018; Gervais et al., 2019; Legarra, 2016; Yang et al., 2017).

Our results regarding the genetic group structure of the house sparrow system otherwise mirror what has been found previously regarding wing length and body mass (Muff et al., 2019), namely group differences in mean genetic value and V_A , which respectively suggest outer-descended birds have shorter wings and lower body mass, and lower V_A in these phenotypes compared to inner. Additionally, we find a lower V_A in tarsus length among innerdescended birds. Note that we have no guarantee against further genetic substructures within the genetic groups (e.g. inner is made up of five different islands), which could bias the results for groupspecific V_A (Wolak & Reid, 2017). To sum up, our genetic group model results indicate that outer and inner island sparrows may be genetically divergent in some of their phenotypes, possibly caused by TABLE 3 Posterior statistics for the random effect variances of the genetic group animal models, as derived from pedigree-based and genomic genetic group models for the three investigated phenotypic traits. For each parameter, the posterior mode and mean (formatted mode; mean) are reported in the first row and the 95% HPD CI in the second row. The parameters denote variance explained by year of measurement (σ_{year}^2), island of measurement (σ_{island}^2), permanent environmental effects (σ_{ID}^2), group-specific V_A ($\sigma_{G_r}^2$, r = 1 for inner, r = 2 for outer and r = 3 for other) and residual environmental effects (σ_e^2).

	Wing length	Wing length		Body mass		Tarsus length	
Parameter	Pedigree	Genomic	Pedigree	Genomic	Pedigree	Genomic	
$\sigma^2_{\rm year}$	0.03; 0.04	0.05; 0.06	0.05; 0.06	0.04; 0.05	0.01; 0.01	0.01; 0.01	
	(0.01, 0.09)	(0.02, 0.14)	(0.01, 0.16)	(0.01, 0.14)	(0.00, 0.04)	(0.00, 0.03)	
$\sigma^2_{ m island}$	0.09; 0.12	0.07; 0.08	0.11; 0.14	0.10; 0.14	0.00; 0.01	0.00; 0.01	
	(0.02, 0.35)	(0.02, 0.21)	(0.03, 0.45)	(0.02, 0.47)	(0.00, 0.03)	(0.00, 0.02)	
σ_{ID}^2	0.34; 0.35	0.43; 0.44	1.03; 1.05	1.09; 1.11	0.35; 0.36	0.35; 0.35	
	(0.21, 0.52)	(0.30, 0.61)	(0.75, 1.43)	(0.83, 1.48)	(0.31, 0.41)	(0.31, 0.40)	
$\sigma_{G_1}^2$	1.85; 1.86	1.57; 1.58	1.38; 1.40	1.12; 1.14	0.30; 0.30	0.29; 0.29	
	(1.57, 2.20)	(1.29, 1.93)	(1.00, 1.90)	(0.75, 1.61)	(0.23, 0.38)	(0.22, 0.38)	
$\sigma_{G_2}^2$	2.23; 2.27	1.94; 1.97	2.23; 2.28	2.27; 2.32	0.15; 0.16	0.07; 0.08	
-	(1.68, 3.08)	(1.37, 2.71)	(1.41, 3.41)	(1.37, 3.55)	(0.08, 0.27)	(0.01, 0.20)	
$\sigma_{G_3}^2$	1.59; 1.65	1.38; 1.43	0.83; 0.96	0.40; 0.53	0.29; 0.31	0.35; 0.36	
	(0.89, 2.79)	(0.75, 2.42)	(0.20, 2.45)	(0.06, 1.71)	(0.14, 0.57)	(0.17, 0.67)	
σ_{ϵ}^2	0.98; 0.98	0.98; 0.98	2.88; 2.89	2.90; 2.90	0.02; 0.02	0.02; 0.02	
	(0.93, 1.04)	(0.93, 1.03)	(2.73, 3.05)	(2.74, 3.06)	(0.02, 0.02)	(0.02, 0.02)	

FIGURE 2 The estimated posterior distribution of genetic group effects γ_r (i.e. mean genetic value) in the models for wing length (top), body mass (middle) and tarsus length (bottom). Posterior effects for the different genetic groups are shown in different colours. Posteriors from genomic models have solid lines, whereas the pedigree-based model posteriors are shown with dotted lines. Since inner is assumed to be the baseline mean, $\gamma_1 = 0$ is shown as a vertical line.





FIGURE 3 The estimated posterior distribution of group-specific additive genetic variances $\sigma_{G_r}^2$ in the models for wing length (top), body mass (middle) and tarsus length (bottom). Posterior variances for the different genetic groups are shown in different colours. Posteriors from genomic models have solid lines, whereas the pedigree-based model posteriors are shown with dotted lines.

local adaption to different habitats, and that the adaptive potential (i.e. V_A) of these possibly locally adapted phenotypic traits could depend on their group ancestry.

Genetic group models are not only attractive because they eliminate the source of bias in estimation of V_A caused by genetic substructures in the base population, but also because they help identify differences in V_A for admixed and/or genetically differentiated populations. Since higher V_A implies higher potential for rapid evolutionary change due to selection, such as adaptation to any environmental changes, the identification of differences is important especially in the light of the current rapid change in environments due to anthropogenic effects (Gienapp et al., 2017; Wood & Brodie III, 2016). However, the estimation of group-specific V_A obviously increases computational requirements and imposes higher demands on the data. The model is dramatically simplified if the assumption of differences in the groupspecific σ_G^2 for r = 1, ..., R is dropped, while still allowing for differences in the mean genetic (i.e. breeding) values γ_r . Thus, only a single, homogeneous V_{Δ} without admixture-induced bias is estimated (Wolak & Reid, 2017). Such a simplified model will only need to estimate the group-specific global ancestry proportions π_{ir} for each individual *i*, as given in Equation (5). In this simplified case, however, one does not need to make the computationally heavy detour via the inference of local ancestries of each allele, as we did here, but could directly estimate π_{ir} using faster methods (such as Raj et al., 2014).

A main difference in the methodology discussed here compared to Rio, Moreau, et al. (2020) is that we assume segregation variances to be negligible. As previously mentioned, segregation variances are expected to be negligible under the infinitesimal model, which is commonly assumed to hold for most complex traits, in particular in natural populations (Hill, 2012; Hill & Kirkpatrick, 2010). The inclusion of segregation variances would therefore unnecessarily increase the demands on computational power and data quality, and potentially prohibit the use of the model. To illustrate that omitting segregation terms is indeed not critically affecting the parameter estimates, we also fitted modified models which do include segregation variances to confirm that segregation variances are indeed negligible (Figure S2; Tables S1 and S2, Supporting Information S4.2).

Another interesting assumption of the genomic model is that it allows allele effects to depend on the local ancestry of the allele, not merely the variant. Such group-specific allele effects can, for example, result from the groups having different strengths of LD between SNPs and quantitative trait loci (QTL, Rio, Mary-Huard, et al., 2020). Among-group differences in strength and extent of LD can be caused by differences in effective population sizes leading to different levels of genetic drift, and by differences in local selection pressures on the trait in question. While we do not explicitly model LD between loci, we thus implicitly account for differences in LD across groups through the group-specific allele effects. Hagen et al. (2020) indeed found that the levels of LD in house sparrows were generally higher and remained higher over longer distances along the chromosomes on islands in our study system with smaller effective population sizes (generally outer islands, which recently went through a strong population bottleneck, see Baalsrud et al., 2014). Group-specific allele effects can also be caused by QTL having epistatic interactions with loci whose allele frequencies differ between groups (Rio, Mary-Huard, et al., 2020). Loci with group differences in allele frequencies are common in our study system (Figure S3, Supporting Information S4.3), as there is evidence for genetic differ entiation between the islands (Niskanen et al., 2020; Saatoglu et al., 2021). Allowing for group-specific allele effects may thus be justified, but further investigations would be required to confirm their presence in the study system.

An important prerequisite when using genetic group models is that there is at least some level of admixture between genetic groups. If the groups were isolated without any admixture, we might otherwise not be able to disentangle evolutionary processes from environmental effects (Hoffmann et al., 2017; Wood & Brodie III, 2016). Conversely, having purebred individuals present in the dataset is not a requirement in general. In fact, in a wild system with dispersal such as the one we consider, it is more realistic to assume all individuals as admixed, if we look far enough back in time. However, the decision to include purebreds must sometimes be made out of necessity. Here, for example, our choice to consider individuals with both parents missing in the pedigree as purebred increased the comparability of our results to those derived from the pedigree-based genetic group animal model. Additionally, the purebreds were needed to be used as reference panels in the local ancestry inference.

In a more general set-up, it might not always be obvious how the 'purebred' reference populations of the different genetic groups should be defined or selected, in particular when pedigree data are missing. Since individuals labelled as 'purebreds' will usually not actually be purebred in a strict sense, but are rather individuals with a mix of ancestries from various populations, the purebred (or reference) population can in principle be defined in any desired way that is relevant for the study at hand. As an example, subpopulations present (or rather, individuals captured) in certain areas the first year of a multi-year study can be used as reference panels in local ancestry estimation, which resembles what we did in the sparrow example. This approach requires previous knowledge about the system-for example, we knew from previous studies that the different islandgroup populations are somewhat genetically differentiated (Muff et al., 2019; Niskanen et al., 2020; Saatoglu et al., 2021). Alternatively, population structure estimation tools can detect the identity of genetic groups, global ancestry proportions and/or the number of groups (e.g. Kuismin et al., 2017; Raj et al., 2014), or even local ancestries directly (e.g. Utsunomiya et al., 2020). These methods do not generally rely on reference panels. When using a tool where only global ancestry proportions in a given group are detected, individuals with global ancestry higher than some threshold (e.g. 0.99) can be assigned as purebred in that group and serve as reference panels to find the local ancestries of the remaining (admixed) individuals

(Geza et al., 2019; Schubert et al., 2020). In short, use of the genomic genetic group model requires either previous knowledge about the population structure, or investigating it with additional tools.

In summary, we believe that the genomic genetic group animal model is a useful new tool that allows researchers to estimate and compare group-specific additive genetic values and variances in wild admixed populations. Thanks to the generalization from pedigreebased to genomic animal models, researchers are no longer restricted to using long-term data collected over many generations. To generate the data necessary to fit a genetic group animal model, it is now sufficient to sample phenotype and genotype information from individuals in an admixed natural population at a given time. In a conservation management perspective, the genomic genetic group approach thus opens up possibilities for researchers to estimate V_A within different populations to find the ones with greatest need of management to ensure sufficient adaptive potential, and also to examine how V_{Λ} in a population is affected by admixture (e.g. immigration), to actually quantify how management actions and/or natural dispersal will potentially aid in evolutionary rescue (Whiteley et al., 2015).

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CONFLICT OF INTEREST

The authors have no conflict of interest to declare.

AUTHORS' CONTRIBUTIONS

S.M., K.A. and H.J. conceived the idea for the study; K.A. developed and implemented the model, analysed the data and wrote the manuscript; S.M. helped develop and implement the model; H.J. provided the sparrow data; S.M. and H.J. provided methodological support, and participated in writing and editing the manuscript. All authors read and approved the final manuscript.

PEER REVIEW

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DATA AVAILABILITY STATEMENT

For the pedigree, phenotype and genotype data, use the Dryad Digital Repository https://doi.org/10.5061/dryad.m0cfxpp10 (Niskanen et al., 2020). For the natal island information, use the Dryad Digital Repository https://doi.org/10.5061/dryad.xgxd254h1 (Saatoglu et al., 2021).

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REFERENCES

- Aase, K. (2021). Genetic group animal models in the genomics era (Master's thesis). Norwegian University of Science and Technology.
- Andrews, K. R., De Barba, M., Russello, M. A., & Waits, L. P. (2018). Advances in using non-invasive, archival, and environmental samples for population genomic studies. In P. A. Hohenlohe & O. P. Rajora (Eds.), *Population genomics: Wildlife* (pp. 63–99). Springer International Publishing.
- Baalsrud, H. T., Sæther, B. E., Hagen, I. J., Myhre, A. M., Ringsby, T. H., Pärn, H., & Jensen, H. (2014). Effects of population characteristics and structure on estimates of effective population size in a house sparrow metapopulation. *Molecular Ecology*, 23, 2653–2668.
- Bérénos, C., Ellis, P. A., Pilkington, J. G., & Pemberton, J. M. (2014). Estimating quantitative genetic parameters in wild populations: A comparison of pedigree and genomic approaches. *Molecular Ecology*, 23, 3434–3451.
- Bowler, D. E., & Benton, T. G. (2005). Causes and consequences of animal dispersal strategies: Relating individual behaviour to spatial dynamics. *Biological Reviews*, 80, 205–225.
- Browning, B. L., Zhou, Y., & Browning, S. R. (2018). A one-penny imputed genome from next- generation reference panels. *The American Journal of Human Genetics*, 103, 338–348.
- Bulmer, M. G. (1971). The effect of selection on genetic variability. *The American Naturalist*, 105, 201–211.
- Carlson, S. M., Cunningham, C. J., & Westley, P. A. H. (2014). Evolutionary rescue in a changing world. Trends in Ecology & Evolution, 29, 521–530.
- Chang, C. C., Chow, C. C., Tellier, L. C. A. M., Vattikuti, S., Purcell, S. M., & Lee, J. J. (2015). Second-generation PLINK: Rising to the challenge of larger and richer datasets. *Gigascience*, 4, s13742–015.
- Charmantier, A., Garant, D., & Kruuk, L. E. B. (2014). *Quantitative genetics in the wild*. Oxford University Press.
- Dias-Alves, T., Mairal, J., & Blum, M. G. B. (2018). Loter: A software package to infer local ancestry for a wide range of species. *Molecular Biology and Evolution*, 35, 2318–2326.
- Evans, L. M., Tahmasbi, R., Vrieze, S. I., Abecasis, G. R., Das, S., Gazal, S., Bjelland, D. W., De Candia, T. R., Goddard, M. E., Neale, B. M., Yang, J., Visscher, P. M., Keller, M. C., & Haplotype Reference Consortium (2018). Comparison of methods that use whole genome data to estimate the heritability and genetic architecture of complex traits. *Nature Genetics*, 50, 737–745.
- Falconer, D. S., & Mackay, T. F. C. (1996). Introduction to quantitative genetics (4th ed.). Longman.
- Frankham, R., Ballou, J., & Briscoe, D. (2010). Introduction to conservation genetics (2nd ed.). Cambridge University Press.
- García-Cortés, L. A., & Toro, M. Á. (2006). Multibreed analysis by splitting the breeding values. *Genetics Selection Evolution*, *38*, 1–16.
- Gervais, L., Perrier, C., Bernard, M., Merlet, J., Pemberton, J. M., Pujol, B., & Quéméré, E. (2019). RAD-sequencing for estimating genomic relatedness matrix-based heritability in the wild: A case study in roe deer. *Molecular Ecology Resources*, 19, 1205–1217.
- Geza, E., Mugo, J., Mulder, N. J., Wonkam, A., Chimusa, E. R., & Mazandu, G. K. (2019). A comprehensive survey of models for dissecting local ancestry deconvolution in human genome. *Briefings in Bioinformatics*, 20, 1709–1724.
- Gienapp, P., Fior, S., Guillaume, F., Lasky, J. R., Sork, V. L., & Csilléry, K. (2017). Genomic quantitative genetics to study evolution in the wild. *Trends in Ecology & Evolution*, 32, 897–908.

- Grabenstein, K. C., & Taylor, S. A. (2018). Breaking barriers: Causes, consequences, and experimental utility of human-mediated hybridization. *Trends in Ecology & Evolution*, 33, 198–212.
- Gravel, S. (2012). Population genetics models of local ancestry. *Genetics*, 191, 607–619.
- Grueneberg, A., & de los Campos, G. (2019). BGData–A suite of R packages for genomic analysis with big data. G3: Genes, Genomes, Genetics, 9, 1377–1383.
- Hagen, I. J., Lien, S., Billing, A. M., Elgvin, T. O., Trier, C. N., Niskanen, A. K., Tarka, M., Slate, J., Sætre, G. P., & Jensen, H. (2020). A genome-wide linkage map for the house sparrow (*Passer domesticus*) provides insights into the evolutionary history of the avian genome. *Molecular Ecology Resources*, 20, 544–559.
- Hayes, B. J., Visscher, P. M., & Goddard, M. E. (2009). Increased accuracy of artificial selection by using the realized relationship matrix. *Genetics Research*, 91, 47–60.
- Henderson, C. R. (1984). Applications of linear models in animal breeding. University of Guelph Press.
- Hill, W. G. (2012). Quantitative genetics in the genomics era. Current Genomics, 13, 196–206.
- Hill, W. G., & Kirkpatrick, M. (2010). What animal breeding has taught us about evolution. Annual Review of Ecology, Evolution, and Systematics, 41, 1–19.
- Hill, W. G., & Weir, B. S. (2011). Variation in actual relationship as a consequence of Mendelian sampling and linkage. *Genetics Research*, 93, 47–64.
- Hoffmann, A. A., Sgrò, C. M., & Kristensen, T. N. (2017). Revisiting adaptive potential, population size, and conservation. *Trends in Ecology* & *Evolution*, 32, 506–517.
- Keller, L. F., Grant, P. R., Grant, B. R., & Petren, K. (2001). Heritability of morphological traits in Darwin's finches: Misidentified paternity and maternal effects. *Heredity*, 87, 325–336.
- Kruuk, L. E. B. (2004). Estimating genetic parameters in natural populations using the 'animal model'. *Philosophical Transactions of the Royal Society of London Series B: Biological Sciences*, 359, 873–890.
- Kruuk, L. E. B., & Hadfield, J. D. (2007). How to separate genetic and environmental causes of similarity between relatives. *Journal of Evolutionary Biology*, 20, 1890–1903.
- Kuismin, M. O., Ahlinder, J., & Sillanpää, M. J. (2017). CONE: Community oriented network estimation is a versatile framework for inferring population structure in large-scale sequencing data. G3: Genes, Genomes, Genetics, 7, 3359–3377.
- Legarra, A. (2016). Comparing estimates of genetic variance across different relationship models. *Theoretical Population Biology*, 107, 26–30.
- Lenormand, T. (2002). Gene flow and the limits to natural selection. Trends in Ecology & Evolution, 17, 183–189.
- Lo, L. L., Fernando, R. L., & Grossman, M. (1993). Covariance between relatives in multibreed populations: Additive model. *Theoretical and Applied Genetics*, 87, 423–430.
- Lynch, M., & Walsh, B. (1998). Genetics and analysis of quantitative traits (Vol. 1). Sinauer Associates Incorporated.
- Muff, S., Niskanen, A. K., Saatoglu, D., Keller, L. F., & Jensen, H. (2019). Animal models with group-specific additive genetic variances: Extending genetic group models. *Genetics Selection Evolution*, 51, 7.
- Niskanen, A. K., Billing, A. M., Holand, H., Hagen, I. J., Araya-Ajoy, Y. G., Husby, A., Rønning, B., Myhre, A. M., Ranke, P. S., Kvalnes, T., Pärn, H., Ringsby, T. H., Lien, S., Sæther, B. E., Muff, S., & Jensen, H. (2020). Consistent scaling of inbreeding depression in space and time in a house sparrow metapopulation. *Proceedings of the National Academy* of Sciences of the United States of America, 117, 14584–14592.
- Nyquist, W. E., & Baker, R. J. (1991). Estimation of heritability and prediction of selection response in plant populations. *Critical Reviews in Plant Sciences*, 10, 235–322.
- Ponzi, E., Keller, L. F., & Muff, S. (2019). The simulation extrapolation technique meets ecology and evolution: A general and intuitive

method to account for measurement error. *Methods in Ecology and Evolution*, 10, 1734–1748.

- Powell, J. E., Visscher, P. M., & Goddard, M. E. (2010). Reconciling the analysis of IBD and IBS in complex trait studies. *Nature Reviews Genetics*, 11, 800–805.
- Quaas, R. L. (1988). Additive genetic model with groups and relationships. Journal of Dairy Science, 71, 1338–1345.
- Raj, A., Stephens, M., & Pritchard, J. K. (2014). fastSTRUCTURE: Variational inference of population structure in large SNP data sets. *Genetics*, 197, 573–589.
- Ranke, P. S., Skjelseth, S., Hagen, I. J., Billing, A. M., Pedersen, A. A. B., Pärn, H., Ringsby, T. H., Sæther, B. E., & Jensen, H. (2020). Multigenerational genetic consequences of reinforcement in a bird metapopulation. *Conservation Genetics*, 21, 603–612.
- Reid, J. M., Arcese, P., Nietlisbach, P., Wolak, M. E., Muff, S., Dickel, L., & Keller, L. F. (2021). Immigration counter-acts local micro-evolution of a major fitness component: Migration-selection balance in freeliving song sparrows. *Evolution Letters*, 5, 48–60.
- Rio, S., Mary-Huard, T., Moreau, L., Bauland, C., Palaffre, C., Madur, D., Combes, V., & Charcosset, A. (2020). Disentangling group specific QTL allele effects from genetic background epistasis using admixed individuals in GWAS: An application to maize flowering. *PLoS Genetics*, 16, e1008241.
- Rio, S., Moreau, L., Charcosset, A., & Mary-Huard, T. (2020). Accounting for group-specific allele effects and admixture in genomic predictions: Theory and experimental evaluation in maize. *Genetics*, 216, 27–41.
- Ronce, O. (2007). How does it feel to be like a rolling stone? Ten questions about dispersal evolution. *Annual Review of Ecology, Evolution, and Systematics*, 38, 231–253.
- Rue, H., Martino, S., & Chopin, N. (2009). Approximate Bayesian inference for latent Gaussian models by using integrated nested Laplace approximations. *Journal of the Royal Statistical Society: Series B* (Statistical Methodology), 71, 319–392.
- Saatoglu, D., Niskanen, A. K., Kuismin, M. O., Ranke, P. S., Hagen, I. J., Araya-Ajoy, Y. G., Kvalnes, T., Pärn, H., Rønning, B., Ringsby, T. H., Sæther, B. E., Husby, A., Sillanpää, M. J., & Jensen, H. (2021). Dispersal in a house sparrow metapopulation: An integrative case study of genetic assignment calibrated with ecological data and pedigree information. *Molecular Ecology*, 30, 4740–4756.
- Schaeffer, L. R. (1991). CR Henderson: Contributions to predicting genetic merit. Journal of Dairy Science, 74, 4052–4066.
- Schubert, R., Andaleon, A., & Wheeler, H. E. (2020). Comparing local ancestry inference models in populations of two-and three-way admixture. *PeerJ*, 8, e10090.
- Själander, M., Jahre, M., Tufte, G., & Reissmann, N. (2019). EPIC: An energy-efficient, high-performance GPGPU computing research infrastructure.
- Slatkin, M., & Lande, R. (1994). Segregation variance after hybridization of isolated populations. *Genetics Research*, 64, 51–56.
- Speed, D., & Balding, D. J. (2015). Relatedness in the post-genomic era: Is it still useful? *Nature Reviews Genetics*, *16*, 33–44.
- Stanton-Geddes, J., Yoder, J. B., Briskine, R., Young, N. D., & Tiffin, P. (2013). Estimating heritability using genomic data. *Methods in Ecology and Evolution*, 4, 1151–1158.

- Tang, H., Peng, J., Wang, P., & Risch, N. J. (2005). Estimation of individual admixture: Analytical and study design considerations. Genetic Epidemiology: The Official Publication of the International Genetic Epidemiology Society, 28, 289–301.
- Toosi, A., Fernando, R. L., & Dekkers, J. C. M. (2010). Genomic selection in admixed and crossbred populations. *Journal of Animal Science*, 88, 32–46.
- Utsunomiya, Y. T., Milanesi, M., Barbato, M., Utsunomiya, A. T. H., Sölkner, J., Ajmone-Marsan, P., & Garcia, J. F. (2020). Unsupervised detection of ancestry tracks with the GHap R package. *Methods in Ecology and Evolution*, 11, 1448–1454.
- VanRaden, P. M. (2008). Efficient methods to compute genomic predictions. Journal of Dairy Science, 91, 4414–4423.
- Walsh, B., & Lynch, M. (2018). Evolution and selection of quantitative traits. Oxford University Press.
- Wang, B., Sverdlov, S., & Thompson, E. A. (2017). Efficient estimation of realized kinship from single nucleotide polymorphism genotypes. *Genetics*, 205, 1063–1078.
- Weir, B. S., Anderson, A. D., & Hepler, A. B. (2006). Genetic relatedness analysis: Modern data and new challenges. *Nature Reviews Genetics*, 7, 771–780.
- Whiteley, A. R., Fitzpatrick, S. W., Funk, W. C., & Tallmon, D. A. (2015). Genetic rescue to the rescue. *Trends in Ecology & Evolution*, 30, 42–49.
- Wilson, A. J., Reale, D., Clements, M. N., Morrissey, M. M., Postma, E., Walling, C. A., Kruuk, L. E. B., & Nussey, D. H. (2010). An ecologist's guide to the animal model. *Journal of Animal Ecology*, 79, 13–26.
- Wolak, M. E., & Reid, J. M. (2017). Accounting for genetic differences among unknown parents in microevolutionary studies: How to include genetic groups in quantitative genetic animal models. *Journal* of Animal Ecology, 86, 7–20.
- Wood, C. W., & Brodie, E. D., III. (2016). Evolutionary response when selection and genetic variation covary across environments. *Ecology Letters*, 19, 1189–1200.
- Wright, S. (1922). Coefficients of inbreeding and relationship. The American Naturalist, 56, 330–338.
- Yang, J., Zeng, J., Goddard, M. E., Wray, N. R., & Visscher, P. M. (2017). Concepts, estimation and interpretation of SNP-based heritability. *Nature Genetics*, 49, 1304–1310.

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