# Tag use to monitor fish behaviour in aquaculture: a review of benefits, problems and solutions 

Macaulay $\mathrm{G}^{* 1}$, Warren-Myers $\mathrm{F}^{1}$, Barrett LT ${ }^{1}$, Oppedal F ${ }^{2}$, Føre $\mathrm{M}^{3}$, Dempster $\mathrm{T}^{1}$<br>${ }^{1}$ Sustainable Aquaculture Laboratory -Temperate and Tropical (SALTT), School of BioSciences, University of Melbourne, Victoria 3010, Australia<br>${ }^{2}$ Institute of Marine Research, Matre Aquaculture Research Station, 5984 Matredal, Norway<br>${ }^{3}$ NTNU Department of Engineering Cybernetics, 7491 Trondheim, Norway

*Corresponding author gmacaulay@student.unimelb.edu.au

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#### Abstract

A variety of tagging techniques are now available to monitor fish behaviour, physiology, and their environmental experience. Tagging is frequently used in aquaculture research to monitor free-swimming individuals within farmed populations. However, for information gathered from tagged fish to be representative of farmed populations, tagging must not fundamentally affect fish behaviour, physiology, or survival. Here, we systematically review studies that used tags to monitor farmed fish behaviour and test factors that affect tag retrieval and tag-related mortality. Most studies using tags assessed movement and


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swimming behaviour in salmonids, predominantly in Europe and North America. Mortality of tagged fish was 10 times higher in sea-cages (mean $=25 \%$, range $=0-61.5 \%, \mathrm{n}=22$ studies) than in tanks (mean $=2.5 \%$, range $=0-17 \%, n=23$ studies), while mortality of tagged fish in sea-cages was markedly higher in longer trials (from 4\% in single day trials to $36 \%$ after 100 days). Higher-than-usual mortality rates among tagged fish, together with largely unknown sub-lethal effects on behaviour, should caution against using tagging studies to make decisions related to farm management. Moreover, key metrics such as mortality rates of tagged and untagged fish or evidence of sublethal effects are often unreported. We make several recommendations to improve future tagging studies and increase transparency in reporting. A greater insight into the causes of tagged fish mortality in sea-cages is required to secure animal welfare and data validity in studies that use tags to assess fish behaviour in aquaculture.

## Introduction

Tagging technology has revolutionised the way researchers study animal behaviour. Now, with the miniaturisation of tags, data storage capability with efficient flash memory, satellite detectability, extended battery life and the capacity for telemetry tags to relay environmental, behavioural, and physiological data in real-time, information about individual animal behaviour has never been so easily accessible (Hussey et al. 2015). For fish especially, the use of tags has been essential in gleaning information about these otherwise difficult to observe animals (Lucas \& Baras 2000). Tags developed for studying wild fish behaviour are largely created with the purpose of long-term monitoring of free-swimming fish, with the goal of conservation or fisheries management (Crossin et al. 2017). For instance, tags have

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revealed important insights into mortality rates of wild fish (Paulik 1963; Linard \& Matley 2020) and the migration and movement behaviour of bluefin tuna (Block et al. 2005), basking sharks (Dewar et al. 2018), manta rays (Braun et al. 2015) and saithe (Uglem et al. 2009). These tags, originally designed for studying wild fish ecology, have now transitioned into the world of farmed fish. Wild fish tagging studies can provide valuable information regarding tagging fish in aquaculture, for instance, many researchers discount the first several days of data post-tagging in the wild to allow time for fish behaviour to return to normal. While wild fish studies can inform tagging practices in aquaculture, the environments of wild and farmed fish differ substantially.

The finfish aquaculture industry has grown significantly over the past three decades. The 'more farmed than captured' milestone was reached for freshwater fish in 1986, diadromous fish in 1997, and the production of marine fishes continues to approach wild capture (FAO 2020). In 2014, fish for consumption raised in aquaculture surpassed wild fisheries (FAO, 2020) and by 2030 aquaculture is predicted to account for two thirds of all fish consumed (WorldBank, 2020). Despite the industry's swift expansion and immense production output, much is still unknown about the behaviour of many farmed fish species. Industrial aquaculture is still in its infancy compared to industrial terrestrial livestock farming, with many more fish species being farmed (Huntingford et al. 2011) and many of these being not fully domesticated (Teletchea \& Fontaine 2014). As such, there is considerable room for improvement in management and animal husbandry to ensure finfish production is environmentally sustainable and ethically sound. Both public and industry concern for farmed fish welfare is increasing (Noble et al. 2018; Bovenkerk \& Meijboom 2020). Securing fish welfare can benefit both fish and farmer; as healthy fish have better growth (Huntingford \&

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Kadri 2014), farmers will spend less on treatments or stock loss associated with fish in poor condition and the shelf life of the final product is longer in less stressed fish. Fish behaviour and welfare are inextricably linked, and changes in behavioural patterns such as feeding, responses to environmental conditions or movement inside the production environment are often indicative of the welfare status of fish. Therefore, improving aquaculture management requires a detailed knowledge of how farmed fish behave.

With hundreds of thousands to millions of individuals cultured at high stocking densities in one farm, effectively tracking fish behaviour in aquaculture is inherently difficult (Føre et al. 2018). Typically, farmers will use their own personal experience and visual assessment of fish from the surface or subsurface cameras to monitor how a portion of their stock behave (Skøien 2017). Using data from environmental sensors as a surrogate for behavioural monitoring is also common (Andrewartha et al. 2015), with the view that if environmental conditions are typical, so should be the behaviour and welfare of the animals. In addition to traditional methods, different technologies for monitoring group behaviour in aquaculture exist. For instance, sonar can be used in tandem with environmental sensors to provide an instantaneous view of how Atlantic salmon (Salmo salar) distribute themselves vertically within sea-cages in response to changes in the cage environment (Oppedal et al. 2011a). Other approaches based on optics coupled with machine vision techniques can also provide insight into cage dynamics such as feeding activity (Måløy et al. 2019). While such technology is available, it is not commonly used in the industry to monitor general behaviour but is gaining pace as a tool in salmon farming to control and monitor feeding (e.g. www.cageeye.com). However, recent advancements in tagging technology could transform

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the way behaviour is monitored in aquaculture as it enables acquiring data histories on the individual level, something optical and hydroacoustic methods cannot do.

Individual based behavioural monitoring has been explored in terrestrial animal agriculture as part of the 'Precision Livestock Farming' (PLF) framework. Farmers implementing PLF use various technologies to continuously monitor the behaviour of individual animals in real-time, so they can take swift action when an animal welfare issue occurs (Berckmans 2014). For example, accelerometers attached to dairy cows monitor the amount of time cows spend lying down; too much time spent lying down indicates lameness and has significant consequences for welfare (Darr \& Epperson 2009). This ‘smart-farming' approach has recently been adapted to aquaculture (Føre et al. 2018). For instance, farmed oysters have been equipped with tags that monitor their cardiac activity (Andrewartha et al. 2015). These tagged individuals act as 'sentinels', much like canaries in a coal-mine, with their activity levels representing those of the rest of the population. A similar approach has been advocated for finfish aquaculture (Føre et al. 2018) with behavioural monitoring of individuals to inform farm management. Sampling individual fish using tags permits the observation of how individuals behave in their environment within the greater population (Føre et al. 2017). Tagging individual 'sentinel fish', could provide welfare related information useful to monitor and manage feeding, parasites and disease, and environmental conditions. However, while using tags can provide useful information, there are challenges.

A primary assumption of this approach is that the behaviour of tagged individuals is representative of untagged individuals. To ensure that this assumption is not compromised, tags must not alter the behaviour or physiology of the tagged individuals to the extent that they no longer represent those of the whole population (Perry et al. 2001). Tagging a fish also

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requires catching, handling, anaesthesia, tag attachment, and recovery, all of which may change behaviours after release leading to erroneous data (Jepsen et al. 2015). In aquaculture, incorrect data could lead to flawed management decisions that negatively impact fish welfare and ultimately production. Furthermore, a tag or tagging procedure that leads to negative sublethal effects creates poor fish welfare. Hence, the effects of tags and tagging on farmed fish should be first understood, and then minimised wherever possible. The 3Rs guidelines for the human use of animals for scientific research include the following: (1) Replacement, which encourages removing the use of animals entirely when possible, (2) Reduction, which limits the number of animals used for experimentation to a minimum, and (3) Refinement which ensures that experimental methods are sophisticated enough so that experimental animals undergo the least amount of suffering possible (Russell \& Burch 1959). If tags cause suffering, they should be replaced with other technologies to monitor behaviour. If different methodologies are not appropriate, then the number of tagged fish used should be reduced to the minimum number that still provides relevant results. Finally, tagging methodologies should be refined to minimise potential suffering.

There is a broad body of literature on the methods and challenges of tagging wild fish (see reviews by Cooke et al. (2011), Thorstad et al. (2013) and Brownscombe et al. (2019)) which can partly inform best use practices in aquaculture. However, challenges associated with tagging fish in the wild differ to those on a farm. As farmed fish are typically stocked at high densities (Fernö et al. 2006), rates of external tag loss may be higher due to interactions with conspecifics or the cage structure. High population densities in spatially static enclosures also promote the swift spread of disease and parasites (Murray \& Peeler 2005). Infection of tagging-induced wounds may therefore be more likely to occur in a high-density environment

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and may lead to altered behaviour in the tagged fish, rendering it unrepresentative of the farmed population. In addition, common husbandry processes or environmental conditions on farms can induce stress that can suppress the immune system and cause altered behaviour or elevated mortality of tagged individuals if they are already under additional stress from tagging (Sykes et al. 2012; Stehfest et al. 2017). Aquaculture also forces spatially and seasonally restricted environments on fish, of which may include extreme experiences that normally would be avoided in the wild.

Retrieving tags in aquaculture seems simpler than retrieving tags from the wild, as farmed fish are in confinement. However, if only a few individuals are tagged out of potentially hundreds of thousands of fish in an enclosure, recovering these tags during harvest can be difficult as hundreds of tons of fish must be checked manually or via sensing methods at harvest point or in processing facilities (Føre et al. 2017). This is especially true for sea-cage aquaculture, which is conducted in floating cages attached to moorings in the sea that are open to the wider marine environment and the farmer has little control over environmental conditions. In tank aquaculture, which is conducted on-land in closed containment, the farmer can heavily control environmental conditions and may be more likely to retrieve lost tags as they will not fall out of net cages as they might in sea-cage aquaculture.

Tag recovery is arguably more important in aquaculture than in wild fish studies, as an unaccounted-for tag could render an entire batch of fish unsellable, to avoid customers potentially encountering tag components (plastic, metal, fibreglass) in their food. Furthermore, deploying many tagged individuals in one production unit may be unfeasible depending on tag type. Because acoustic telemetry tags (originally designed for wild fish), were not designed to be used in restricted spaces, a concern with using these tags in

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aquaculture is that too many acoustic tags using the same frequency may cause signal interference or 'tag clashes' (Kolarevic et al. 2016; Martos-Sitcha et al. 2019). However, this potential issue can be partly remedied through more advanced signal processing in the receiver unit, optimising receiver deployment and designating unique acoustic frequencies for each cage or tank (Føre et al. 2017). Irrespective of these challenges, tags are powerful tools that are becoming more commonly used in aquaculture research to obtain important information about individual farmed fish behaviour.

The unique possibilities and challenges of using fish tags to deliver information on behavioural attributes important for production in aquaculture have received little attention since an early review by Baras and Lagardère (1995). Here, we summarise the different types of tags used to investigate behaviours of fish in farming contexts, assess trends in reporting of data, and analyse factors that affect tagged fish mortality and tag retrieval in fish farms. Building upon our results, we make recommendations about the use of tags as an information gathering tool in aquaculture.

## Methods

We searched the Web of Science database in May 2020 for publications (from all years) that used tags to study fish behaviour in aquaculture using the following search terms: (aquaculture OR fish farm) AND fish AND behav* AND (tag* OR telemetry OR bio-telemetry OR biotelemetry OR satellite OR bio-log* OR biolog* OR PIT OR remote sensing OR acoustic telemetry OR depth OR movement OR swimming depth OR vertical distribution OR thermoregulation OR habituation). This initial search returned 573 results. These papers were manually checked by title and abstract and eliminated if the subject matter was clearly

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unrelated to the search terms, or otherwise after reading the full text. Additional publications missed by our initial search were discovered by examining reference lists of appropriate articles and through additional investigative searches using Google Scholar and Scopus databases up until August 2020.

Several papers included multiple tagging studies (e.g. multiple species, tag types, cage/tank sizes, cage/tank designs and stocking densities). The data in these papers was split accordingly into separate studies, so the effect of these different aspects of the study on fish mortality could be evaluated (Table S1). Studies that gave the number of untagged fish, density, biomass or mentioned the presence of untagged fish were classed as studies where tagged fish were co-located with untagged conspecifics.

## Statistics

Beta regression models (betareg package: Cribari-Neto \& Zeileis (2009)) in R (version 4.0.2, R Core Team, 2020) were used to test the relationship between two response variables (TagMortality: the mortality rate of tagged fish; TagRetrieval: the proportion of tags retrieved) and predictor variables (see Table 1 for list). Both responses were initially calculated at the level of experimental units (tanks or cages), and then aggregated to study level by computing the mean of all comparable experimental units. Beta regression models were selected after examination of response variable distributions and residuals using the DHARMa package (Hartig 2016). Beta regression models cannot fit proportions of exactly 0 or 1 , so where these occurred, we assigned dummy values of 0.0001 or 0.9999 , respectively. Where values were reported as a range (e.g. 'stocking density was $22-25 \mathrm{~kg} / \mathrm{m}^{3}$ ) the midpoint of the range was used for the analysis. Studies were proportionally weighted within the model according to the number of experimental units that contributed to the study mean value (i.e. number of tanks

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or cages). Many potentially important predictor variables were not ubiquitously reported by tagging studies. We did not include the following variables due to these being under-reported and decreasing the model sample size: fish weight, tag weight, tag volume, tag length.

Predictor variables included in models are listed in Table 1.We fitted a set of models for each response variable, each designed to maximise the sample size for a given predictor variable of interest. Final model specifications are outlined in Table 2. Analysis of Deviance (Type II Wald $\mathrm{X}^{2}$ test) was used to test the significance of predictor variables (Anova function, car package: Fox \& Weisberg 2019). To test the sensitivity of the findings to particularly influential studies, we fitted models without 2 studies that were outliers in their long duration (Sykes et al. (2012) ~1100 days; Jurajda et al. (2016) ~730 days). We also fitted models with and without 2 studies that used submerged cages (and reported on tagged fish mortality) (Korsøen et al. 2012; Wright et al. 2019) to test the sensitivity of findings to submerged cages, which can have high levels of mortality. The final models excluded the two long studies and the two submerged cage studies due to the model being sensitive to their influence. Tanks studies were also excluded in the final models due to the low mortality rates in these environments. Duration of tank and sea-cage studies were of similar length.

The ggeffects package (Lüdecke, 2018) was used to predict tagged fish mortality with marginal effects from sea-cage volume.

## 1. Results

The literature search yielded a total of 83 studies (from 49 research articles) that used tags to monitor behaviour of fish in aquaculture environments, with the first study that used tags to monitor farmed fish behaviour being conducted in 1989 (Gusar et al. 1989).

### 1.1 Study location, species, and experimental units

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Most studies were conducted in Europe and North America (Norway (43\%), USA (12\%), Canada (10\%), Fig 1). Sea-cages (of varying sizes) were the most common study environment (61\%) followed by tanks (33\%) and ponds (6\%). Atlantic salmon (Salmo salar) was the most studied species, followed by rainbow trout (Oncorhynchus mykiss) and Atlantic cod (Gadus morhua) (Fig. 2). It was most common for all tagged fish to be monitored in a single cage, tank, or pond (59\% of all studies in a single experimental unit, $12 \%$ in 2 experimental units, $29 \%$ in 3 or more experimental units).

### 1.2 Study focus

Assessments of movement and swimming behaviour (3D spatial location of fish) were the most common study focus. Response to increasing stocking density and measuring responses to various types of stress were the second and third most prevalent study focuses, respectively. Other areas of focus, addressed by a smaller number of studies, included: feeding behaviour, respiration and swimming activity (e.g. acceleration, tail beat frequency), response to cage designs, parasites, aggression, environmental conditions, surface activity and coping styles (Fig. 3).

### 1.3 Tag types

Transmitter tags were the most widely used to study behaviour in aquaculture. These transmitter tags provided the following types of data: location of fish in 3-dimensions, swimming trajectory, acceleration, pressure, or muscular activity (Fig. 3). DSTs (Data Storage Tags) were the second most widely used tag, recording a variety of types of data including swimming depth, body temperature, heart rate, gastro-intestinal blood flow and acceleration. 24\% of studies that internally implanted telemetry, DST or PIT tags also externally attached non-electronic external marker tags (*External marker tags = coloured tag, Floy, T-bar,

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numbered Peterson's discs + opercular loop tag FT4) to the same fish to allow researchers to identify tagged individuals for the purpose of tag retrieval or monitoring tagged individuals via video or direct observation. Three studies used only external marker tags as a method to observe individuals for data collection and did not use an accompanying electronic recording tag (Laursen et al. 2011; Bui et al. 2013; Staven et al. 2019). PIT tags were used for identification of individuals, but also used in combination with PIT antennae to register activity of individuals relating to a behaviour (e.g. surface activity, size-dependent swimming).

### 1.4 Tag attachment

$60 \%$ of studies internally attached electronic tags with surgical with surgical implantation into the abdominal cavity (Wright et al. 2019), or the dorsal musculature behind the head (Kristiansen et al. 2004; Nilsson et al. 2013). 19\% of studies had internally attached tags with a non-electronic external marker tag also (e.g. Macaulay et al. 2020). 20\% of studies externally attached tags to the musculature adjacent to the dorsal fin. In final models, tag attachment was not a significant predictor of either tagged fish mortality or tag retrieval (Table 2).

### 1.5 Fish

### 1.5.1 Untagged conspecifics

Of the 52 studies that mentioned tagged fish being co-located with untagged conspecifics only 4 reported a measure of mortality of the untagged fish (e.g. Føre et al. (2017) reported "less than 5\%"; Wright et al. (2018) reported "less than 1\%"; Leclercq et al. (2018) reported "no significant mortalities occurred"; Macaulay et al. (2020a) reported only total mortality). This means a baseline mortality rate for untagged (control) fish was lacking throughout majority of studies, making it impossible to compare mortality rates of tagged and untagged fish.

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For the studies that presented data on untagged conspecifics (and reported sample sizes per replicate), mean percentage of the entire population that were tagged per unit replicate was $4.9 \%$, with a median of $0.58 \%$. This large difference between the mean and median values reflects the influence of relatively few studies that had a high percentage of tagged fish, where most or all fish were tagged due to the small number of fish used.

### 1.6 Tagging effects

### 1.6.1 Sublethal effects

Reporting on the physical effects of tags (e.g. skin/scale loss, necrosis, bacterial infection, fouling) was uncommon and brief when done so. Examples of how these effects were reported in externally tagged fish include: Sykes et al. (2012), who reported cases of ripped opercula and chronic skin lesions around the tag attachment site in FT4 opercular loop tagged fish, and reduced growth rates in externally tagged fish compared to PIT tagged fish. Ferrer et al. (2020) reported no signs of haemorrhaging or tissue damage in their externally tagged fish. Examples of how physical effects were reported in internally tagged fish include: Bauer and Schlott (2004) reported inflammation surrounding the incision in tagged carp but no 'severe' tissue necrosis after 121 days. Conversely, Rillahan et al. (2009) reported no incidences of infection, Føre et al. (2017) found no obvious visual signs of impaired health in tagged individuals and Svendsen et al. (2020) found no signs of poor wound healing or infection after inspection of tagged fish at the end of the study duration.

Behavioural effects of the tag, such as tag influence on swimming performance (McFarlane et al. 2004), were more commonly reported on than physical sublethal effects, but were not reported in great detail when done so. Pilot trials conducted in tanks were used in some studies to determine whether the tag had an effect of fish behaviour (e.g. Rillahan et al. (2009)

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internal attachment; Stehfest et al. (2017) external attachment), while others kept the fish in small containers for longer recovery times prior to deployment in larger sea-cages or ponds to monitor behaviour post-tagging (e.g. Jurajda et al. (2016) internal attachement). Some studies reported normal tagged fish growth or food consumption to indicate that tagging had no significant effect on the fish or results; e.g. Rillahan et al. (2009) compared behaviour and feeding of cod before and after internal tag implantation in tanks; Kolarevic et al. (2016) stated all individuals increased their body weight during this study (after being internally tagged with transmitters and externally tagged with marker tags); Føre et al. (2017) found no obvious visual signs of impaired health or lowered growth on the internally tagged individuals identified at slaughter. Korsøen et al. (2012), however, reported that internally DST tagged salmon had significantly lower growth in submerged cages compared to control cages with surface access, and that submerged tagged fish exhibited irregular swimming patterns for several days.
$16 \%$ of studies mentioned the 'tag weight to fish body weight ratio' rule of thumb that states tags should not weigh more than $2 \%$ of the fish's weight in air. Of these studies, tag weights ranged from $0.2-3 \%$ of the fish's weight in air (Table S1). Recovery time from tag attachment varied widely, from no recovery period (fish placed straight back into experimental cage/tank after tag attachment) to up to 7 days in a recovery area (Table S1).

### 1.6.2 Mortality

Mean tagged fish mortality across a range of trial durations was 10 times higher in sea-cages (mean $\pm$ SE $=25 \% \pm 4.07$, range $=0-61.5 \%, n=22$ studies), compared to tank studies (mean $\pm$ $S E=2.5 \% \pm 1.15$, range $=0-17 \%, n=23$ studies). While mortality was variable in sea cages there was a clear positive correlation between study duration and tagged fish mortality; the

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longer the study, the greater the mortality (Fig. 4). Model predictions (TagMortProp ~ UnitVol_m3 + StudyDays, weights = No.treatrep) forecast 3\% mortality on day 1 increasing to a predicted $28 \%$ mortality on day 100 in for a sea-cage volume of $2000 \mathrm{~m}^{3}$ and up to a predicted $36 \%$ mortality on day 100 in a $5000 \mathrm{~m}^{3}$ sea-cage..

33 \% of studies did not explicitly report mortality of tagged fish, or only stated there was 'some mortality' but did not give numbers. Only studies that clearly indicated mortality were included in the analyses.

We found no correlation between tag volume, tag weight, or tag:body weight ratio on fish mortality. However, as these tag attributes were not widely reported throughout the studies, the sample size we assessed was small.

### 1.7 Tag retrieval

Tank data was left out of the final model for tag retrieval, as there was $100 \%$ tag retrieval in tank studies. Study duration was the only significant predictor of number of tags retrieved in sea-cages (Table 2). There was a negative relationship between tags retrieved and study duration in sea-cages; the longer the study the less tags were successfully recaptured.

## 2. Discussion

Our analysis revealed significantly higher rates of tagging-related mortality occurred in seacages compared to tanks, and that this effect increased with study duration (Fig. 4). Fewer tags were successfully retrieved at the completion of longer-term studies in sea-cages. These relationships did not occur in tank environments, and therefore reveal a disparity in outcomes when tags are used to record fish behaviour in tanks versus sea-cages. This has clear implications for studies aiming to obtain industry-relevant data on fish behaviour. Few studies

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compared tagged fish mortality to untagged fish mortality. Deprived of baseline mortality rates, we cannot truly determine to what extent tagging influences survival.

### 2.1 Tagging effects

### 2.1.1 Mortality

The discovery that mortality rates were $10 \times$ higher for tagged fish in sea-cages compared to tanks highlights fundamentally different outcomes for fish in these two study environments. This result has several implications for researchers, including: 1) results obtained in tanks may not be generalisable to sea-cage environments, 2) high mortality rates in sea cages experiments calls into question how representative tagged fish are of untagged fish, and 3) with such high mortality rates in sea-cages, tagging in this environment creates significant animal welfare issues. Why such high rates of mortality of tagged fish occur in sea-cage studies, and why it differs from tank studies is unclear. However, experiments conducted in small tanks are typically done under more controlled culture conditions, where researchers can largely prevent poor water quality and adverse events. In comparison, experiments conducted in extensive outdoor systems, such as sea-cages, create exposure to depth and seasonally variable and potentially unfavourable environmental conditions (temperature, salinity, dissolved oxygen, current, pollutants and pathogens, all of which can affect fish behaviour, welfare, and survival). Data first obtained from controlled tank experiments must be ground-truthed at industrial scale if it is to have commercial relevance.

Though mortality was variable in sea-cages between different studies, more tagged fish died as study length increased; the model fit indicated that 36\% of tagged fish died by trial day 100 in sea-cage with a volume of $2000 \mathrm{~m}^{3}$. This mortality rate is likely an order of magnitude higher than the average mortality rate experienced by fish in sea cages across 100 days. As an

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example, Atlantic salmon mortality over a three-month period would be approximately $3 \%$ (Stien et al. 2019). Ideally, tagged fish mortality should be within the bounds of the normal mortality rates of untagged fish in the cages. Beyond that, representativeness of the data delivered from tagged fish is questionable. However, it may be argued that the surviving fish have coped after surgery and represent normal behaviour, while the mortalities were not. This is demonstrated in a recent DST heart rate study fish with moribund fish displaying deviating results (Hvas et al. 2020b).

Mortality on untagged fish in tagging studies was largely underreported; only 2 of the 49 research articles we reviewed reported the percentage of untagged fish that died over the duration of their studies. Føre et al. (2017) reported less than 5\% mortality of the untagged fish population, while average tagged fish mortality across both cages and all tag types (acoustic transmitters and DSTs) was $19 \%$. Clearly, there is a large difference between untagged and tagged fish mortality in this study. Wright et al. (2019) reported mean mortality of untagged fish was $0.4 \%$ in standard cages and $0 \%$ of tagged fish. However, mean untagged fish mortality across both fresh and seawater snorkels was $0.5 \%$, while mean tagged fish mortality in these submerged cages was $30 \%$. This stark difference between tagged fish and untagged fish mortality in submerged cages highlights the risks of submerging tagged physostomous fish and indicates that adding additional stressors to tagged fish should be avoided, unless measuring stress responses is the purpose of the study. Without a baseline mortality for untagged fish, it is impossible to compare mortality rates of tagged and untagged fish. These results highlight an urgent need to better understand tag-related mortality in sea cage environments through specific testing of all aspects that could contribute to poor outcomes.

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While tags at present have only been used as research tools, the ambition to use them for 'sentinel fish' for real-time monitoring of behaviour to inform management decisions in commercial environments remains (e.g. Føre et al. 2018). The high rate of mortality in seacage studies challenges the achievability of this concept given the present technology level, both from a welfare and data accuracy/ representativity perspective. Currently, there are issues of an ethical and welfare nature when it comes to tag use in sea-cage finfish aquaculture. An experience that causes elevated rates of mortality is also an experience that causes elevated levels of stress and suffering that ends in mortality. Tagging methods in aquaculture therefore must improve the survival rates of tagged fish to minimise the issue of animal welfare. Future use of data from tagged sub-populations of fish in management decisions related to farm management requires that we first know and understand the causes of mortality and sub-lethal effects of tags on fish behaviour.

### 2.1.2 Sublethal physical effects

Although reports of sublethal physical effects were uncommon, this does not necessarily mean these effects were not present, but rather may be untested or under-reported. There has been extensive research into wound closure methods following tag insertion for internal attachment in wild fish telemetry. Common sublethal physical effects of following tag insertion include necrosis and inflammation (Lowartz et al. 1999; Wagner et al. 2000; Wagner \& Cooke 2005). Despite Sykes et al. (2012) detailing cases of ripped opercular and chronic skin lesions around the tag attachment site in FT4 opercular loop tagged fish, the focus of this study was to explicitly test tagging effects for this tag type. Researchers that are using tags to get answers on behavioural questions may not be as concerned with necrosis or inflammation at the tagging site if the fish's behaviour appears normal throughout the study. Nevertheless,

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because an injured or stressed fish may behave differently in the long term and can be more vulnerable to infection due to supressed immune function (Wagner \& Cooke 2005), any sublethal physical effects found on tagged fish should be stated plainly. Tagging technology is continually evolving to increase the capacity for gathering data, whilst simultaneously reducing the burden tags have on their individual bearer. Using smaller tags for future tagging trials will probably contribute to reduce the invasiveness of the procedure, and recent experiments have demonstrated the application of tags weighing as little as 600 mg (MartosSitcha et al. 2019).

### 2.1.3 Sublethal behavioural effects

If tags alter the behaviour of tagged fish, then the validity of any data provided by the tag is compromised as is the welfare of the tagged fish. Both Korsøen et al. (2012) and Wright et al. (2019) studied the behaviour of salmon in surface modified cages using DSTs. Korsøen et al. (2012) reported that submerged tagged salmon swam irregularly for several days, while Wright et al. (2019) found that DST tagged salmon in depth-modified snorkel cages experienced a $38 \%$ mortality rate compared to $0 \%$ for tagged individuals in a standard seacage. Salmon are a physostomous fish, meaning they require surface access to refill their swim bladder and regulate buoyancy (Fange 1953). In both studies, the added weight of the tag would reduce the salmon's maximum neutral buoyancy depth, reducing the capacity of the tagged fish to cope in these deep cages with restricted surface access (Macaulay et al. 2020b). Similarly, Perry et al. (2001) found that tagged Chinook salmon (Oncorhynchus tshawytscha) swam shallower than untagged conspecifics. In studies on wild salmonid smolts, concerns around tag weight negatively affecting fish lead to the largest individuals being chosen for tagging (Newton et al. 2016). Because fish size influences smolt survival, initially choosing only

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the largest fish for tagging biases results and leads to erroneous representations of true behaviour and survival (Deng et al. 2015). Stehfest et al. (2017) asserted that ‘only tagged fish that fully regained balance and responded to manual agitation after recovery were released back into the cage' in their study, although they did not report how many fish were unable to regain balance fully, or comment on whether the fish that regained balance were truly representative of the untagged population. Clearly, tags can influence buoyancy in fish, and so affect their behaviour and survival rate, which has inherent consequences for the validity of the data derived from tags as well as the welfare status of the tagged fish.

### 2.1.4 Tag:fish weight rule

Through our literature search, 9 research articles mentioned the tag:fish weight ratio. This general rule-of-thumb, originating from wild fish telemetry studies in the 1980s (Winter 1983), states that a tag should not weigh more than $2 \%$ of the fish's weight in air, to reduce negative impacts of tag weight on buoyancy, balance and swimming ability (Adams et al. 1998). Researchers often state this rule to imply that the tagging procedure did not affect the findings of their study. However, there are arguments against widely employing this rule across fish taxa (Jepsen et al. 2005; Cooke et al. 2011). For instance, tags weighing up to 12.7$14 \%$ of Atlantic salmon smolt and post-smolt weight have been used for migration studies, with no reported negative effects on survival (Lefèvre et al. 2013; Newton et al. 2016). However, other studies indicate that higher tag weight to body weight ratios reduce fish growth (juvenile Atlantic salmon: Larsen et al. (2013); lake sturgeon Acipenser fulvescens: Sutton and Benson (2003)). Nonetheless, tag weight alone is a poor scale on which to base expectations of the effect of tags on fish behaviour and physiology, as two tags that have identical weights in air can have very different densities, affecting tagged fish buoyancy

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accordingly (Brown et al. 1999). Of the studies found through our systematic review, $40 \%$ reported tag weight in air, $17 \%$ reported tag weight in water, and $56 \%$ reported tag length and diameter, such that tag volume could be calculated (Table S1). Tag volume is also important. An internally implanted tag that is neutrally buoyant but has a large volume will take up more space in the body cavity, pressing on vital organs, potentially also sutures, and could reduce swim bladder capacity (Macaulay et al. 2020b). A large neutrally buoyant tag that is externally attached could cause significant drag or epidermal damage (Brown et al. 1999; Jepsen et al. 2015). In addition, measuring fish weight in water (or fish body density) is important. As fish weigh much less in water, basing tag weight on fish weight in air may still cause negative effects on fish buoyancy (Macaulay et al. 2020b). Evidently, the tag:fish weight rule should be used with caution. Other more relevant tag and fish measurements discussed above will likely give a better indication of the effects of the tag's physical characteristics on fish physiology and behaviour.

### 2.2 Study focus

Because tagging is a rather invasive means to gather behavioural data, other methods should be considered before tagging is chosen, if they can deliver suitable data to answer the research question. The first principle of the 3Rs guidelines for the human use of animals for scientific research, is Replacement, which encourages removing the use of animals entirely when possible (Russell \& Burch 1959). This principle should be the foundation of tag use, especially in aquaculture where tagged fish mortality rates are presently too high (Fig. 3). As Brownscombe et al. (2019) states, 'studies that involve tagging animals simply for the sake of tagging' should be avoided. Most studies used telemetry tags to study movement and swimming behaviour, highlighting that tags are frequently used to track how fish use space in

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fish farms. Many of these studies simply used tags to track the swimming depth of individual tagged fish (e.g. Johansson et al. 2009; Korsøen et al. 2012; Stehfest et al. 2017; Wright et al. 2019). Combining this individual-based data to explain behaviour of a group of animals is not the best use of tags when less intrusive technologies are available to measure group-based behaviours, such as depth distribution or biomass measurements. For example, echo-sounder technology has been used effectively in several studies to study the swimming depth distribution of salmon in sea-cages (for review see Oppedal et al. (2011b)). Although the echosounders used so far cannot deliver fine-scale measurements of parameters such as swimming speed, acceleration or movement in the horizontal plane, many studies do not require this type of data to answer questions about vertical distributions or swimming densities in production units. Camera systems can be of great value in gathering data on behaviours that echo-sounders cannot. Furthermore, setting up a tagging experiment can be more logistically challenging than using alternative technologies. An example of an unused technology in aquaculture is the split-beam technology used in fisheries, which can track individuals within schools over short time periods and give representative measures of individual behaviours in the 3D plane (e.g. Handegard 2007) and even tail-beat frequency (Handegard et al. 2009).

When other less-invasive technologies cannot provide the fine-scale individual data needed to answer a research question, applying tagging technology using more innovative and rigorous approaches can yield deeper insights. For example, Nilsson et al. (2013) tagged different sized salmon with PIT tags and deployed PIT registering antennae at different depths. To discover previously unknown size-dependent vertical distribution of farmed salmon, with larger fish generally swimming deeper. These results have direct consequences

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for size sampling at farms, as sampling fish from one depth only, will likely mean one size class is over-represented (Folkedal et al. 2012). Deciding whether a study would really benefit from tag data is an important consideration for aquaculture behavioural research.

### 2.3 Tag types

Different tag types each come with their own advantages and limitations. Tag type selection will largely depend on research question, type of data required and the study species and environment. Telemetry tags were used more often than DSTs, indicating the benefits of obtaining data wirelessly via transmission. Real-time relay of behavioural or physiological data means that farmers could monitor behaviour and make immediate management decisions (Berckmans 2014). This type of tag use relates back to the concept of 'sentinel' animals in livestock farming, where individuals act as representatives for the wider population (Andrewartha et al. 2015). However, since this concept is yet to be applied in the aquaculture industry, the advantage of not necessarily having to recapture tagged fish to gather data is likely what resulted in the preference for this tag type. In comparison, data from DSTs cannot be remotely collected so the tag must be retrieved to access the data. Despite this, DSTs can still provide valuable information, for instance in the form of high-resolution data that may shed light on fish recovery after surgery. Heart-rate DSTs have shown that the heart rate of tagged Atlantic salmon did not return to baseline levels for 4 days on average (up to 6 days) after surgery and anaesthesia (Føre et al. 2020). This has implications for interpretation of data from transmitters or other electronic tags, as for salmon at least, the first weeks of data may be unrepresentative of untagged fish. Similarly, in telemetry studies on wild fish, researchers often discount the first few days of data to allow fish to return to baseline behaviours. Heart rate patterns in moribund fish, caused by the surgery procedures, deviated

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largely from normal fish and stresses the variable representativity of HR in following trials (Hvas et al. 2020b). Furthermore, DSTs can validate behavioural data gathered at the group level by other methods (e.g. echosounders), such as environmental preference seeking (Johansson et al. 2009).

While PIT tags technically fall under the telemetry tag category, they are more often used to simply identify individuals due to their small size and limited capacity to relay more complex data. However, their application to study behaviour is increasing. As PIT tags are quick to insert, do not require sutures and are generally small ( 24 mm max ), negative physical effects of PIT tags are assumed to be less common than for larger electronic tags, however, dependent on fish and PIT tag size PIT tags can still have adverse outcomes for fish welfare (Vollset et al. 2020). Although PIT tags cannot provide information such as heart rate or muscular activity, using these lightweight tags in more creative ways can yield insight to individual fish behaviour. For example, counting registrations of individuals at PIT antennae has shed light on feeding habits and coping styles of halibut (Kristiansen \& Fernö 2007), the learning ability of salmon to refill their swim bladder at underwater air domes (Macaulay et al. 2020a), and the spatial learning skills of mulloway (Argyrosomus japonicus). However, PIT tag use is limited by the short detection range (only a few cm ), one tag occupying the detection range can block other tags from being detected, and that when tags approach antennae diagonally, identification of the tag's unique code may fail to be registered (Brännäs \& Alanärä 1993; Baras \& Lagardère 1995) and their use in small fish can lead to increased mortality and tag loss (Vollset et al. 2020).

External non-electronic marker tags were commonly used in addition to the transmitter or DST to aid in the identification of tagged individuals (20\% of studies). Non-electronic external

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markers dimensions were not widely reported (Table S1) as coloured T-bar anchor tags are light and small and thus thought not to affect fish buoyancy greatly. T-bar tags that are inserted into the dorsal musculature require extra handling and on top of internal surgery. While we did not find that fish tagged with both external markers and internal surgery suffered higher rates of mortality (Table 2), it does not mean that external marker dimensions should not be reported on more thoroughly. Three studies used only physical marker tags to identify individuals and make behavioural observations directly. This method of tagging for behavioural studies relies on being able to see the tag clearly either directly or via video, and so is often conducted in lab conditions or with few other conspecifics, and therefore is inapplicable at commercial scale.

### 2.4 Tag attachment

Tag attachment method did not predict tagged fish mortality or tag retrieval. However, as most studies (83\%) used internal attachment via surgical implantation, the sample size of studies using externally attached electronic tags was too small to provide robust insights on comparative effects of internal and external tags. In wild fish studies, tag retention is species specific, dependent on behaviour, physiology, and environment (Baras \& Lagardère 1995; Bridger \& Booth 2003). Regardless of species, both external and internal attachment methods have some general advantages and disadvantages. Arguably, surgical implantation is the most invasive method of tagging. Fish must endure complete anaesthesia, ventral cuts, insertion of a foreign body near vital organs, suturing, and a higher risk of infection compared to other attachment methods (Jepsen et al. 2002). Historically, the positives of internally tagging are that tags are often closer to the fish's centre of gravity, interfering less with balance, and largely have longer retention times compared to external tags (Baras \& Lagardère 1995),

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although we did not find longer retention times for internal tags in this review. Advantages of external tags include quicker attachment and handling times and tagged fish being easily visible (Bridger \& Booth 2003). However, external tags can affect balance and cause drag (Thorpe et al. 1981; Jepsen et al. 2015), and in high velocity waters or with constantly high speed swimming species, extended periods of drag can erode dorsal muscles (Baras \& Lagardère 1995). Moreover, external tags can cause irregular swimming and scraping behaviour and are more prone to being entangled in structures (Collins et al. 1999). Larger external tags influence the amount of inflammation and damage to the epidermis at the site of attachment (Thorstad et al. 2000). Furthermore, externally tagged fish may be more prone to stress during common husbandry practices that involve crowding, which could ultimately lead to death (e.g. Sykes et al. (2012)). It is likely that the perceived disadvantages of external electronic tags resulted in a preference for internal tagging for monitoring fish behaviour at farms.

### 2.5 Tag retrieval and data loss

As expected, tag retrieval decreased with study duration in open sea cages compared to tanks where $100 \%$ retrieval is possible. We did not include tags that malfunctioned or provided corrupt data in the 'lost' tags category, meaning that the number of useable tags is often even smaller than the total number retrieved. Mortality also results in loss of useable data in many studies. A combination of tag loss, tag malfunction and mortality can lead to studies having a very low data capture success rate. For example, Johansson et al. (2009) implanted 58 salmon with DSTs to measure the effect of different stocking densities on how salmon respond to thermal stratification in sea-cages. A total of 30 out of 58 DST tagged salmon were retrieved (15 were dead or undiscovered, 13 fish with tag loss). However, only 12 retrieved tags were

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useable in the normal-density group, and a total of 11 tags were useable in the high-density group, across replicates. This means that only 23 of the original 58 tagged fish ( $\sim 40 \%$ ) were retrieved with useable data. Similarly, Solstorm et al. (2018) used only 50\% of tag data in final analyses. Bauer and Schlott (2004) reported that 3 out of 4 tags stopped working during one trial period intended to observe overwintering behaviour of carp in a pond. Given the amount of useable data that long-term tagging studies can provide, together with higher instances of mortality as study length increases, the present application of tags in commercial aquaculture as a monitoring tool should be limited to short-term use, with the caveat that study length must be greater than recovery time.

### 2.6 Recommendations for future research to improve tagging in farm environments

Our results reveal a significant knowledge gap regarding tagging protocols for aquaculture which consistently lead to low mortality rates. This needs to be remedied to secure the welfare of tagged fish and reliability of data produced from tagging in production settings. A first step toward this should be targeted research around the effects of (1) characteristics of the tag itself, (2) various aspects of the tag attachment procedure; and (3), how this interacts with a variety of biotic aspects of fish in production environments. In Table 3 we make a series of recommendations for more accurate and transparent reporting of tag use and outcomes for behavioural studies in aquaculture environments that should enable more robust assessments of data quality.

Studies that compare physical condition, growth, behaviour, and morphology between tagged and untagged conspecifics, as well as sham fish (that undergo a 'tagging procedure' without the insertion/attachment of the tag), are required. An additional handling treatment with no surgery (just handling and anaesthetic) would also be of value to determine the

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impact of handling and anaesthesia only. Contrasting the effect of biotic factors (such as lifestage, sex and source of fish) on the severity of negative tagging effects is also vital (Bridger \& Booth 2003). Additionally, investigating the effects of common events unique to the farm environment, such as various husbandry and farm operations, on physiology and behaviour should shed light if it is appropriate to carry out these routine procedures with tagged fish present. Aside from comparing the adverse effects of tagging between tagged, sham and untagged fish, testing different features of the tagging procedure will be important, including anaesthesia, attachment methods, incision sites, wound closures, sterility, antibiotic use, surgeon experience, recovery time and tag coatings (Bridger \& Booth 2003; Cooke et al. 2011). Furthermore, understanding the effects of essential abiotic factors such as dissolved oxygen levels, pH , salinity, and temperature of water on the outcome of surgery is important (Cooke et al. 2011).

A better understanding of the factors that contribute to sublethal or lethal tagging effects in an industrial setting will help to reduce adverse consequences of tagging. Based on our findings, tagged fish in sea-cage culture environments had worse outcomes. This is a challenging result for the use of these tags to monitor behaviour in industrial aquaculture settings. Mariculture continues to use even larger cages (e.g. Oldham et al. (2018)), with interest growing in placing these cages in more exposed locations, further away from populated coastlines to increase production and reduce adverse ecological impacts associated with nearshore farms, such as disease and parasites (Froehlich et al. 2017). Thus, further research is required if the approach of using tags to aid behavioural monitoring is to succeed in these future farming environments.

### 2.6.1 Experimental design

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When planning a tagging study in aquaculture, a clear research question is essential. What do you want to know about farmed fish behaviour? Will the data that a tag provides be suitable (and necessary) to answer that question? Wasted time, effort and resources can result from inadequately planned studies that do not have clear aims and that simply use tags for the sake of using tags (Koehn 2012; Brownscombe et al. 2019). While wasting time and money can be costly for research, it is also costly for the fish.. The ethical and welfare issues associated with putting fish through the stressful event of tagging that increases their likelihood of experiencing sublethal and lethal effects to provide data that might be unusable is unacceptable. Replacing the use of tagged fish when possible with alternative technologies (e.g. with cameras, echosounders) must be considered before tags are employed in aquaculture settings.

We found that $59 \%$ of studies had all tagged fish in a single cage, tank, or pond. Not all work in a single experimental unit is designed poorly, especially if the authors are not generalising their results to situations outside of that unit, such as for proof-of-concept work. However, it is inappropriate to make generalisations about fish behaviour in many cages or many farms using data from just one cage. Increased replication of experimental units (sea-cages, tanks, ponds) is then needed and we caution against using multiple individual tagged fish in one production unit as pseudo-replicates. A minimum of three replicates per treatment is advised, with more being desirable. For example, Føre et al. (2017) used two sea-cages within close proximity of each other, with comparable stocking densities, environment conditions and farm management in their study testing the concept of using acoustic tagged fish as 'sentinels' for commercial salmon aquaculture. They found that individual behaviour and vertical distribution of tagged fish differed between the two similar and co-located replicate cages.

On average, tagged fish represented $4.9 \%$ (median $0.6 \%$ ) of the entire population per unit replicate (among studies that reported numbers of untagged conspecifics). Increasing the number of tagged fish per unit replicate would increase sample size and representativeness, but this must be balanced against the animal welfare and research costs of tagging unnecessarily large numbers of fish. Furthermore, looking at the correlation between tag retrieval and study duration, more research is needed to ensure there is less tag loss over time, opposed to simply adding more tagged fish to compensate for tag loss.

### 2.6.2 Reporting on tagging effects

Transparency is necessary when reporting on mortality in tagging studies (Klinard \& Matley 2020). In Table 3, we make a series of recommendations for what study aspects to report on for research using tags to monitor farmed fish behaviour in aquaculture environments. Firstly, if there was no mortality of tagged fish, this should be stated clearly. Secondly, if there is mortality during the tagging procedure, the number of individuals that died should be stated to highlight the potential negative effects of the tag, tagging method or combination of both. Only considering fish that survived may lead to bias, as these survivors could exhibit tag related morbidity and thus might not be representative of the whole untagged population. Reporting on the mortality rate of untagged fish is also crucial. Without knowledge of standard mortality among untagged individuals, we cannot know how much tagging influences survival. This unique opportunity exists in aquaculture, as mortality is recorded frequently at fish farms (it is much more difficult to estimate mortality of untagged wild fish). Any sublethal physical effects found on tagged fish should also be stated plainly, because an injured or stressed fish may behave differently in the long term, and sublethal behavioural effects (e.g. scraping, swimming behaviour) should be reported for the same reason.

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Tagging technology is becoming increasingly more available, affordable, and more sophisticated. Simultaneously, the negative effects of these tags should decrease as tags evolve and become more miniaturised and attachment methods less invasive in the future. Currently, tagging is still a stressful experience for fish. Researchers using tags should be aware of the potential lethal and sublethal effects of tags on fish behaviour in aquaculture and how these deviations from normal behaviour can bias results and create poor welfare. Through a better understanding of the negative effects of tags, a best practice for tagging in production environments can be established, leading to the improved welfare and data validity of tagged fish.

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Table 1 - Names and description of predictor and response variables tested.

| Predictor <br> variables | Description | Type |
| :--- | :--- | :--- |
| StudyDays <br> UnitVol_m3 | Duration of trial (days). <br> Volume $\left(m^{3}\right)$ of the tank/cage or pond the trial was conducted in in <br> cubic metres. | Numeric |
| Attach | Method used to attach the tag to the fish (3 levels): Internal (tags <br> surgically implanted); External (tags were externally attached to <br> musculature or skin); Both (both internal and external tags used within <br> the study). | Factor |
| Response | Description | Proportion of tagged fish that died per unit replicate. Only included in |
| variables | models if numbers of tagged fish were given. | Numeric |
| TagMortProp | Proportion of tags retrieved per unit replicate. Only included in models | Numeric |
| TagsRetrvdProp | if reported. | proportion |

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Table 2 - Results and description of beta-regression models (betareg package for R; CribariNeto \& Zeileis (2009)) testing for an association between tagged fish mortality (Model 1) or proportion of tags retrieved (Model 2) and study characteristics in sea-cages. Associated pseudo- $R^{2}$ and sample size ( $n$ ) of each model are included. Bold text indicates a significant change in the deviance of the model when the associated predictor variable is added to the model. Model terms were using the Anova function in the car package for R, using type II sums of squares (Fox \& Weisberg 2019).

| Response 1: TagMortProp sea-cage Model 1 n=17 | betareg(TagMortProp ~ UnitVol_m3 + Attach + StudyDays, weights $=$ No.treatrep) |  |  |
| :---: | :---: | :---: | :---: |
| Pseudo model $\mathrm{R}^{\mathbf{2}}=0.54$ |  |  |  |
| Term | X2 | Model df | P |
| UnitVol_m3 | 0.86 | 1 | 0.3536 |
| Attach | 0.46 | 2 | 0.7945 |
| StudyDays | 14.20 | 1 | 0.0001 |
| Response 2: TagsRetrvdProp seacages | betareg(TagsRetrvdProp ~ UnitVol_m3 + Attachment + StudyDays, weights $=$ No.treatrep) |  |  |
| Model $2 n=19$ <br> Pseudo model R2 = 0.6 |  |  |  |
| Term | X2 | Model df | P |
| UnitVol_m3 | 0.53 | 1 | 0.4664 |
| Attach | 0.76 | 3 | 0.8587 |
| StudyDays | 28.79 | 1 | <0.0001 |

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Table 3 - Summary of recommendations for experimental design and reporting for tagging studies in aquaculture.

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## Experimental design recommendations

| Considerations and suggestions | Description |
| :--- | :--- |
| Replacement | Are there alternative technologies available to use to monitor <br> farmed fish behaviour that could answer the research question <br> without the use of tags? For example, video cameras or echo- <br> sounders? <br> If possible, have at least three separate unit replicates <br> (ponds/tanks/cages) and divide equal number of tagged <br> individuals among these to avoid pseudo-replicates |
| Appropriate replication at the level of treatment | Compare the growth, behaviour, physiology of tagged fish to (a) <br> untagged conspecifics, (b) 'sham' or procedural control fish and (c) <br> fish that were handled and anaesthetised only. <br> Compare the growth, behaviour and physiology of tagged fish of |
| Tagging effects | different life-stage, sex or source |
| Test how different abiotic factors such as temperature, pH, |  |
| Biotic factors | dissolved oxygen and salinity affect the outcome of surgery <br> Compare the growth, behaviour and physiology of tagged that |
| Abiotic factors | have undergone different tagging procedures. For example, <br> different recovery times, attachment methods/locations, <br> anaesthesia, wound closures, surgeon experience. |
| Aspects of tagging procedures |  |

## Recommendations for reporting

| Before tagging | Description |
| :--- | :--- |
| Tag dimensions | Record tag brand, length, width, volume, weight in air, weight in <br> water. Ensure tag dimensions are reported for external marker <br> tags also. <br> Tagged fish source |
| Where did the tagged fish come from? |  |
| Method of collecting fish | Were fish crowded? Was a net used? |

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| Number of fish tagged | Record the total number of fish that underwent the tag attachment procedure. |
| :---: | :---: |
| Mortality | Record the total number of fish that did not recover or died after tag attachment. |
| During trial | Description |
| Number of tagged and untagged fish per replicate | Record the total number of tagged and untagged fish placed in each unit replicate. |
| Behaviour and physical health | Record any behavioural abnormalities and physical signs of damage of tagged fish throughout the trial compared to untagged fish. |
| Mortality | Total number of both tagged and untagged fish that died throughout the trial in each unit replicate. |
| Farm procedures/experimental conditions | Note any farm or animal husbandry procedures that occurred during the trial (or any experimental manipulation). |
| Environmental conditions | Volume and depth of tank/cage/pond. Water temperature, dissolved oxygen levels and salinity over the trial period. |
| Post-trial | Description |
| Tag retrieval | Total number of tagged fish successfully retrieved per unit replicate and number of tagged fish that were undiscovered. |
| Weight and length | Mean fish weight in air and length after trial for both tagged and untagged fish. |
| Useable tags | Total number of tags retrieved that provided complete and useable data sets per unit replicate. |
| Weight of fish in water | Weighing fish in water after euthanasia (with swim bladder deflated or removed) allows fish density and fish volume to be calculated. |
| Tagged fish condition | Inspection of the tag attachment site: is there any inflammation? Signs of necrosis, infection? Do the tagged fish's physical condition appear visibly different to the untagged fish? |

## Publications



Figure 1 - Heat map of number of publications from different countries on monitoring behaviour in farmed fish using electronic tags. benefits, problems and solutions. Reviews in Aquaculture, 18. DOI: https://doi.org/10.1111/raq. 12534


Figure 2 - Number of studies on monitoring behaviour using tags on different species of farmed fish.

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Figure 3 - Number of publications returned by the literature review, according to tag type and study focus. Publications that used multiple tag types appear multiple times. Note: Transmitter (movement) = 3D positioning, depth transmitters. Transmitter (activity) = acceleration, swim speed/angle, body angle transmitters. Transmitter (physiology) = electromyogram, opercular pressure transmitter. Transmitter (environment) = pressure, dissolved oxygen transmitter. DST = any kind of archival data tags (e.g. heart rate bio-loggers). External marker tags = coloured tags, floy and T-bar tags, and numbered Peterson's discs.

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Figure 4 - Relationship between tagged fish mortality (\%) and study duration in sea-cages (A) and tanks (B) and the relationship between tags successfully retrieved (\%) and study duration in sea-cages (C) and tanks (D). Dot size is scaled for number of tagged fish per unit replicate and the graph is weighted by number of unit replicates. Fitted lines are predictions from a beta regression model fitted using the betareg package for R (Cribari-Neto \& Zeileis (2009)) (model specification: Response ~ StudyDays).

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