

Preserving permanent mountain grasslands in Western Europe: Why are promising approaches not implemented more widely?

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2 **approaches not implemented more widely?**

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4

5 **Abstract.** *To protect grasslands and maintain the ecosystem services they provide, many*
6 *European countries have been offering agri-environment measures aimed at maintaining*
7 *extensive grazing by cattle, sheep or goats. Yet, despite more than two decades of agri-*
8 *environment measures, semi-natural mountain pastures are still seen as threatened by*
9 *abandonment and subsequent shrub encroachment. Building on a three-round Delphi*
10 *inquiry, we analyse the perception of a range of experts on how measures aimed at*
11 *maintaining mountain grasslands are designed and implemented in Austria, France and*
12 *Norway. Results show that the experts see the need for a stronger involvement of diverse*
13 *regional actors, the need to increase the flexibility given to farmers in managing mountain*
14 *grasslands, and the need to reconceptualise monitoring as a social learning process. While*
15 *these approaches are implemented in some ‘best practice’ examples, they are not*
16 *widespread. Understanding these approaches as requiring double-loop learning may*
17 *contribute to explaining their limited spread. Indeed, they build on a radically different*
18 *conceptualization of farmers and of researchers, and thus of how agri-environment*
19 *measures need to be designed and implemented to be effective. Yet, such radical changes*
20 *are likely to be resisted.*

21 **Keywords:** open landscapes; ecosystem services; single-loop and double-loop learning; agri-
22 environmental schemes; mountain grassland

23 **1. Introduction**

24 Historically, permanent mountain grasslands have been used by farmers as pastures in the
25 summer months, to graze cattle, sheep or goats (Poschlod and WallisDeVries, 2002). These
26 grasslands are semi-natural, i.e. they require management by farmers to be maintained. They
27 are species-rich, but productivity is low as the growing season is short and they are usually
28 located on nutrient-poor soils (Hopkins, 2009). While this makes them less attractive to
29 farmers, they are valued by society for the broad range of ecosystem services they provide.

30 These services are highly interconnected and include regulating services, such as buffering
31 climate extremes, preventing flooding, and purifying water; provisioning services, such as
32 providing high quality fodder for livestock; supporting services such as nutrient cycling,
33 maintaining biodiversity and soil fertility; and cultural services, such as contributing to the
34 aesthetic value of open landscapes and offering a space for recreational activities (Gibon,
35 2005; Quétier et al., 2010; Lindemann-Matthies et al., 2010; Lavorel et al., 2011; Ocak,
36 2016).

37 The land-use changes induced by agricultural modernisation are threatening these extensive
38 grasslands (MacDonald et al., 2000; Eychenne, 2008). Indeed, while in favourable areas
39 agriculture has intensified, in less favourable areas – such as mountain areas – land tends to
40 be abandoned. As a result of abandonment, the semi-natural mountain grasslands are
41 encroached by shrubs and may over time revert to forests (Cocca et al., 2012; Carlson et al.,
42 2014). This change in land-use is linked to changes in the ecosystem services that can be
43 provided (Schirpke et al., 2013).

44 In an effort to counter-act the adverse impact of agricultural practices on the environment,
45 the 1992 MacSharry Reform of the Common Agricultural Policy (CAP) required every
46 Member State to introduce an agri-environment programme¹ (see Council Regulation (EEC)
47 2078/92; Potter and Goodwin, 1998; Strijker, 2005; Isoni, 2015). Since their inception
48 almost 25 years ago, the programmes have evolved over the subsequent 7-year programming
49 periods of the CAP. The programmes are diverse, not least given the high level of
50 subsidiarity which allows the Member States much leeway in the design of their overall
51 programme and of individual agri-environment measures (Beckmann et al., 2009). What
52 they have in common is the basic rationale: participation is voluntary, and the state pays
53 participating farmers to deliver an environmental service. Such payments have been offered
54 to farmers to maintain grazing the mountain pastures in the summer months, so as to keep
55 the landscape open and contribute to preserving the specific biodiversity of these semi-
56 natural grasslands.

57 The agri-environment measures have been relatively successful regarding their uptake,
58 however they have been only partially successful in achieving their conservation goals

¹ While Norway is not a Member State of the EU, it has also implemented agri-environment measures as part of its agricultural policy

59 (Uthes and Matzdorf, 2013; Dedeurwaerdere et al., 2015; Hinojosa et al., 2016). Various
60 reasons for this limited effectiveness have been identified, such as the influence of broader
61 societal changes leading to continued farm abandonment (MacDonald et al., 2000; Marini
62 et al., 2011); a lack of economic attractiveness of the measures which focus on compensating
63 cost incurred and income forgone, rather than being incentive payments (Hasund, 2013;
64 Saunders, 2015); or the design and implementation of the measures (Gross, 2011; Ingram et
65 al., 2013; Girard et al., 2015). Indeed, by prescribing specific management practices, the
66 measures insufficiently acknowledge the spatial diversity of mountain grasslands, the
67 complexity of ecological processes, and the uncertainties regarding the impact of climate
68 change (Komac et al., 2013; Duru et al., 2015; Girard et al., 2015).

69 While there have been a number of studies focusing on why farmers do (not) adopt agri-
70 environment measures (e.g. Morris and Potter, 1995; Schenk et al., 2007; Uthes and
71 Matzdorf, 2013), there is much less literature available on the perception of institutional
72 actors (e.g. Beckmann et al., 2009). However, the views of these institutional experts
73 working in government agencies, in farmer associations, and in environmental NGOs are
74 important, as Member States have been encouraged to design the measures in a decentralised
75 and participatory way. The design of the measures is thus the result of a complex and
76 protracted political process (Rutz et al., 2013). This process starts at EU-level and leads to a
77 broad framework published by the European Commission (see e.g. Regulation (EU)
78 1305/2013 and Regulation (EU) 1306/2013) and ends when the Commission approves the
79 agri-environment programme defined by each Member State. The individual agri-
80 environment measures are designed in a process at (sub-)national level, and are thus
81 influenced by the respective policy arena, with its specific government structures, political
82 ideologies, and administrative culture, as well as relative political power of various policy
83 actors at various scales (Beckmann et al., 2009). While the specific processes that lead to
84 defining a measure vary, in most cases the agricultural administration and farmers' interest
85 groups play a defining role, but the environmental administration as well as researchers and
86 environmental NGOs may also be involved (Beckmann et al., 2009; Benoit and Patsias,
87 2014). Overall, despite nationally varying efforts to include a diversity of actors during
88 development and evaluation, agri-environment programmes can still be seen as following a
89 state-led and expert-led mode of governance, characterised by a top-down approach to
90 designing and monitoring (Morris, 2006; Prager, 2015).

91 This paper aims to add to the discussion why agri-environment measures have so far been
92 limited in their effectiveness in preserving semi-natural mountain grasslands. We propose
93 that while measures targeting the maintenance of grasslands have certainly changed over the
94 last 25 years, the improvements were mostly incremental, i.e. based on single-loop learning.
95 While this might have improved the effectiveness of the administration of the measures in a
96 number of ways, it has not achieved the expressed goal: maintenance of semi-natural
97 mountain grasslands. The changes needed to achieve this goal might well require double-
98 loop learning, which would imply to design and implement measures based on radically
99 different assumptions.

100 The distinction between single- and double-loop learning was developed by Argyris and
101 Schön (1978) in the context of organisational learning. It has been transferred to learning in
102 a policy context (e.g. Grin and Lober, 2007; Pahl-Wostl, 2009; Hall, 2011). In the context
103 of agri-environment measures, we understand single-loop learning as referring to
104 incremental changes, such as adaptations of contractual arrangements or fine-tuning specific
105 aspects of prescribed management practices. This constitutes instrumental or technical
106 learning, based on the experiences gained during the implementation of measures in the
107 previous programming periods. It is concerned with adjusting the measures to address day-
108 to-day problems and with increasing the efficiency of various processes. The aim is thus to
109 improve performance, without questioning established routines, or the underlying
110 assumptions and beliefs. In contrast, double-loop learning does question the assumptions
111 that guide the definition of priorities, of the boundaries of the system under consideration,
112 and of means suitable to achieve the goal. As a result, they present a radical departure from
113 established practices. As Pahl-Wostl (2009) points out, this often implies the need for social
114 learning, as it may lead to changes in the actors involved, and to shifts in the allocation of
115 resources. Distinguishing between changes that build on single-loop vs. double-loop
116 learning thus helps to understand why some proposed changes are resisted by some actors.
117 Indeed, as changes building on double-loop learning tend to be a radical departure from
118 mainstream approaches, they tend not be compatible with the dominant policy regime. This
119 may curtail their spread.

120 The next section describes how we collected the data using a Delphi inquiry to ask experts
121 in Austria, France and Norway to share their views on the current state of mountain
122 grasslands and on the agri-environment measures to maintain open landscapes. We then

123 summarize the changes the experts saw as necessary to make these measures more effective.
124 We do so under three broad headings: involving a broader range of stakeholders, increasing
125 the flexibility at farm-level, and reframing monitoring as a social learning process. We then
126 illustrate how these changes have been implemented in ‘best practices’ examples provided
127 by the experts. We close by discussing the extent to which these ‘best practices’ build on
128 double-loop learning, and how this may contribute to explaining why they are not
129 implemented more widely. Indeed, we argue that the changes imply a radically different
130 conceptualisation of farmers and of researchers; and as a result of the designs that are
131 perceived as effective. However, radical changes in the design and implementation process
132 are likely to be resisted.

133 2. Method: the Delphi inquiry

134 The Delphi method of inquiry is a qualitative method through which information is gathered
135 iteratively, involving a panel of subject-matter experts (Hsu and Sandford, 2007; Grisham,
136 2009; Häder, 2009). While the Delphi technique has been used to seek consensus and make
137 predictions, in this study, it was used to reveal commonalities between the three countries,
138 and to enable experts to learn from each other’s experiences and proposals for promising
139 design options. In contrast to interviews, the Delphi inquiry allows reflection on the results
140 of the previous round and allows the experts to reflect on their answers in light of the answers
141 of other experts.

142 Members of the expert panel were recruited in Austria, France, and Norway. Experts were
143 identified informally, mostly through direct contacts of researchers, who were familiar with
144 and engaged in networks related to mountainous grasslands. Further experts were identified
145 through their membership in formal working groups and committees, as well as through
146 referral. The aim was to include all groups who were or who could be involved – directly or
147 indirectly – in the design or administrative implementation of measures. We thus recruited
148 experts from a range of occupational backgrounds: experts working in government agencies
149 (at regional, national, and EU level), in advisory services, in research and education, in
150 NGOs concerned with environmental protection and rural development, as well as in private
151 sector businesses. We did not include farmers because the aim was not to understand the
152 challenges of implementing specific measures in a specific place, but gain an overall view
153 of the administrative implementation process.

154 The target was to ensure that approximately 25 experts from each country (Tab. 1), with an
 155 even spread across the occupational backgrounds (Tab. 2) participate in the first round. This
 156 allowed including a sufficiently broad range of viewpoints in the study, while keeping the
 157 amount of material produced to a manageable size. Some experts at EU-level were also
 158 invited (Tab. 1), as were businesses (Tab. 2). However, the participation rates of these two
 159 groups was low, possibly because the management of mountain grassland is not part of their
 160 core tasks. Non-response from experts is a major challenge in Delphi inquiries and attrition
 161 rates can be high in each round (Padel and Midmore, 2005). In this study, participation fell
 162 from 87 experts involved in the first round to 39 experts in the third round (Tab. 1 and Tab.
 163 2).

164

165 Table 1 – Number of experts participating in each of the three Delphi rounds, from each country and from the
 166 EU-level

	Round 1	Round 2	Round 3
Austria	24	20	12
France	29	22	17
Norway	26	13	7
EU	8	5	3

167

168 Table 2 – Number of experts participating in each of the three Delphi rounds, from each occupational
 169 background

	Round 1	Round 2	Round 3
Government agencies	23	15	9
Advisory services	20	16	12
NGOs	25	16	9
Research and education	18	13	9
Private sector businesses	1	0	0

170

171 Invitations to participate in the Delphi inquiry were sent out by email, which included a link
 172 to the web-based platform used to collect the data. The invitations to the three rounds, were
 173 mailed out in March 2014, June 2014, and January 2015. Each round included 10-15 closed
 174 questions, with a number of pre-defined answer options. These were followed by open
 175 questions and space for comments. In the next section, when quoting from these comments,

176 we include the information regarding the occupational background and the nationality of the
177 expert. After each round, we compiled and analysed answers to refine issues and to develop
178 the next set of questions. At the beginning of the second and third round we reported the
179 summarized findings to the experts and we invited them to comment on them. The questions
180 on the web-platform were translated into German, French and Norwegian and pretested
181 before being put on-line, and the comments were translated back into English for analysis.

182 The first round was a ‘scoping’ round, to get an overview of the perceptions of the experts
183 on major trends. Among other, we asked them to identify the biggest problem in the context
184 of mountain grasslands. The majority of experts (88%) saw maintaining an open landscape
185 as the biggest challenge, mainly due to land abandonment and the resulting shrub
186 encroachment. Preserving mountain grassland was seen as linked to broader societal
187 changes, that undermine the economic and social viability of mountain farms, and induce
188 shifts in farming practices, i.e. the intensification of grasslands closer to the farm and the
189 abandonment of extensive mountain grasslands.

190 In this paper we build on the results from the second and third rounds, which focused on the
191 measures that aim to safeguard the openness of the landscape. The second round was used
192 to explore which institutions should be given more competences in designing and
193 implementing measures. The third round was used to assess which stakeholders are currently
194 included and which should be included, as well as to ask experts for ‘best case’ examples.

195 3. Changes suggested to more effectively maintain open landscapes

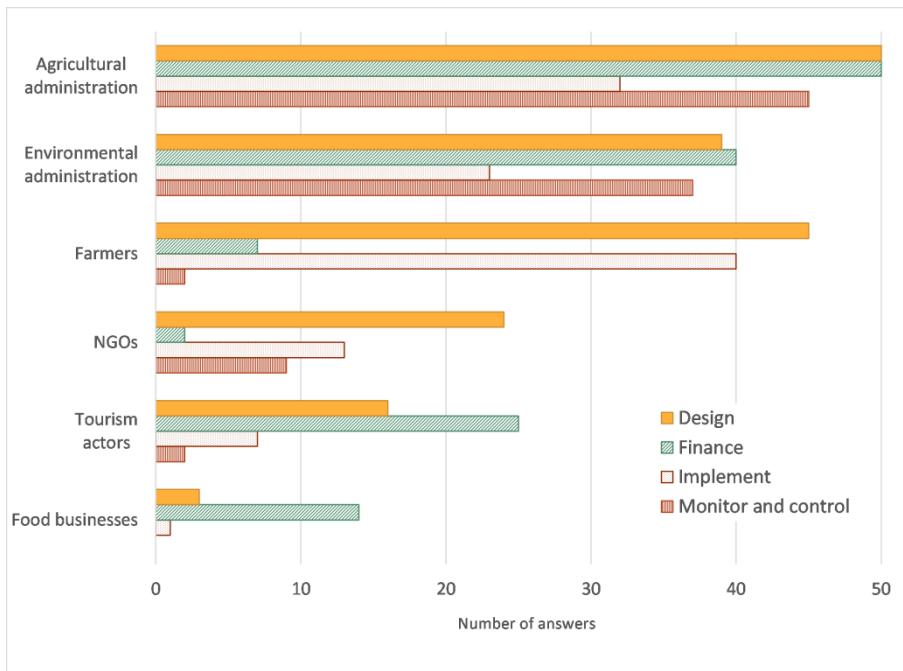
196 3.1. Broaden the range of involved stakeholders

197 The experts were asked which stakeholder groups should be involved in various aspects of
198 designing and implementing measures that aim at keeping the landscape open. For this
199 question we distinguished between four tasks, which broadly speaking can be seen as the
200 four stages of an agri-environment measure. With ‘designing’ we labelled the process of
201 defining the measure, i.e. the conditions tied to the payments, before the agri-environment
202 programme is submitted to and approved by the Commission. While discussions on who will
203 cover the cost of the payments to farmers and of administering and monitoring the measures
204 is usually part of the design process, we have included ‘financing’ as a separate category to

205 be able to assess the perceived potential for various stakeholders in covering the costs of the
206 measures. Regarding ‘implementation’ we distinguish between administrative bodies, where
207 implementation is the process leading up to the contract with a farmer; and farmers where
208 implementation refers to complying with the contract on-farm, i.e. the use appropriate
209 management practices. ‘Monitoring and controlling’ are the processes that ensure that the
210 management practices do comply with the terms of the contract and contribute to achieving
211 the aim of maintaining the semi-natural mountain grasslands.

212 We also distinguished between six groups of stakeholders. The ‘agricultural administration’
213 included all offices linked to the ministry of agriculture, including advisory services, the
214 chamber of agriculture and the administration linked to implementing various aspects of the
215 CAP. The ‘environmental administration’ are the offices in charge of environmental
216 protection and of administering protected areas. ‘Farmers’ include individual farmers as well
217 as farmer associations. ‘NGOs’ are all civil society actors, e.g. environmental groups and
218 consultants. ‘Food businesses’ refer to all stakeholders along the food chain, i.e. processors
219 and retailers; whereas ‘tourism’ are associations and local/regional public bodies involved
220 in promoting tourism and cultural heritage in the region.

221 The responses show that experts thought that both the agricultural administration and the
222 environmental administration should be strongly involved (Fig. 1). However, the required
223 close coordination between agricultural and environmental administrations remained a
224 challenge. On the one hand, this may be due to the fact that in some countries they are linked
225 to different administrative structures, at different levels of government, i.e. national and
226 regional. On the other hand, experts reported that there still is a “rift between nature
227 protection and farming” (Austria, advisor). An Austrian working at an NGO attributed this
228 to persistent animosities and poor communication, including negative media reports by
229 agricultural groups against Natura 2000. Similarly, environmental awareness had led to a
230 fundamental suspicion towards farmers, which have been portrayed as “polluters, cheaters,
231 profiteers” (France, NGO).



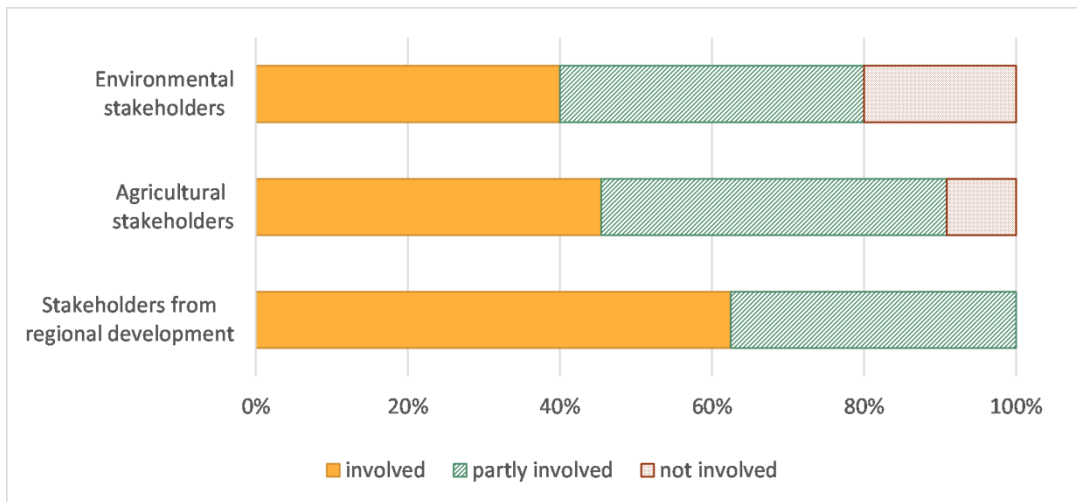
232

233 Figure 1: Number of experts indicating that a stakeholder group should be involved in a task linked to measures
 234 targeting the openness of landscapes (n=59)

235 The responses also show that a whole range of stakeholders should be involved in most tasks.
 236 By giving important roles to various actors such as NGOs, tourism and food businesses, the
 237 experts clearly conveyed that they saw the value of a territorial approach to environmental
 238 protection, rather than a sectoral approach with a one-sided focus on agriculture. The
 239 literature indicates various ways in which actors could be involved in a context-specific
 240 manner, e.g., through rules within geographic indications or other food labels (see e.g.
 241 Lamarque and Lambin, 2014), or financial support by tourism actors for specific measures
 242 (e.g. maintenance of traditional wooden fences, see e.g. Blumentrath et al., 2014).

243 We then asked experts whether those stakeholder groups they thought *should* be involved in
 244 designing measures, were actually participating in the process. Only around half of the
 245 experts indicated that this was the case (Fig. 2). As a researcher from Norway commented:
 246 “researchers generally score low on influence, as do environmental organizations and
 247 cultural heritage organizations. To ensure sustainable development in rural mountain areas
 248 all three should have a stronger influence”. Other experts also pointed out that farmers’
 249 associations or associations for the maintenance of the cultural landscape often have valuable
 250 suggestions, but are not sufficiently involved in the process (see also Beckmann et al., 2009).

251



252

253 Figure 2: Percentage of experts who state whether – and to what extent – a stakeholder group should be
 254 involved in the process of designing measures (n=29)

255

256 A respondent from a French NGO highlighted that while a number of stakeholders are
 257 consulted, this ‘participation’ is mostly a form of tokenism (see Arnstein, 1969). Indeed,
 258 local needs and a territorial approach tended to carry little weight, as “regulations are often
 259 the outcome of highly centralized negotiations at national level, with a strong and powerful
 260 sectoral agricultural representation, which negotiates on issues that do not necessarily
 261 represent the specific interests of mountain farming” (researcher, France).

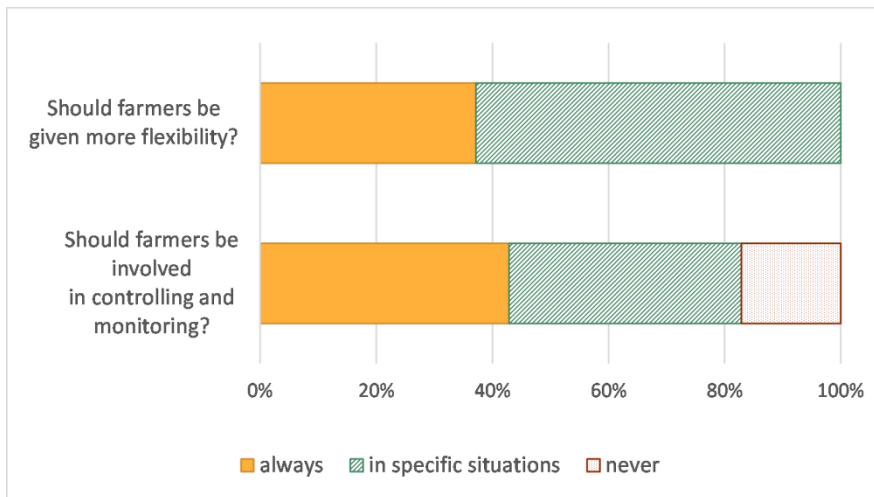
262 The experts clearly conveyed the need for a more inclusive, participatory process to improve
 263 the design and implementation of the measures. An advisor from Norway suggested that
 264 “the process should start with openness, like a public meeting or a survey, which then should
 265 be analysed and elaborated in a smaller working group where relevant actors participate. It
 266 is important to have an open process and make room for feedback”. This approach was
 267 mirrored by an expert from a French NGO: “Ideally, the process would be initiated through
 268 a local demand, be it the farmers, the villagers or the municipality. Then researchers and
 269 locally active associations should jointly establish a diagnostic. Based on this, various
 270 scenarios and possible projects could be discussed.” The experts thus advocated an
 271 integrated territorial approach guided by a committee, which would ensure that the needs of
 272 the different actors at regional level are heard. This approach implies not only a shift to more
 273 participation, but also a shift towards more flexibility in the agri-environment measures, so
 274 that their implementation can be tailored to the specific local context.

275 The overall understanding of the experts was that the roles of stakeholders at different levels
276 should be more differentiated: the EU and national level should be primarily responsible for
277 setting the financial framework and for setting (environmental) goals. Regional stakeholders
278 should be in charge of operationalizing these goals by identifying the most economically
279 efficient and environmentally effective way to use the available funds, given the specific
280 regional and local context. The experts saw the collaboration between farmer associations,
281 environmental NGOs and researchers (ecologists, agronomists, but also social science
282 researchers) at local or regional level as very promising. It would give a stronger role to a
283 diversity of local stakeholders (rather than prioritizing the agricultural administration or
284 farmer lobby groups), and this “would enable more dynamism, responsiveness and
285 coherence regarding the local context” (France, researcher). The aim of policies should thus
286 be to “create a framework that encourages initiatives, AND trust local actors to implement
287 the initiatives” (France, NGO, emphasis in original).

288 3.2. Enable flexibility at farm level

289 The experts in the Delphi inquiry also pointed out the need to increase flexibility at farm
290 level. Indeed, when experts were asked whether farmers should be given more flexibility,
291 none of the experts said ‘never’, while 37% said ‘always’ (Fig. 3). Experts who selected ‘in
292 specific situations’ were asked to elaborate why and how. They distinguished between goals
293 and practices to achieve them: while farmers should not be flexible on the goals to achieve,
294 they should be granted more flexibility in *how* to achieve them. This would enable farmers
295 to select the practices most appropriate in a specific year, to the specificities of a grassland
296 plot, and to their farming system.

297



298

299 Figure 3: Percentage of experts indicating whether farmers should be given more flexibility in implementing
 300 measures to maintain an open landscape, and whether farmers should be involved in controlling and monitoring
 301 (n=35)

302

303 Indeed, some agri-environment measures seem to micro-manage farming practices (e.g.
 304 through specific cut-off dates that need to be implemented in all areas and all years). As a
 305 French researcher pointed out, many farmers “live a large mismatch between the practical
 306 realities of their job, the way they would like to farm, and the requirements stipulated in the
 307 contract, which impose new practices.”

308 We interpret this demand for flexibility as a recognition of the site-specificity of effective
 309 management practices and thus the importance of local knowledge. Blanc (2009) details
 310 how, through observing nature and their animals, herders can acquire a detailed ecological
 311 knowledge, noting how the palatability of grass species changes over time, how
 312 microclimate, topography and soils influences species composition, all of which influence
 313 quantity and quality of grass at different times and different places of the grassland. Herders
 314 can thus have a dynamic understanding of the heterogeneity of the grassland, with season
 315 and weather patterns influencing the choice of how to optimally use the available feed
 316 resources in relation to the needs of their herd (Blanc, 2009). This dynamic and relational
 317 understanding of the appropriate use of the pastures cannot be reduced to standardized
 318 technical management rules (see also Girard et al., 2015; Peltola and Tuomisaari, 2015).

319 Over 60% of experts stated that farmers should be given more flexibility ‘in specific
 320 situations’ (Fig. 3). This conveys the importance of specifying how and under what

321 conditions this flexibility is given. Indeed, the experts were well aware that flexibility may
322 be linked with a range of risks to effectively achieving the environmental goals. Some
323 experts saw farmers as strongly constrained by economic pressures. For example, a measure
324 may stipulate that pastures can only be grazed starting on the 15th of August to protect certain
325 ground-breeding bird species. Yet there may be years when weather conditions and grass
326 regrowth would make grazing possible starting early August. Flexibility may thus result in
327 a conflict between on-farm economic interests (avoiding the purchase of feed) and broader
328 societal interests (securing bird populations). As an advisor from Norway noted: “in some
329 cases the farmers can have a too narrow view, especially if there are national and/or regional
330 concerns that should get priority over the local”.

331 Overall, the experts clearly warned against designing measures based on a naïve assumption
332 that farmers are “benevolent landscape managers” (France, government). They pointed out
333 that farmers do not see themselves as “landscape gardeners” or “flower caretakers” (Austria,
334 researcher). While mountain farmers will engage in environmental protection, they are
335 unlikely do so at the expense of agricultural production. It is important to acknowledge these
336 tensions, as a one-sided focus on grassland biodiversity would not be compatible with
337 farmers’ primary self-identity as food producers.

338 This was all the more important as some measures were seen by farmers as narrowly aiming
339 to preserve the past, which led to further tensions. Indeed, experts stated that farmers do not
340 want to live in a “reservation” (France, government), nor do they want to practice “museum
341 agriculture” (Austria, advisor). While they cherish many traditions, farmers are also well
342 aware that grassland management practices need to change to adapt to an evolving economic,
343 technological and social context, as well as to the impacts of climate change. It is thus
344 necessary for measures to enable the evolution of practices, identifying new ways to achieve
345 ecological goals.

346 The experts pointed out that an important impact of increasing the flexibility in the
347 implementation of measures would be to enable farmers to overcome a sense of frustration
348 and demoralisation, even “de-responsabilisation” (France, government) and the “feeling that
349 they can no longer manage their farm autonomously” (Austria, NGO). Indeed, as the
350 literature shows, farmers highly value their autonomy (Stock and Forney, 2014), which has
351 been undermined by restrictions imposed by expert-led, centrally defined measures that are
352 implemented in a top-down manner. The aim of increasing flexibility is thus to reground

353 farming practices in regional specificities, and to revalue the competencies and knowledge
354 of farmers. When designing and administering measures that increase flexibility, the
355 inherent tensions need to be carefully taken into account, i.e. between increasing farmers'
356 autonomy and avoiding a temporary 'opting-out'; between ecological and food production
357 goals; between maintaining traditional practices and enabling change.

358 3.3. Reconceptualise monitoring as part of a social learning process

359 The experts also raised the need to ensure effective implementation so as to achieve the
360 ecological goal of the measures, i.e. the conservation of semi-natural grasslands, and
361 ultimately of their specific biodiversity. Thus, how monitoring and controlling is organised
362 plays an important role (Cundill and Fabricius, 2009). We asked the experts whether farmers
363 should be involved in monitoring and controlling: 43% of experts indicated that farmers
364 should 'always' be involved, and 40% indicated that this should only be the case in specific
365 situations (Fig. 3). Those who answered 'in specific situations', were asked to elaborate.
366 Through their answers the experts highlighted a flaw in the question design, i.e. the need to
367 differentiate between monitoring and control. Indeed, individual farmers could not be
368 expected to control their own implementation of practices: they would be "judge and
369 claimant" at the same time (France, advisor).

370 Interestingly, in the second round of the Delphi inquiry, only two experts indicated that
371 farmers should be involved in monitoring and controlling (Fig. 2), whereas in the third
372 round, 15 experts indicated that they should 'always' be involved (Fig. 3). It is likely that
373 this discrepancy is – in large part – due to the fact that in the second round the question was
374 asked in relation to the current design of agri-environment measures. In the third round, the
375 experts were asked the same question, but specifically referring to their recommendation to
376 design measures so as to allow more flexibility and adaption to local conditions, while also
377 avoiding overburdening the administration. This can be interpreted as an indication that the
378 experts saw the role of farmers as dependent on the broader context: if the measures are
379 designed adequately and farmers are given appropriate training and support, farmers could
380 and should take a more active role, not least in monitoring.

381 As the literature shows, involving farmers in monitoring can promote social learning, e.g.
382 through joint visits of plots, discussion of possible measures, discussions of the outcome of
383 the implemented measures and search for improvements (Prager, 2015). As experts pointed

384 out, this approach is also successfully implemented in organic group certification as well as
385 in some cheese cooperatives with Controlled Designation of Origin (AOC cheeses). Such a
386 collective approach would require that, as part of the requirements to participate in the agri-
387 environment measure, a farmer “accepts that the group takes a critical look at his practices”
388 (France, NGO). As the expert explained, the aim would be to develop a dynamic where
389 farmers as a group take responsibility for outcomes, based on mutual feedback and shared
390 learning (i.e. the aim is not to create a league against external control, as some form of
391 external control will still be necessary). The experts saw this approach as most promising if
392 the group participates in defining the outcomes, rather than outcomes being imposed
393 externally.

394 As the experts pointed out, actively involving farmers in monitoring will require resources,
395 i.e. time for the participatory processes, as well as funds – either through additional funds,
396 or through shifting the allocation of current funds. Indeed, there was a need to offer more
397 support to farmers, either individually or as a group: “When giving farmers flexibility you
398 also need to support them with training and information to ensure the effectiveness of the
399 flexibility provided” (EU, NGO). Depending on the composition of the group, it may
400 promote a dialogue between farmers and researchers, thus revaluing local knowledge linked
401 to traditional practices as well as integrating scientific ecological knowledge (see Gross,
402 2011). It may also promote a dialogue between farmers and a range of local stakeholders,
403 such as those involved in tourism, wildlife conservation, or rewilding, allowing them to
404 clarify diverging interests and identify ways to address them.

405 The experts in the Delphi inquiry pointed out that an appropriate legal framework is required
406 to encourage collective action, experimentation and social innovation. As an example, an
407 expert (France, NGO) referred to the pastoral law (‘loi pastorale’) in France, which enabled
408 forming ‘pastoral groups’, which are now important actors in the management of mountain
409 grasslands (see Charbonnier, 2012; Eychenne and Lazaro, 2014).

410 4. ‘Best practice’ examples of alternative approaches

411 In the third round of the Delphi inquiry, the experts were asked to name ‘best practice’
412 examples, where (elements of) their suggestions were successfully implemented. These
413 examples illustrate how the identified shortcomings can be addressed, i.e. how various

414 stakeholders can be involved, how flexibility at farm level can be increased, and how a
415 framework encouraging learning processes can be built. We assigned the examples provided
416 to three broad groups and briefly characterize them based on published documents.

417 4.1. Contracts based on management plans

418 This approach is characterized by individualized management plans, in contrast to the ‘one-
419 size-fits-all’ approach of many agri-environment measures. Contracts based on management
420 plans address the complaint by farmers that the standardized practices prescribed in a
421 measure do not ‘make sense’ in their particular region or for their particular grassland. As a
422 result, the requirements are perceived as a meaningless burden, because they are either
423 ineffective or a much stricter management constraint than would be necessary to achieve a
424 comparable outcome.

425 A contract-based approach was implemented in France in a territorial agri-environment
426 measure (Mesure Agro-Environnementale Territorialisée – MAET H09; see Agreil et al.,
427 2011), in Norway as part of the Special Environmental Measures in Agriculture (Særskilte
428 Miljøtiltak I Landbruket – SMIL; see Blumentrath et al., 2014), and in Austria through the
429 tool ‘nature conservation plan for alpine pastures’ (see Aigner et al., 2007). In each case, the
430 contract is based on an individualized agreement between a farmer and an agricultural or
431 environmental administration. The specific practices to be implemented are based on an
432 assessment by an ecologist and negotiated between this ecologist and the farmer. As a result,
433 the defined practices are perceived by the ecologist as efficient to achieve the protection
434 goal, and by the farmer as feasible given his/her farming system and constraints.

435 Another example mentioned by a Delphi expert from France is the territorial pastoral plan
436 (Plan Pastoral Territorial – PPT), specifically the one that has been implemented in
437 Belledonne (see PPT Belledonne, 2010). In line with the priorities defined in the broader
438 territorial plan, individual farmers develop a plan to manage their pasture. This starts with a
439 joint on-site visit of the mountain grassland to understand the ecological setting, followed
440 by discussions how herding practices could be adapted to achieve the goals of the territorial
441 plan. This is then formalized in a five-year plan, which is the foundation of the contract with
442 a government body (Agreil et al., 2011).

443 In Norway, measures are available for targeted efforts to maintain landscape elements such
444 as meadows, wetlands, traditional buildings and paths. These elements are documented and
445 used to draw up individual contracts with farmers. This involves a participative process
446 including farmers, local public management bodies, and researchers (Blumentrath et al.,
447 2014; Daugstad et al., 2014).

448 One drawback of this individualized approach, are higher transaction costs (Matzdorf and
449 Lorenz, 2010; Franks, 2011). As a result, in France the approach is currently implemented
450 only in specifically designated areas, e.g. Natura 2000 areas. Another drawback is that the
451 negotiation process may be challenging for individual farmers. This concern is partly
452 addressed in the implementation of the MAET H09, where the farmer pays a pastoral
453 association to draw up the contract. Another way to overcome the barriers of transaction
454 costs for the government bodies as well as the administrative burden on individual farmers,
455 was pioneered in the Netherlands. There, contracts were not negotiated with individual
456 farmers, but collective management plans could be negotiated by a ‘Local Environmental
457 Cooperative’ (see Franks, 2011; Franks and Emery, 2013; Westerink et al., 2014; van Dijk
458 et al., 2015).

459 4.2. Contracts based on outcomes

460 The second group of approaches provides flexibility not so much through negotiating tailor-
461 made contracts that stipulate which practices need to be implemented, but through agreeing
462 on an outcome that must be achieved, giving farmers the flexibility how to achieve it. The
463 approach also builds on a cooperative understanding between farmers, the implementing
464 agency and environmental protection administration (Stolze et al., 2015).

465 The example of outcome-based contract mentioned by the Delphi experts are the ‘prairies
466 fleuries’ (flowering meadows) in France, and a similar pilot scheme which is being
467 implemented in Austria (Ergebnisorientierter Naturschutzplan, i.e. ‘results-based nature
468 conservation plan’). The ‘flowering meadows’ was implemented in France in the framework
469 of an agri-environment measure (MAE H07), where farmers were provided a predefined list
470 of 20 species, and each contracted plot needed to have at least four of these 20 species. The
471 approach was attractive to farmers as it offered flexibility in management (Plantureux et al.,
472 2011; Nettiier et al., 2012; de Sainte Marie, 2014). Moreover, a competition conveyed social
473 recognition for the quality of their grassland and their contribution to the maintenance of

474 biodiversity (Magda et al., 2015). Importantly, the competition was a joint initiative of the
475 Federation of Natural Parks, the Chamber of Agriculture, associations for Controlled
476 Designation of Origin (AOC) cheeses, bee-keeping, and environmental protection, thus
477 extending the range of actors involved.

478 While the ‘prairies fleuries’ were mostly applied to more intensively used grasslands, it
479 would seem that the principle is transferrable to semi-natural mountain pastures. For a
480 discussion of the strengths and limits of result-oriented measures, see e.g. Matzdorf and
481 Lorenz (2010), Nettier et al. (2012), Burton and Schwarz (2013), Stolze et al. (2015), and
482 Russi et al. (2016).

483 4.3. Incentives based on collaborative learning

484 As an example of an approach that explicitly builds on co-learning, French experts referred
485 to the action research project ‘Alpages Sentinelles’ (sentinel alpine pastures). Although no
486 formal contract or payments were offered, the project illustrates how an open-ended learning
487 process can be implemented. Three drought years (2004, 2005 and 2006) had raised the
488 awareness for the impact of climate change on the grassland in a national park. To strengthen
489 the adaptive capabilities of all actors, the park administration initiated a collaborative
490 learning process, involving farmers, herders, researchers, and park officers (Dobremez et al.,
491 2014). At the end of each grazing season, a team made of the herder, an officer of the national
492 park, and an expert on pastoralism visited the grassland. Through sharing their observations
493 about the status of the grassland in various plots, the team discussed possible causes for the
494 observed outcome, and options to modify grassland use to improve its ecological status.

495 Moreover, once a year, results and observations were discussed in a broader transdisciplinary
496 team, which also includes extension agents and researchers (agronomists, ecologists,
497 meteorologists). During this day-long workshop, data from ecological measurements and
498 weather data were linked with observations from the field. This enabled integrating insights
499 by researchers with observations by herders on growth patterns, and by farmers on labour
500 and economic constraints. This allowed raising the understanding of all actors while
501 discussing management options, e.g. regarding routes on the pasture, herd size, herd
502 composition and complementary feed sources. As a result, actionable knowledge was co-
503 produced using an iterative process of inquiry; a process, which allowed taking into account
504 on-going shifts in both core questions and context (Dobremez et al., 2014).

505 5. Double-loop learning to reconceptualise roles and relationships

506 The preservation of semi-natural grasslands has long been recognized as a priority given the
507 ecosystem services they provide. However, preserving open landscapes remains an ongoing
508 challenge in mountain areas. The experts in the Delphi-inquiry identified a number of
509 changes that could increase the likelihood that agri-environment measures achieve their goal
510 of maintaining an open landscape. These changes are characterised by inclusive participatory
511 processes in the design of measures, increased flexibility for farmers in their implementation,
512 and reframing monitoring as social learning. The call for such changes is not new (see e.g.:
513 Berkes and Folke, 2002; Gunderson and Light, 2006; Stenseke, 2006; Pahl-Wostl, 2009;
514 Williams and Brown, 2016). Moreover, the ‘best practice’ examples highlighted by the
515 experts show that these principles can be – and have been – implemented successfully. The
516 question thus is: why are they not more wide-spread?

517 We argue that these changes are characteristic of double-loop learning as they build on
518 radically different assumptions and beliefs compared to those underlying current governance
519 arrangements. They imply a shift in boundaries, not only in who is involved and how, but
520 also in what is considered relevant when assessing the effectiveness of a measure. They build
521 on reconceptualising the motivations and roles of farmers and researchers, and thus how
522 administrative processes need to be designed and implemented to enable effective agri-
523 environment measures.

524 Firstly, roles are reconceptualised. The role of farmers is shifted from passive recipient of
525 measures designed and specified by others, to active participants in the process of defining
526 which practices will be both feasible and effective in their specific context. Indeed, both best
527 practice examples – individually negotiated contracts and outcome-based schemes –
528 explicitly value farmers’ knowledge. Moreover, farmers are no longer reduced to
529 economically rational actors that engage in a market exchange for a conservation good (see
530 Falconer and Whitby, 2000). Rather, participation in an agri-environment measure is
531 understood to be motivated by monetary compensation, by the compliance with social
532 norms, and by the expression of personal values (Schenk et al., 2007; Burton and
533 Paragahawewa, 2011; Nettiier et al., 2012; Ingram et al., 2013; Russi et al., 2016). To enable
534 intrinsically motivated behaviour, a supportive governance and an appropriate
535 administrative design is needed (DeCaro and Stokes, 2008). As Dedeurwaerdere et al. (2016)
536 point out, this design should enable a feeling of competence (i.e. feeling efficacious in

537 relation to the task at hand, not least through discussion and feedback), a sense of autonomy
538 (i.e. being free to make important choices and direct one's action without unwanted
539 pressure), and promote relatedness (i.e. a sense of belonging to a group that values collective
540 goals).

541 The role of researchers is also reconceptualised. Rather than them being primarily involved
542 ex-ante in designing measures, and ex-post in evaluating their economic efficiency and
543 ecological effectiveness, they engage in an adaptive and iterative mode of inquiry. As the
544 example of the 'Alpages Sentinelles' shows, a transdisciplinary approach allows integrating
545 scientific knowledge, e.g. the outcome of mechanistic models of abiotic resource flows in
546 the plant-soil-atmosphere system, with the experiential knowledge of herders, and with the
547 perspective of various stakeholders. The approach is thus designed as a social learning
548 process, and explicitly acknowledges complexity, i.e. that situations are dynamic and options
549 are context specific (Pahl-Wostl, 2015). Redefining of the role of researchers thus also builds
550 on a radical change in the public understanding of science, shifting from a view of scientific
551 knowledge as objective, reliable, and authoritative; towards acknowledging that it is
552 preliminary, tentative, uncertain, and fragmented (Aufvenne et al., 2014).

553 Secondly, reconceptualising the roles of farmers and researchers implies a radical shift in
554 administrative processes. Indeed, the experts involved in the Delphi inquiry highlighted that
555 current design, implementation and monitoring processes were problematic. In many cases
556 these processes followed a technical-rational approach, which Morris and Reed (2007) have
557 characterized as the 'McDonaldization' of agri-environment programmes, i.e. a rationality
558 centering on efficiency, calculability, predictability, and control. These norms tend to value
559 activity, control, comfort, and clarity; rather than reflection, learning, and inclusivity (see
560 Allan and Curtis, 2005; Allen and Gunderson, 2011). However, the latter norms are needed
561 to engage in context-sensitive, open-ended learning processes. These learning processes
562 imply collaborative, territorial approaches, as well as iterative social learning amongst state,
563 market, and civil society actors. And indeed, the experts in the Delphi inquiry have drawn
564 attention to the need for most actors to be involved in most tasks linked to agri-environment
565 measures (see Fig. 1).

566 To enable such social learning processes, the evaluation of measures needs to be embedded
567 in a comprehensive iterative learning cycle to improve implementation and design (see
568 Cundill and Fabricius, 2009). Regarding economic efficiency, it requires a radical shift from

569 a one-sided focus on selected transaction costs, e.g. the cost of contract negotiation between
570 the government body and the farmer, as well as the cost of on monitoring to ensure
571 compliance (see Falconer and Whitby, 2000; Beckmann et al., 2009). As Gorrdard et al.
572 (2016) have shown, public planning processes can restrict the type of knowledge and values
573 taken into account to those considered valid for the decision process (e.g. utilitarian values
574 that can be monetized) and marginalize those related to nature, culture, and sustainability
575 (e.g. disregarding non-market costs and amenity value, discounting long-term effects,
576 neglecting cross-scale effects). Evaluation also needs to go beyond monitoring focused on
577 biophysical indicators and include social and procedural aspects (see Waylen and
578 Blackstock, 2017).

579 6. Conclusion

580 The maintenance of semi-natural alpine grasslands has remained a challenge in Austria,
581 France and Norway, despite the implementation of agri-environment measures to maintain
582 summer grazing. The changes that have been implemented in the measures over the last two
583 decades may well have been limited to single-loop learning, i.e. on incremental adaptations
584 to streamline various administrative aspects and to improve prescribed management
585 practices. Indeed, the experts who participated in the Delphi inquiry pointed out that
586 systemic issues linked to the design and implementation process have not been (sufficiently)
587 addressed. We argue that this would require double-loop learning, i.e. the questioning of
588 underlying assumptions. Indeed, the ‘best practice’ examples identified by the experts build
589 on a radical departure from the dominant assumptions regarding the role of both farmers and
590 researchers, and thus of how the design and implementation processes should be
591 administrated.

592 This could contribute to explaining why such agri-environment measures have not spread
593 more widely. Indeed, implementing the lessons of this double-loop learning requires a
594 transformation of the broader governance arrangements and the structural context, i.e. the
595 norms and values used to assess what is effective and efficient. It would require a
596 fundamental transformation of institutional arrangements towards one that enables and
597 encourages more critical, inclusive, and reflexive practice (Buizer et al., 2011; McLoughlin
598 and Thoms, 2015; Wyborn, 2015). However, as Pahl-Wostl (2009) pointed out, such a

599 transformation requires a change in boundaries and power structures, which makes it likely
600 that even individual changes are resisted (see also Eggers, 2006).

601 We posit that unless the wider structural context changes, an effective reconceptualising of
602 roles and relationships is unlikely, which will impede the spread of innovative governance
603 arrangements and management practices. However, innovative approaches that build on
604 exploratory and iterative social learning processes are important elements in the effort to
605 maintain semi-natural mountain pastures in the face of climate change, rural demographic
606 changes, and farm abandonment.

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