

The Environmental Impact of Green Consumption and Sufficiency Lifestyles Scenarios in Europe: *Connecting local sustainability visions to global consequences*

Gibran Vita^{a*}, Johan R. Lundström^b, Edgar G. Hertwich^c, Jaco Quist^d, Diana Ivanova^a, Konstantin Stadler^a and Richard Wood^a

^aIndustrial Ecology Programme and Department of Energy and Process Engineering, Norwegian University of Science and Technology, Trondheim, Norway.

^bEnvironmental and Energy Systems Studies, Lund University, Sweden

^cCenter for Industrial Ecology, School of Forestry and Environmental Studies, Yale University, New Haven, CT, USA.

^dFaculty of Technology, Policy, Management, Delft University of Technology, Netherlands

*corresponding author: gibranvita@gmail.com

Keywords: sustainable lifestyles, backcasting, participatory modelling, Environmentally-Extended Multiregional Input-Output (EE-MRIO), environmental footprints, sufficiency, green consumption, quality of life.

The sustainability transformation calls for policies that consider the global consequences of local lifestyles. We used stakeholders' visions of sustainable lifestyles across Europe to build 19 scenarios of sufficiency (net reductions) and 17 of green consumption (shift in consumption patterns). We applied Environmentally Extended Multi-Regional Input-Output analysis to model scenarios by assuming widespread adoption of the proposed lifestyles changes. Finally, we estimate the domestic and foreign implications for land, water, carbon and human toxicity potential. We distinguish the options with most potential from those that are seemingly fruitless or present backfire risks. While our method allows for testing a large number scenarios under a consistent framework, further work is needed to add robustness to the scenarios. However, we do find a range of indicative results that have strong potential to contribute to mitigation efforts. Services: We find that a local and sharing service economy has a maximum reduction potential of 18% of the European carbon footprint (CF). Clothing & Appliances: Sharing and extending lifetimes of clothes and devices could diminish CF by approximately 3%. Transport: Reducing motorized transport by remote work and active travel could mitigate between 9-26% of CF. Food: Vegan diets could spare 4% of the land and reduce up to 14% of CF. Bio-economy: Switching to biomaterials and bioenergy tend to reduce carbon and toxic emissions at the risk of increasing water and land use. Housing: Passive housing and decentralized renewable energy reduces carbon emissions up to 5 and 14%, respectively. We characterize the sensitivity of our results by modelling income rebound effects and confirm the importance of deterring expenditure in resource intensive goods.

1) Introduction

Sustainable lifestyles can be broadly defined as “living well within earth’s limits”(Jackson 2011; O’Neill et al. 2018). Encouraging sustainable lifestyles is a central strategy towards the 12th UN’s Sustainable Development Goal of “Responsible Consumption and Production”(Akenji and Bengtsson 2014). This goal stems from recognizing that the global environmental crisis is ultimately driven by resource-intensive lifestyles, needs and wants (Vita et al. 2019; Vásquez et al. 2018a).

Europeans live some of the worlds’ most unsustainable lifestyles (Ivanova et al. 2016, 2017). Driven by the level of consumption and living standards, European households emit up to 20 t CO₂ per capita/yr (Ivanova et al. 2017). Only 20% of those emissions are related to household fuels, while most emissions are embodied in consumer products and services (Ivanova et al. 2016, 2017). Further, Europe is a net importer of resources and carbon emissions with about half of its footprint occurring abroad (Tukker et al. 2016). Thus, alternative consumption and lifestyle changes are indispensable to reach environmental goals, especially in wealthy nations (Bjørn et al. 2018a; Rogelj et al. 2018).

Informing the transition to sustainable lifestyles was the main goal of the EU FP7 funded project GLAMURS (Green Lifestyles, Alternative Models and Upscaling Regional Sustainability). From 2014-2017, GLAMURS applied theoretically-based and empirically-grounded frameworks to research the main obstacles and prospects for sustainable lifestyles in Europe (Dumitru et al. 2017) (see glamurs.eu). Empirically, the project compared the lifestyles of average citizens with the lifestyles of members of local grassroots sustainability initiatives (Vita et al. 2018), conducted action research with those local initiatives, and organized backcasting workshops where multiple stakeholders developed visions and pathways towards sustainable lifestyles.

The purpose of this paper is to present a novel approach and analysis related to the environmental impact of sustainable lifestyle options, which was done as part of the GLAMURS project. **The overarching objective of this article is to assess the environmental consequences of lifestyle scenarios obtained from a range of backcasting visions. Our hypothesis is that footprint reductions can be achieved through widespread adoption of sustainable lifestyle options proposed by stakeholders. In the paper, we approve or disprove our hypothesis for each envisioned lifestyle option and discuss the environmental potentials/pitfalls of lifestyles changes.**

We start out from the visions produced during backcasting workshops across several European countries. We identified consumption-related elements from the visions and modeled them as scenarios of changed or reduced household demand. We evaluated the environmental outcomes by running a simulation through the EXIOBASE Environmentally-Extended Multiregional Input-Output Model (EE-MRIO) (Moran et al. 2018; Wood et al. 2017).

Linking qualitative methods to global models of consumption and resources allows us to compare stakeholder views with the environmental and social consequences implied in social change. Naturally, such a modelling effort is subject to at least two considerations. First, there is no standardized methodology to translate from narratives to quantitative modelling (Kemp-Benedict 2004; O'Brien et al. 2014). Although backcasting is common in scenario analysis (O'Neill et al. 2017; Schanes et al. 2019), it is not commonly linked to life-cycle oriented modelling due to the complexity of both, the demand of current lifestyles and the global supply chains serving this demand. Whilst EE-MRIO databases are becoming increasingly detailed and capable of providing product-specific results, such analyses are generally indicative rather than very precise.

Second, economy-wide scenario modelling are typically meant either to predict or characterize counterfactual developments (Distelkamp and Meyer 2019; Bjørn et al. 2018b; Rogelj et al. 2018). This is not the case of backcasting scenarios, where stakeholders normatively describe their visions of sustainability -regardless of expert judgments about “feasibility”. Thus, backcasting scenario evaluation is meant to characterize the broad implications of a vision. Here, the results should be regarded as a first iteration that provides a sense of direction and magnitude of environmental consequences of lifestyles options.

Our modelling decisions follow recent parametrization approaches of scenario simulation with EE-MRIO (Moran et al. 2018; Wood et al. 2017), whilst giving more weight to the stakeholder visions. To strengthen our quantitative evaluation, our scenarios do not model changes in single goods, but rather reflect a bundle of goods associated to a particular lifestyle choice.

This paper seeks to inform the transition to sustainable lifestyles by combining participatory modelling with Multiregional Input-Output Analysis to evaluate a range of scenarios that : 1) Reflect the lifestyles envisioned by different stakeholders 2) Characterizes sufficiency and green

consumption alternatives assuming widespread adoption of sustainable lifestyles, and 3) Discuss the implications for environmental footprints and quality of life of different scenarios.

1.1 Overview of sustainable lifestyles, green consumption and sufficiency

Recent efforts explore demand-side options for reducing consumption (**sufficiency**) or consuming less polluting goods (**green consumption**) (Schanes et al. 2016; Girod et al. 2014; Wynes and Nicholas 2017; Dietz et al. 2009; Gardner and Stern 2008; Bjørn et al. 2018a). Most studies point to plant-based diets, conserving energy, curtailing travel and living car-free as the most promising actions to reduce impact while enhancing human well-being (Schanes et al. 2016; Girod et al. 2014; Wynes and Nicholas 2017; Dietz et al. 2009; Gardner and Stern 2008; Ivanova et al. 2018; Ahmad et al. 2017; Westhoek et al. 2014).

Sufficiency scenarios represent lifestyles that seek to reduce material consumption and aspire to a higher quality of life (Jackson 2005). Sufficiency assumes that once basic needs are satisfied, well-being relies more on health, social relationships, time affluence, and other factors (O'Neill et al. 2018; Vita et al. 2019). Sufficiency lifestyles are supported by the proposal of voluntary simplicity (Jackson 2005) and align with alternative economic models such as de-growth or steady state (D'Alisa et al. 2015; Steinberger and Roberts 2010; Brand-Correa and Steinberger 2017). Sufficiency or “de-growth” assumes the satisfaction of human needs through material and non-material needs in a steady state economy (Vita et al. 2019). While a sufficiency paradigm lowers the risk of rebound effect of monetary savings, it also implies employment challenges such as shorter working hours and the necessary adjustments to protect livelihoods.

By contrast, **green consumption** stands here for consumption that relates to “green growth” economic models (Lorek and Spangenberg 2014). The main assumption is that economic growth may be compatible with sustainability, due to increasing eco-efficiency via technological improvement, servicing and shifting to a circular economy (Akenji 2014). Green consumption options rely on clean technologies (e.g., renewable energies, biotechnology) and reducing waste by closing material cycles as much as possible through extending lifetimes, re-use, retrofit, remanufacturing, and recycling (Steen-Olsen and Hertwich 2015). Under this paradigm, people aspire to a sustainable use of resources without needing to change current lifestyles and economic practices in a fundamental way (Akenji 2014).

Demand-side policies aim to incentivize sustainable lifestyles through behavioral ‘nudges’ and infrastructures that encourage sufficiency or green consumption (Creutzig et al. 2018; Ürge-Vorsatz et al. 2018). However, the whole spectrum, scale and effectiveness of demand-side solutions remains understudied (Creutzig et al. 2018). A broader perspective would include radical lifestyles changes, typically founded on needs-centered views on well-being (Vita et al. 2019), new social norms (Nyborg et al. 2016a), grassroots innovations (Vita et al. 2018), shared economies (PWC 2015) and others (see (Creutzig et al. 2018; Jackson 2005; Baumann and Vita 2015; Akenji 2014; Wiedenhofer et al. 2018)).

Unlike top-down deployment of low-carbon technologies or economic instruments (Wiebe 2016; European Commission 2014), policies for lifestyle changes require of citizens’ engagement and approval in order to succeed (O’Brien 2015; Nyborg et al. 2016b). Even benevolent top-down policies that do not resonate with the target group are bound to generate resistance, be costly or even create social distress (Sekulova et al. 2017). Further, non-participative public planning restricts the communities’ role in launching initiatives to tackle social and environmental challenges (O’Brien 2015; Sekulova et al. 2017).

1.2 Participatory visioning and economy-wide modelling for scenario assessment

Backcasting can be used as a participatory process suitable to embed stakeholder and citizens’ views into decision making (Vergragt and Quist 2011; Quist et al. 2016b). It literally means “looking back from the future” and when done in a participatory way consists of collectively envisioning a desirable future and paths forward to get there (Robinson 1990). Planning through backcasting can smoothen tensions between top-down policies and the actual needs of citizens and stakeholders (Vergragt and Quist 2011; Quist and Vergragt 2006).

Participatory modelling has gained popularity, with the long-overdue recognition that involving stakeholders is key in addressing socio-ecological issues (Brand-Correa et al. 2018; Jordan et al. 2018; Carlsson-Kanyama et al. 2008). The challenge is to find a balanced tool that is supportive of, and supported by, stakeholders while providing comprehensive and transparent insights of the implications of different pathways (Jordan et al. 2018).

Studies on demand-side options often vary in scope and methods, hindering comparisons or meta-studies (Hertwich 2005b; Hertwich and Katzmayer 2004; Schanes et al. 2016). Assessing options through a consistent economy-wide model allows for: 1) Considering global supply-chains and

trade, 2) Aggregate effects at the European level while isolating household potential 3) Product granularity to build specific scenarios 4) Comparison between scenarios and with respect to status-quo baseline 5) Multi-criteria assessment of trade-offs and synergies by comparing multiple resource and emission footprints.

Understanding the global impacts of the sustainable lifestyle scenarios is not a trivial task in today's globalized economy. Could upscaling the envisioned changes lead to footprint reductions? We use EXIOBASE (Wood et al. 2015), a state of the art EE-MRIO, to evaluate the scenarios' potential to mitigate footprints of land, water, carbon and human toxicity. We employ a multi-indicator dashboard to discuss potentials and pitfalls of scientifically assessed and stakeholder-inspired, visions of sustainable lifestyles.

2) Method: Environmental Assessment of alternative consumption scenarios

In this paper, we expand the spectrum of options for sustainable lifestyles while involving stakeholders' views. We selected visions of sustainable lifestyles produced by European citizens, sustainability frontrunners, public managers, and other stakeholders compiled in the GLAMURS project (Quist et al. 2016b, 2016a). We then translated the qualitative scenarios into an EE-MRIO framework, which made it possible to systematically quantify and compare the environmental implications of a range of sufficiency and green consumption scenarios.

Figure 1 summarizes the procedure and methods used in this research. We conducted backcasting workshops where stakeholders described visions of sustainable lifestyles. We then identified the visions that imply alternative consumption scenarios and the goods that would need to change or reduce in each scenario. We use the backcasting information to parameterize our model in terms of whether the changes occur only in household consumption, or also in production recipes and which is their adoption rate. We then simulate the scenario as a "shock" with economy-wide effects (Wood et al. 2017). Finally, we calculated the environmental consequences and compared them to current European impact in order to determine the potential of realizing such scenario.

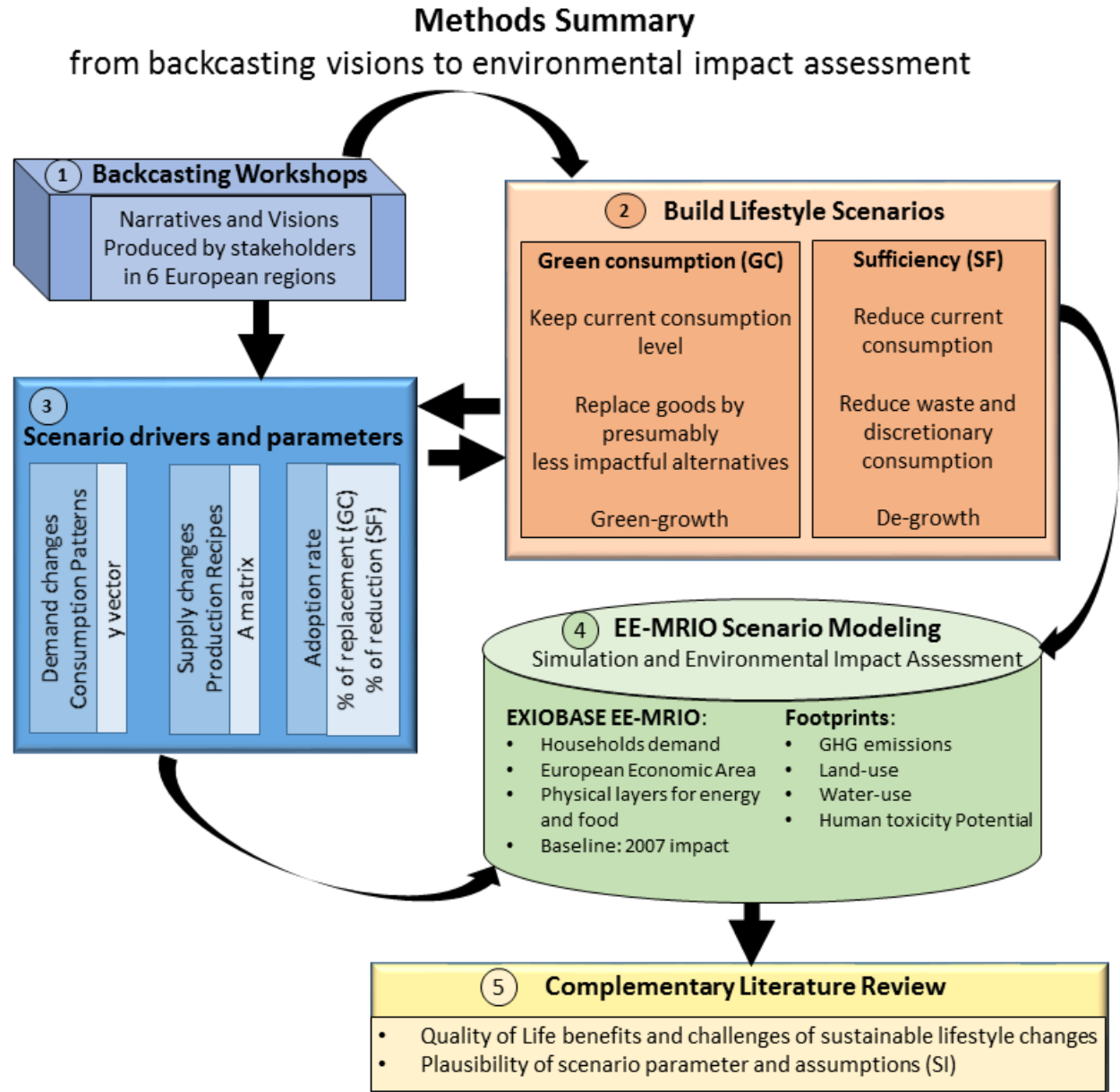


Figure 1 Schematic illustration of steps and framework to model the environmental impact of envisioned consumption scenarios from backcasting workshops. More detail on the steps to translate from qualitative backcasting to quantitative scenarios available in the Supplementary Information (SI).

From backcasting visions to lifestyle scenarios

The data to build consumption scenarios derives from the project GLAMURS, an interdisciplinary research project on sustainable lifestyles (Dumitru et al. 2017). Two backcasting workshops with typically 30-40 participants were conducted in each study region (Table 1): Banat Timis, Romania; Halle, Germany; Danube-Bohemian Forest, Austria; Galicia, Spain; Lazio and

Rome, Italy; and the Rotterdam-Delft-The Hague metropolitan region, the Netherlands (Quist et al. 2016b).

During two series of visioning and backcasting workshops, stakeholders from different societal spheres, including civil society, policy, knowledge and business developed and discussed visions for sustainable lifestyles in the future, including lifestyles changes. More details about the backcasting workshops and their participants can be found in reports of the GLAMURS project (Quist et al. 2016a, 2016b; Dumitru et al. 2017).

Table 1 List of backcasting workshop participants by country and type of participant. The table summarizes two workshops that produced reports (Quist et al. 2016a, 2016b; Dumitru et al. 2017) which constitute the basis of our analysis. NB: Romania had less participants due to weather events.

Total	Austria	Germany	Italy	NL	Romania	Spain
Nr. of participants	32	35	31	37	15	41
Business	10	10	0	0	3	0
Civil society	5	15	16	18	2	18
Government	14	4	3	4	2	11
Knowledge	3	5	12	14	9	12
Other	0	1	0	1	0	0

For the work reported in this paper the backcasting vision reports were scanned for statements proposing lifestyles options that involve consumption changes. We then classified according to their consumption category (e.g., food, transport, etc.). We interpreted the visions statements as literally as possible to set up consumption scenarios that are explicit about the goods and services that would decrease, increase or substitute each other. For example, to model scenarios based on statements such as “clothes will be produced locally and with low transport,” we reduced transportation requirements of the clothing sectors (“*Local Clothing*”) and quantified the environmental consequences. Another example is a scenario where all food would be vegan or vegetarian, meaning full replacement of animal products. This modelling decision implies that our analysis does not show a “feasible” reduction but rather the “maximum potential” of mainstreaming such a lifestyle.

Despite a great amount of sustainable lifestyle options proposed by stakeholders, we could only model those that can be translated into “alternative consumption options”. Text excerpts from

the backcasting reports that were used to build scenarios are provided in Supplementary Information (SI).

We further identified whether the vision corresponds to a sufficiency scenario – implying net reductions in consumption– or green consumption –implying consuming more eco-efficient alternatives. We end up with 19 sufficiency scenarios, 17 green consumption. Additionally, the researchers introduced 5 sensitivity scenarios, to provide a contrast to some of the sustainable lifestyle scenarios.

2.1 Footprints and Database

We use an environmentally-extended input-output framework to calculate the current environmental pressures of European consumption as a baseline (year 2007), and then compare it with the resulting footprints from the modelled scenarios. Environmental footprint, \mathbf{fp} , represents the total consumption impacts from European households. We calculate \mathbf{fp} as a function of household demand, \mathbf{y} , as follows:

$$\mathbf{fp} = \mathbf{s}(\mathbf{I} - \mathbf{A})^{-1}\mathbf{y} + \mathbf{dhe} \quad (1)$$

where \mathbf{s} is the intensity coefficient vector resulting from dividing the total resource or emission required for the production of a given good by its economic output (e.g. CO₂/EUR), \mathbf{I} is the identity matrix and \mathbf{A} is the technical coefficient matrix, representing the inter-industry requirements. The \mathbf{dhe} vector represents direct household emissions from the combustion of fuels for transport, cooking and heating.

Our modelling is based on EXIOBASE2, an Environmentally Extended Multiregional Input-Output (Wood et al. 2015) database. EXIOBASE2 represents the production and consumption of 200 economic goods for 43 countries and 5 rest-of-world regions for the year 2007. Satellite accounts for resources and emissions are available for each sector and country. For each footprint, we consider the resources and pollutants in **Table 2**. Our unit of analysis is the final demand of households of the European Economic Area, hereafter referred as Europe. See SI for details on countries included and EXIOBASE2 coverage.

Table 2 Environmental footprints, including factors of productions and chemicals covered.

Footprint	Coverage	Unit
Carbon Footprint	Global Warming Potential of CO ₂ , CH ₄ , N ₂ O (combustion and non-combustion) and SF ₆ . Includes direct household emissions (GWP 100, IPCC 2007).	Mt CO ₂ equivalent
Human Toxicity Potential	NO _x , NH ₃ , dioxins (PCDD_F), HCB, PM10, As, Cd, Cr, Cu, Hg, Ni, Pb, Se, Zn and SO _x (combustion and non-combustion). Non baseline characterization factors (CML, 2001)(CML-Leiden University).	Mt 1,4-dichlorobenzene-equivalent
Land Footprint	Total land use: forests, pastures and arable land	M km ²
Water Footprint	Total blue water consumption. Includes direct household water consumption.	Km ³

2.2 Modelling consumption changes with EE-MRIO

The global EE-MRIO described above accounts for different production recipes, trade supply chains and household consumption patterns across nations. The parameters that ultimately drive the scenarios are changes in consumption, production recipes and uptake rates (Figure 1). The basis of the model to simulate backcasting scenarios is to perturb the EE-MRIO by modifying the consumption patterns in the \mathbf{y} vector or production recipes in the \mathbf{A} industry matrix (Wood et al. 2017). The magnitude of the perturbations follow the uptake rates stated in Table 3. The full mathematical model to simulate changes in consumption using an EE-MRIO has been adopted from Wood et al. (Wood et al. 2017).

Here we model visions of alternative consumption patterns in households (\mathbf{y} vector of final demand per product), and/or changes in industrial recipes (\mathbf{A} matrix of technical coefficients). We assume a regular functioning of welfare institutions (health, education, pensions etc.) by holding all services provided by governments and social institutions (NPISH) constant.

We model three types of scenarios (Wood et al. 2017):

1. Change in households' demand (Change in \mathbf{y}): Either a reduction in consumption or consuming different goods. In both cases, the scenario modelling consists of simulating a demand change in the relevant goods.
2. Change in industries' demand (Change in \mathbf{A}): When the envisioned scenario depends on changes in inter-industries production recipes and inputs. For example, to produce *Natural Fibres* implies reducing the inputs of synthetic textiles to the apparel sectors.
3. Change at both households' and industries' demand (Change in \mathbf{A} and \mathbf{y}): Some scenarios entail simultaneous changes in household demand and industrial practices.

For example, adopting vegetarian diets would imply that households reduce their purchase of meat directly (y) but also that restaurants have less demand for meat products (A).

While sufficiency scenarios imply a net reduction in the consumption of specific goods, green consumption scenarios imply that the reduced consumption of one product (i) is substituted by increasing the demand of another product (g). As substitute, products may differ in price or energy content per functional unit, the extent of replacement is affected by the relative differences (p) between the products, with no differences having a unitary value.

Expenditure was kept as the monetary functional unit for most services and aggregated product categories, as no physical layer could be derived. The original model allowed for price differences in product substitutes but did not explicitly consider the physical utility delivered by goods (e.g., energy use, calories provided) (Wood et al. 2017). In this research, we enhanced the model by introducing a physical layer to balance food and energy goods to ensure food and energy sufficiency in our scenarios.

For food and energy, which make up nearly half of the EXIOBASE 2 goods, prices underlying the EXIOBASE 2 model (Wood et al. 2015) were used to convert to mass or volume. Further, data on energy content was applied in order to convert to physical functional units i.e. kcal or TJ by weight in kilograms (or by volume in m^3), as explained in the SI and data file. Deriving physical functional units allows us to introduce the current living standards as a constraint by keeping the same level of nutrition (kcal) or energy use (kWh) while shifting the means of provision, as proposed by green consumption scenarios. This allows us to model reductions in food and shelter without falling in a situation of food scarcity or energy poverty.

The differences in prices or energy content per kilogram of fuels and food that modulate product substitution are modelled as follows:

$$p_{ig} = \frac{p_g}{p_i} \quad (2)$$

Where p_{ig} determines the proportion of expenditure shifted in a given scenario. For example, a value of 0.5 would mean 50% of the expenditure of reduced products, i is shifted to increased products, g . This would be the case if a substitute energy carrier delivered twice as dense as the current i.e. double energy per weight. For monetary layers, an example would be buying textiles

for do-it-yourself clothes is five times cheaper than in-store apparel i.e. $p \approx 0.2$. Differences in price and energy densities modulate the substitution share in products demanded by households and industries alike (Wood et al. 2017).

While differences in energy densities are modelled for all food and energy, price differences between substitute goods modelled in monetary terms were rarely assumed, reported in the “price deflator” row in the Supplementary Data modelling parameters. Differentiating price and quality between comparable goods is limited by the product aggregation in EE-MRIO analysis (Girod and de Haan 2010)

Table 3 is a full account of the envisioned consumption scenarios modelled in this paper. The “visions” column describes the actions to achieve sustainable lifestyles articulated by the backcasting workshops participants. Since our goal is to understand the possible environmental outcomes of scaling up the envisioned lifestyles, we assumed aggressive uptake rates to reflect a maximum potential. However, we consider technical or physical limitations when relevant (i.e., food waste cannot be totally eliminated, minimum daily caloric intake (Vásquez et al. 2018b), etc.). Assumptions are detailed in the SI. When pertinent, we model “sensitivity scenarios” to provide an opposite case for comparison. For example, we model *Industrial Materials* as a contrast to a scenario of building with *Natural Materials*. Sensitivity scenarios, however, do not represent stakeholders’ visions.

It should be noted that scenarios of either reduced consumption or reduced inputs to production are applied directly and thus imply a reduction in the GDP of the economy, given that all other variables remain constant (see discussion and limitations). In the discussion we consider economic challenges and quality of life benefits associated with the scenarios. In the SI, we characterize the sensitivity of considering an economic rebound effect for the scenarios that represent monetary savings.

Table 3 Scenarios built from backcasting visions. The values for γ and $-$ parameters indicate the assumed adoption level in household demand or inter-industry demand, respectively, where the value indicates the degree of substitution in the case of green consumption e.g. 1 is full substitution of products. For sufficiency, the value indicates the level of reduction, where 1 represents a total ban of a bundle of goods. (See SI for details on assumptions). Visions marked with * are modelled through physical balances (kcal or kWh) and baseline energy are introduced as a constraint to be kept constant. E.g. Interpretation Key: Animal free clothing proposes a vegan fashion industry, which imply replacing animal textiles with plant-based textiles. This is classified as green consumption (GC) because it keeps clothing consumption constant but with different, presumably more sustainable, materials. The adoption rate is full ($\gamma = 1, A=1$) because it implies a total ban of animal textiles both in household consumption and in industrial recipes.

	Visions	Description	Modelled changes in consumption	SF/GC	γ	A
Clothing	Animal Free	No clothing of animal origin (vegan clothing).	Substitute wool, furs, leather, and replace with textiles/plant-based fibers.	GC	1	1
	Durable Fashion	Reduces textile consumption e.g., clothes swap, second hand use, repairs	Reduces clothes and wearing apparel by 80%. Shift 20% of spending by textile materials (fibers and wool) and leather.	SF	0.8	0
	Natural Fibres	No petroleum-based clothes. Only natural fibres, e.g., wool, fur, cotton	Replace plastic/rubber inputs to clothing sectors with natural fibres by 90%.	GC	0	0.9
	Local Clothing	Only local clothing clothes and fibers.	Reduce by 50% the transport inputs to sectors of clothing and apparel.	SF	0	0.5
Construction	Minimum Construction work	Minimal construction due to large scale co-habitation and downsizing. Only minimal repairs and renovation takes place.	Reduce all construction work and materials by 90%	SF	0.9	0.9
	Repair Renovate	Intensive refurbishment and renovation of existing residential buildings.	Shift 5% of all overall expenditure (except for food) to increase construction work and building materials.	GC	0.5	0.9
	Natural Materials	Building with natural construction materials: wood, clay, stone and sand.	90% decrease in cement, bitumen, metals and foundry work. Increase in wood, clay, sand, stone and non-metallic mineral products.	GC	0.9	0.9
	Industrial Materials	Building and renovation with industrial materials: concrete and metals	Reduce wood, clay, sand, stone and non-metallic mineral products. 90%. Increase in concrete and metals.	SS	0.9	0.9
Food - Diet	Processed Food*	Shift towards more processed food and ready to eat food products.	Reduce all raw and plant-based foods, as well as live animals, by 80%. Replace with processed food products.	SS	0.8	0
	Food Sufficiency*	Limits food consumption to 2586 kcal/day. Reduces food surplus.	Reduce all food product spending by 27%, corresponding to the average surplus calories in Europe (Hiç et al. 2016; Vásquez et al. 2018a).	SF	0.27	0
	Mediterranean Diet*	High consumption of plant-based food, fish, dairy, and wine. Less meat.	Decrease non-fish meat products by 80%, increase all others foodstuff. Hotels and restaurants (H/R) change their inputs.	GC	0.8	0.8
	Vegetarian*	Vegetarian food with dairy and eggs but no meat.	Reduce meat and fish to 100%. Replace with plant-based food, dairy, and processed food. Hotels and restaurants change their inputs.	GC	1	1
	Vegan*	Vegan food (no red/white meat, eggs, or dairy products).	Eliminates all food animal products. Increase all other food. Hotels and restaurants change their inputs.	GC	1	1
	Healthy Vegan*	Vegan food and eliminates processed foods, sugars and beverages.	Eliminates all food animal products, processed food, sugar and beverages. Hotels and restaurants change their inputs.	GC	1	1
Food SC	Local Food	Shift towards locally sourced food, including hotel/restaurant sector.	Reduce transport needs of food industries by 50%.	SF	0	0.5
	Organic Food	Food and animals are produced without agrochemicals.	Reduce fertilizers, chemicals and medicines as inputs to food and H/R products by 100%.	SF	0	1
	Seasonal Food	Less vegetables grown in greenhouses through seasonal consumption	Reduce inputs of fuels and electricity to vegetable sector by 30%.	SF	0	0.3
	Less Waste	Reduce food waste at the household level.	Reduce all food product spending by 12% (Vanham et al. 2015) (corresponding to estimated calories that currently go to waste).	SF	0.12	0
Man. products	Share & Repair	Collaborative ownership of appliances and tools. Second-hand buying/renting, tool library and repair cafés. Shift to services.	Reduced consumption of machinery and electronic apparatus and their retail/trade by 50%. 10% of expenditure shifts go to renting apparatus.	GC	0.5	0
	Offline Minimalist	Less media, Internet, telecommunication equipment etc.	80% reduction of media, machinery, electric apparatus, telecommunication devices and services related.	SF	0.8	0
	Durable appliances	Extended appliance lifetime, increased reparability lowers consumption	80% reduction of general appliances, office equipment devices and precision instruments.	SF	0.8	0
	No Chemicals & Plastics	Reduces use of chemicals and plastic, e.g., bottled beverages, plastic bags	90% reduction of chemicals, fertilizers, cleaning agents, plastics and rubbers at the household.	SF	0.9	0
Mobility	Frequent Flyer	Flies frequently.	Reallocate 2% of all product spending, except on food, towards air transport.	SS	0.02	0
	Cycling & Flying	Cycling increases, reducing land transport but people fly with the savings.	50% reduction of products related to local land mobility, shifting expenditure to air mobility.	GC	0.5	0
	No Flying	Stops flying.	Eliminates all air transport services.	GC	1	0
	Renewable Fuels	Public transport and private vehicles use mostly liquid biofuels.	Substitute 90% of all fossil transport fuels by bio gasoline, biodiesel, ethanol fuels and others. Including direct household mobility. Inputs to land transport services and motor fuel retail industry shift towards biofuels.	GC	0.9	0.9
	Less Cars (50%)	Expanded public transport, car co-ownership and ride share are deployed.	Substitutes 50% of income spent on private vehicles and fuels with land public transportation (bus, train, metro, etc.).	GC	0.5	0
	Less Transport (50%)	Overall decreased mobility, e.g., through digital lifestyles and efficient cities	50% reduction of all products related to mobility.	SF	0.5	0
	Work from Home (50%)	Reduces need for mobility by working from home, telecommute, living close to work, etc.	Reduces spending on mobility by land by 50%.	SF	0.5	0
	Work from Home (50%) ER	Same as "Work from Home" but ER assumes that more time spent at the home could increase electricity and heating needs.	Reduces spending on mobility by land by 50%, increase electricity and heating fuel spending by 20%.	SF	0.5	0
	Bike Walk Full	Bikes/walks everywhere for land commute. Other mobility constant.	100% reduction of vehicles, fuels and services related to mobility by land.	SF	1	0
	Leisure Services	Increased travel agencies, restaurant food, spa, entertainment, etc. Focus on hedonism and disregards insurances and financial security.	80% reduction expenditure in health, education and financial services and instead spends on entertainment, tourism, hotels and restaurant and shopping.	SS	0.8	0
Services	Non-Market Services	Large-scale collaborative economy and inter-community exchanges, voluntary work, time banks and community services.	80% lower use of all services.	SF	0.8	0
	Community Services	Engaged in recreational, sport and cultural organizations, high communication	Decrease leisure services and tourism by 80%, substitutes with recreational and membership organization services.	GC	0.8	0
	Local Services	Local and decentralized service supply. Local economy favors servicing.	Reduce direct household spending on local mobility by 20% (Wiedenhofer et al. 2018). Reduce transport inputs into all services by 30%.	SF	0.2	0.3
	100 % Fossil Fuels*	Replaces household renewable fuels and electricity with fossil fuels	Full replacement of current renewable electricity and energy with fossil sources.	SS	1	0
Shelter	Renewable Electricity*	Renewable electricity by wind, photovoltaic, solar, geothermal and tidal.	Reduce fossil electricity by 100%, replace with renewable electricity.	GC	1	0
	Passive housing	Passive house standard and energy-efficient dwellings.	Reduce energy spending by 43% (Mosenthal and Socks 2015) (i.e. 40% lower energy need). Shifts 20% of consumption to construction work and insulation.	GC	0.43	0
	No energy Ecovillage	Models a pre-industrial energy use while keeping all else constant.	Decrease spending on energy carriers and grid services by 100%. Models the impacts of current electricity and fuel consumption.	SF	1	0
	High-tech Ecovillage	Decentralized, local, small-scale renewable energy production distributed through micro grids.	Decrease spending on fossil based electricity and overall transmission grid services. Substitute with local generation of renewable electricity: solar, hydro, wind, geothermal. All other fossil fuels for heating remain the same.	GC	1	0
	Water Off-Grid	No conventional water distribution. Water use from natural sources.	100% reduced expenditure on collected and purified water, distribution services of water.	SF	1	0

SF= sufficiency (net reduction), GC= Green consumption (shift in consumption), SS= Sensitivity Scenario, ER = Energy Rebound,

3) Results

3.1 Current status of European impact

Table 4 shows the impact intensity per euro spent for detailed consumption categories. Food is the most water and land intensive category, while mobility and shelter are the most carbon intensive (Ivanova et al. 2016). Transport emits the most human toxins per euro, while services have a relatively impacts per EUR. Table 2 serves as a baseline to interpret the scenario modelling results.

Table 4 Average intensities in impact per euro for consumption categories. Calculated as footprint of each product category divided by the total consumption of that category aggregated for Europe. DCB: dichlorobenzene. Own calculation based on EXIOBASE (Wood et al. 2015). Calculations of energy per kilo for food and fuels can be found in the SD.

European environmental intensity of consumption						
	Carbon (kg CO ₂ eq/EUR)	Human Toxicity Potential (kg 1,4-DCB eq/EUR)	Land (m ² /E UR)	Land (m ² /kg)	Water (liter/EUR)	Water (liter/kg)
Clothing and apparel	0.79	0.70	1.70		31.79	
Construction materials and work	0.75	0.49	3.29		8.27	
Food: Processed	1.11	0.62	3.61	10	118.92	333
Food: Dairy	1.45	0.62	4.70	13	80.49	222
Food: Meat and fish	1.44	0.65	3.63	76	94.67	1972
Food: Plant-based	1.35	0.44	7.81	19	292.80	712
Manufactured products: Appliances, machinery and electronics	0.70	0.71	0.51		8.44	
Manufactured products: Media and communication apparatus	0.55	0.57	0.88		9.15	
Manufactured products: Plastic, paper	3.44	4.19	1.38		41.85	
Transport: By air	2.01	0.77	0.38		6.98	
Transport: By land	2.04	0.94	0.49		8.72	
Transport: By water	3.09	122.28	0.48		9.05	
Services: Information technology	0.37	0.30	0.35		5.07	
Services: Business and financial	0.19	0.16	0.17		2.78	
Services: Health, education and research	0.28	0.23	0.47		8.84	
Services: Renting services and real estate	0.18	0.16	0.19		2.30	
Services: Recreation and tourism	0.50	0.58	0.97		25.30	
Services: Trade and retail	0.39	0.54	0.48		8.90	
Housing: Electricity and fuels	4.46	0.66	1.89		12.18	
Housing: Household commodities	1.06	0.70	2.23		16.76	
Housing: Recycling	1.09	1.10	0.48		7.28	
Housing: Waste treatment	1.16	0.40	0.39		6.67	

3.2 Environmental impact assessments of green consumption and sufficiency scenarios

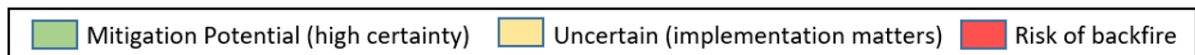
Table 5 summarizes the impact assessments for the envisioned scenarios of green consumption and sufficiency. **Sufficiency** options have higher mitigation potential in the domains of transport, services and clothing, while **green consumption** options show more reductions in the domains of food and manufactured products. We find that large-scale shifts towards plant-based diets, reductions in motorized transport and energy-efficient housing offer the most potential to curb European environmental impacts (Wynes and Nicholas 2017). Reducing manufactured products and clothing hold considerable potential, above 2% across footprints.

While here we contrast green consumption and sufficiency, in practice some of these actions might be complementary. For example, adopting plant-based diets does not exclude preventing food waste or eating organic. For green consumption options, however, the environmental impact of the alternative goods and the volume of consumption, would largely determine the environmental outcome, e.g., the foods chosen to replace meat in diets (Rao et al. 2018).

We mark footprint changes below 2% in yellow to signal outcome where the observed change is relatively small and the practical implementation of such scenario could tip the balance towards reduction or increase. Energy and food scenarios were modelled through a physical energy layers (marked with * in Figure 2 and Table 3) in order to maintain current energy demand (kcal or kWh) and model the isolated effect of shifting food and energy carriers (such as in *Renewable Electricity* or *Vegetarian.*). See SI for modelling of physical layers.

Table 5: Environmental synergies and trade-offs of green consumption and sufficiency scenarios. Mitigation potential (green and positive) or backfire (red and negative) expressed as a percent difference (Δ) with respect to the baseline. Color-coding as follows: yellow: $\Delta \pm 2\%$; light red: $\Delta < -2\%$; dark red: $\Delta < -5\%$; light green: $\Delta > 2\%$; dark green: $\Delta > 5\%$. Yellow color represents small and thus uncertain results. The outcome of these actions would depend on their practical implementation. The values summarize the percentages reported in Figure 2.

Consumption domain	Green Consumption Scenarios	Mitigation potential				Sufficiency Scenarios	Mitigation potential				
		Carbon	Toxicity	Land	Water		Carbon	Toxicity	Land	Water	
Clothing	Animal Free (Ctrl)	-0.8%	-0.5%	-1.2%	-0.5%	Local Clothing	0.5%	1.7%	0.3%	0.5%	
	Natural Fibers	0.0%	-0.1%	-0.3%	-0.3%	Durable fashion	1.8%	2.5%	2.1%	2.1%	
Construction	Repair & Renovate	-0.7%	2.4%	-10.8%	1.0%	Minimum Construction	1.8%	1.3%	3.5%	0.5%	
	Natural Materials	0.5%	0.1%	-1.4%	0.0%	Work					
Food	Mediterranean Diet*	2.7%	0.2%	-0.1%	-0.5%	Food Sufficiency* (Ctrl)	4.9%	2.6%	14.4%	16.0%	
	Vegetarian*	6.4%	3.0%	0.6%	0.2%	Local Food	0.6%	3.6%	0.1%	0.1%	
	Vegan*	13.9%	9.0%	4.7%	14.8%	Organic Food	1.8%	1.0%	0.8%	1.3%	
	Healthy Vegan*	15.7%	12.0%	-2.9%	9.7%	Seasonal Food	0.1%	0.0%	0.0%	0.0%	
Manufactured Products	Share Repair					Less Waste	2.1%	1.1%	5.5%	7.1%	
		Less Chemicals & Plastics	3.9%	4.0%	2.7%	4.4%					
		Offline minimalist	1.5%	2.0%	0.6%	0.6%					
		Durable Appliances	1.5%	2.0%	1.0%	0.7%					
Transport	Less Cars (50%)	8.8%	1.7%	0.8%	0.6%	Less Transport (50%)	14.5%	20.4%	2.0%	1.9%	
	Renewable Fuels	12.1%	1.4%	-5.9%	-5.3%	Work from Home (50%)	13.0%	7.1%	1.9%	1.8%	
	No Flying	2.3%	1.0%	0.3%	0.2%	Work from Home (50%) ER	8.9%	6.1%	-1.0%	1.2%	
	Cycling & Flying (Ctrl)	0.1%	1.3%	0.3%	0.4%	Only Bike and Walk	26.0%	14.2%	3.8%	3.5%	
Services	Community Services		3.1%	23.8%	3.6%	6.6%	Local Services	5.3%	2.9%	0.8%	0.7%
							Non-market Services	17.8%	21.5%	14.6%	15.8%
Housing	High Tech Ecovillage*	7.9%	1.3%	1.7%	0.3%	Low Tech Ecovillage	13.8%	4.9%	4.9%	2.6%	
	Renewable Electricity*	2.9%	0.2%	-3.1%	-0.1%	Water Off Grid	0.5%	0.2%	0.1%	0.1%	
	Passive House	5.6%	1.9%	5.0%	1.1%						



Overall, we find encouraging environmental outcomes from the envisioned consumption scenarios. Switching towards locally sourced, peer-to-peer and community services could mitigate 3-23 % of European environmental impacts. Reducing transport needs, working from home and switching to cycling and walking are options that do not present trade-offs and could mitigate 9-26% of carbon and 2-4% of land and water impacts. Switching to plant based diets has the potential to mitigate between 4-15% across impacts, while reducing food waste and surplus could reduce 2-5% of carbon and save up to 16% of water.

Switching the fibers used in clothing has negligible effects, but making clothes last longer (e.g., through swapping and repairing) could lead to 2% reduction in European impacts. Similarly, sharing and repairing household appliances and devices could yield a 2.5-6% reduction across impacts. Finally, the outcome of alternative housing would depend on the chosen energy carriers. If forestry products are to supply the current heating and cooking needs, carbon emissions could be reduced by 8%, but at the cost of doubling land requirements. Adopting *passive house* standards

or to live at the margins of centralized energy systems show no-trade offs and could reduce 5-14% of European impacts.

The magnitude of our results are in line with previous analyses. Previous assessments associate housing, transport and services to 70% of carbon emissions, while food alone takes up half of the water and land embodied in European consumption(Ivanova et al. 2017, 2016). Clothing, construction, and durable goods together account for about twenty percent of resource use and emissions(Ivanova et al. 2017, 2016). The following section describes results for each consumption category in detail.

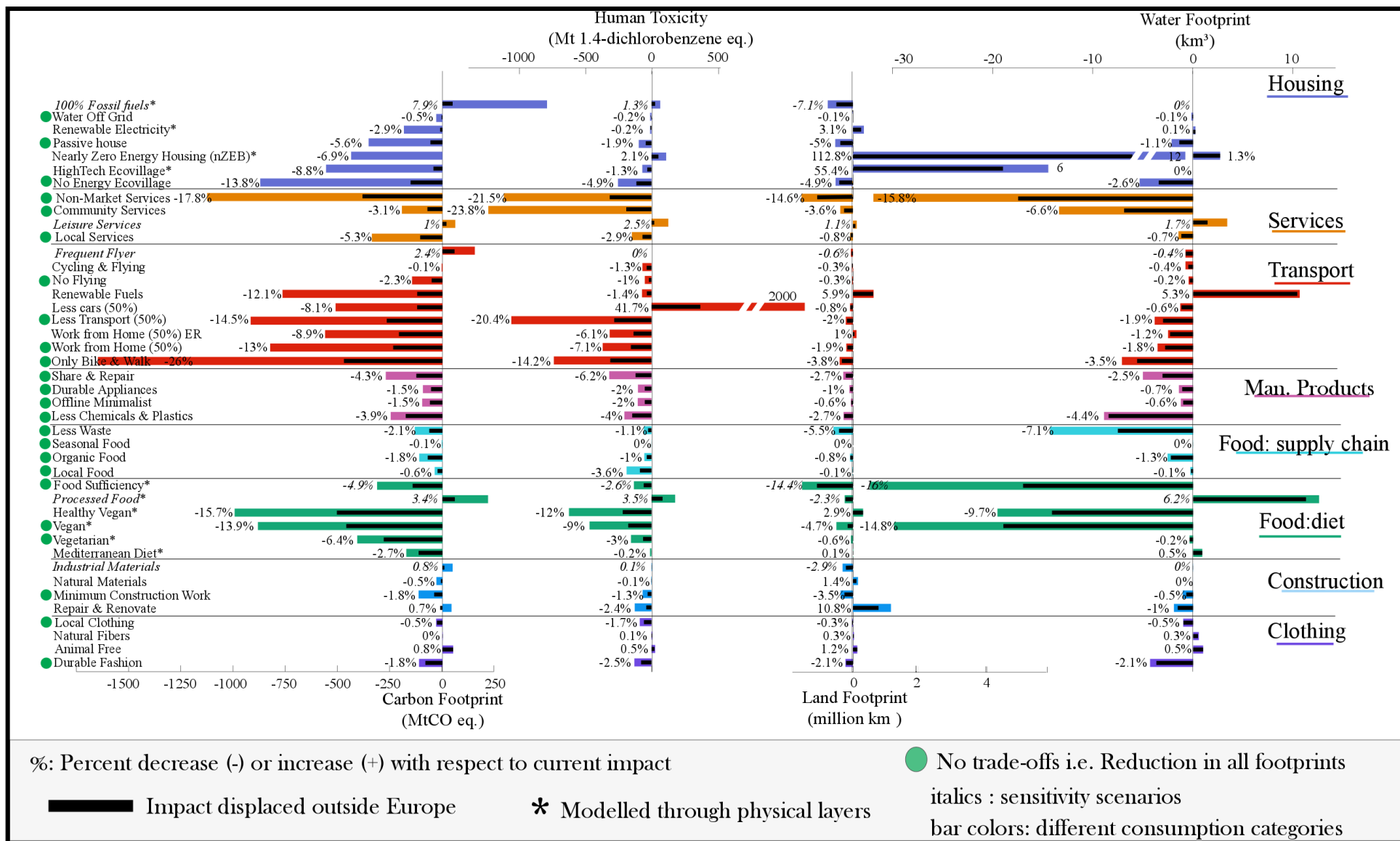


Figure 2 Relative and absolute footprint changes with respect to lifestyle change scenarios. Percent values indicate deviation with respect to baseline: total European household footprints of 2007. Black bars show the impact share that occurs outside the European Economic Area. A green dot indicates the consumption changes that present a positive reduction and no trade-offs across footprints to indicate the “safe options.” Asterisk * indicates lifestyles modelled through physical energy balances of kcal or kWh. ER=energy rebound (see Table 3). To contrast the sustainability visions, we included some worse case scenarios to show the range (indicated by italics).

1

2 ***Clothing***

3 While net reductions in the consumption of clothing and construction may curb impacts, simply
4 shifting materials offers modest reductions with possible trade-offs, as shown in Figure 2.

5 Durable Fashion could halve current impact of clothing, reducing the environmental of
6 Europeans by 1.8-2.5% by extending clothes' lifetimes and increasing secondhand re-use.

7 Lowering clothes miles by preferring Local Clothing reduces human toxicity by 1.7% due to the
8 high toxicity of transportation fuels (Table 4). with marginal reductions in other footprints
9 (Figure 2). Replacing all synthetic fibers with *Natural Fibers* has a negligible mitigation potential
10 across footprints. Phasing out animal fibers for plant-based and synthetic fibers would require
11 **1.2% more land and 0.5% more water** as shown by the *Animal free* clothing scenario. Choosing
12 natural over synthetic clothing materials present negligible carbon reduction potentials with
13 potential increases in other footprints. In sum, only sufficiency scenarios of net reductions in
14 clothing offer mitigation potential.

15 ***Construction***

16 Co-habitation and downsizing of living spaces could Minimize Construction Work, reducing land
17 and carbon footprints by 3.5 and 1.8%, respectively. Intensive Repair & Renovation could increase
18 land use about 11% and slightly reduce other footprints, due to the lower intensity of construction
19 goods with respect to other categories (Table 4).

20 Using more *Natural Materials* in construction results in a carbon reduction of 0.5% but a land
21 increase of 1.4%. *Natural Materials* such as wood, stone, sand and clay require more land but emit
22 less carbon since they require less processing and energy compared to concrete and metals. We
23 model the opposite case in *Industrial Materials* by building with concrete, steel and aluminum.
24 This would decrease land by 3% while increasing carbon footprint by 0.8%. Although
25 construction is not typically associated with lifestyles, 70% of Europeans households own their
26 dwelling(Eurostat 2018) and thus can influence the energy efficiency and materials in their
27 houses. Renovation for thermal performance could decrease energy use per area but expansion of
28 current living spaces would have the opposite effect (Vásquez et al. 2016).

29 As with clothing, the choice of natural over synthetic materials in construction shows a negligible
30 potential reduction in carbon, toxicity and water accompanied by potential increase in land.
31 Again, only sufficiency scenarios in construction offer considerable mitigation potential.
32 Noteworthy, wood materials are rather intensive in “forest land”, while natural fibers rely mainly
33 on croplands (e.g., cotton) (Table 4, Table 2).

34 ***Food: Diets***

35 All low-meat diets provide significant environmental footprint reductions (Figure 2). A
36 *Mediterranean Diet* would lower non-fish meat and increase legumes, oils, vegetables, cereals, fish
37 and dairy, and could reduce carbon emissions by 2.7% at the cost of a slight increase of land and
38 water. A full *Vegetarian* diet would reduce carbon and toxicity by 6.4 and 3.0%, respectively.
39 Removing dairy products and eggs (*Vegan* lifestyle) yields a reduction potential of carbon (14%)
40 and of toxicity and water footprints of 9 and 15%, respectively. With a *Healthy Vegan* diet (reduced
41 sugar, beverages and other processed food products), the carbon and toxicity footprints would be
42 decreased by 16 and 12%, respectively. The slight land footprint increase for *Healthy Vegan* lies in
43 the low price but relatively high calorie of unhealthy vegan foods such as sugar and beverages.
44 Supplying calories with sugar requires less total land than supplying the same calories with oils
45 and nuts, for example. This result is not conclusive, and in practice the outcome would depend on
46 the food products that constitute a *Healthy Vegan* diet (Rao et al. 2018).

47 We model the sensitivity scenario of *Food Sufficiency* by limiting the calorific intake to a sufficient
48 amount for European standards of 2586 kcal/day (O’Neill et al. 2018) and find that such measure
49 may reduce the total carbon footprint by 4%, twice the potential found by a prior study of
50 France (Vieux et al. 2012). *Food Sufficiency* yields a decrease in total agricultural land needed; the
51 water and land footprints may decrease by 16% and 14%, respectively. Our results agree with
52 previous findings that show 20% of European food is supplied in a surplus, which in turn largely
53 drives waste and overeating (Hiç et al. 2016). The *Processed Food* simulates a higher intake of
54 processed food and lower intake of plant-based and staple foods. This would increase all footprints
55 except land, for a similar reason as discussed above with respect to the *Healthy Vegan*, but also
56 because supplying current caloric needs exclusively through *Processed Food* would come at a
57 greater cost, and thus prevent expenditure in other products (see “physical layers” in SI).

58

59

60 ***Food: Supply chain***

61 *Organic Food* could reduce carbon (1.8%), land (0.8%) and water (1.3%) while *Local Food* reduce
62 toxicity footprint (3.6%) due to lower transport needs. The scenario of more *Seasonal Food*, where
63 energy inputs to agriculture reduce by 30% (Girod et al. 2014), has no significant mitigation
64 potential. Europe consumes a large share of imported food, and agriculture requires relatively
65 low energy inputs. However, in a scenario where a larger share of food is produced within Europe,
66 the effects of seasonal food might be more significant.

67 We confirm previous findings of *Organic Food* having lower impact than consuming *Local Food*
68 which reduces food miles (Avetisyan et al. 2014). However, when we add human toxins to this
69 debate, we find that *Local food* is preferable for reducing toxicity in Europe. Policies to favor
70 synergies between *Organic*, *Seasonal* and *Local* agriculture could lead to dynamic effects that yield
71 potential beyond our estimates (Westhoek et al. 2014). *Less Waste* would imply reduction of food
72 consumption by 12% (Vanham et al. 2015) (1.2% of total household expenditure). Our results
73 agree with previous estimates of at least 2% of European carbon to be food waste (Hoolohan et
74 al. 2013) and are within the 2-7% range reported by Usubiaga et al, based on EXIOBASE
75 (Usubiaga et al. 2018). Indeed, we find reducing food waste can reduce by 5.5 and 7% the use of
76 land and water, half of it outside Europe.

77 Combining sustainable diets and supply chains could yield further reductions. A *Vegan* diet with
78 *Less Waste* and *Organic Food* could potentially reduce footprints of up to 18, 11 and 24%, for
79 carbon, land, and water, respectively. Our general findings agree with previous research that
80 reports low-meat diets (Tukker et al. 2010; Rao et al. 2018; Wynes and Nicholas 2017; Schanes
81 et al. 2016) and organic food (Reganold and Wachter 2016; Hoolohan et al. 2013) have lower
82 environmental impact than conventional diets. In sum, we find most reduction potential by
83 shifting to non-meat diets, while reducing food waste and miles yield lower, yet considerable,
84 reduction potentials.

85 ***Manufactured products***

86 *Share & Repair* reduces carbon by 4.3% and toxicity by 6%; assuming increased sharing,
87 reparability, re-use and product-service systems. The scenario of *Durable Appliances* and *Offline*
88 *Minimalist* show comparable reduction potentials. *Durable Appliances* extends useful lives of
89 appliances while *Offline Minimalist* reduces personal electronic devices and media consumption to

90 offer a reduction of 1.5 and 2% for carbon and toxicity, respectively. A scenario of *Less Chemicals*
91 & *Plastics* entails lowering household chemicals and plastics, with a 4% reduction potential in
92 carbon. Reducing chemicals reduces the pressures of foreign land and water, while *Share & Repair*
93 has a significant reduction of carbon and toxicity within Europe.

94 ***Mobility***

95 Replacing all local land transport with biking and walking (*Only Bike Walk*) can potentially reduce
96 carbon by 26% and toxicity by 14%. *Work from Home* implies mainstreaming flexible and remote
97 work, thereby halving current commutes and reducing carbon and toxicity by 13% and 7%. If
98 *Work from Home* becomes widespread, there is a risk of increased use of fuel and electricity at
99 home. We estimate such possibility in *Work from Home ER* at mitigation potentials of only 9%
100 carbon and 6% toxicity. Such rebound could be counteracted by energy efficient housing or
101 decentralized working spaces that workers can reach without motorized transport.

102 Similar to others, we find that shifting to public transport is efficient in reducing carbon (Duarte
103 et al. 2016; Wynes and Nicholas 2017). *Less Transport* implies 50% reduction in all motorized
104 transport, thereby reducing toxicity (20%) and carbon (14%). The *Less Cars* scenario models a
105 large adoption of car-free lifestyles, implying a 50% expenditure shift from private vehicles
106 towards collective transport and shared vehicles. This could reduce carbon up to 8.8% and
107 toxicity by 1.7%. By modelling transport through a top-down MRIO, we do not consider the
108 demand of passenger-kilometers directly. Since 80% of current European commute is done with
109 passenger-cars (Eurostat 2014), shifting monetary demand from private to public transport could
110 lead to a surplus of passenger-kilometers, e.g., more buses, trains and ferries. Thus, bottom-up,
111 country-specific data on fleet inventory and passenger-kilometers by transport mode would
112 increase the accuracy of the model.

113 Adopting *Renewable Fuels* for mobility potentially decreases carbon (12%) and toxicity (1.4%),
114 with the risk of increasing pressures on foreign land and water by 5.8 and 5.3%. This result
115 stresses the importance of considering consequences abroad in policies such as the EU 2020
116 energy strategy (European Commission 2014). *No Flying* could reduce carbon by 2.3% while the
117 sensitivity scenario of *Frequent Flyer* shows that carbon could increase by 2.5%. Shifting demand
118 from other goods towards flying frequently would actually reduce the land and water footprint,

119 due to relative low water and land intensity, and high price of air travel, compared to other goods
120 (Table 4).

121 *Cycling and flying* portrays a scenario of commuting by walking, cycling and public transport but
122 flying with the savings. We find that the carbon reductions of active transport would be offset by
123 the rebound effect of flying, with the risk of increasing toxic emissions by 3%. This result suggests
124 that air transport should be discouraged as active transport is encouraged, to prevent a rebound
125 effect.

126 ***Services***

127 The *Local Services* scenario portrays a lifestyle that mostly takes place within the neighborhood.
128 It entails a moderate reduction of short distance mobility coupled with preference for locally
129 sourced services that require less transport logistics. Favoring *Local Services* could reduce carbon
130 (5.3%) and toxicity (3%) footprints. The lifestyle of *Community Services* portrays reduced tourism
131 and leisure to be more engaged in recreational, sport and cultural organizations. Citizens would
132 be active in community organization and communications, leading to a reduction of toxicity (24%)
133 and water (6.7%) due to a combined effect of reduced transport needs and shifting toward services
134 with lower impact intensity, such as organizations and club membership.

135 Non-market Services envisions communities where citizens largely supply each other with
136 services through collaborative economies, voluntary work, time banks and community services,
137 reducing all impacts by 15-20%. Even if services are less impactful per euro compared to physical
138 goods (Table 4) their consumption volume makes them relevant for impact mitigation, as shown
139 by *Community Services*.

140
141 [Scenario of non-market economy models possibilities of nearly zero marginal cost to produce](#)
142 [goods and services supported by global collaborative commons and internet of things \(Rifkin](#)
143 [2015; Grubler et al. 2018\)](#). The premise of such a self-provision scenario relies on regional
144 [exchange networks organized towards satisfying most needs of their members and even use their](#)
145 [own alternative currencies \(Sekulova et al. 2017\)](#). This is the premise of the gift economy and
146 [conviviality tools \(Sekulova et al. 2017; Dumitru et al. 2016; Illich 1971\)](#). However, this result
147 should be interpreted cautiously because switching to *Non Market Services* would imply economic
148 de-growth and possibly lower incomes, which are macroeconomic effects beyond our scope.

149
150 *Leisure services* is a sensitivity scenario to contrast *community services*. We find that increasing
151 *Leisure Services* would slightly increase current footprints by shifting expenditure in health and
152 education towards entertainment, tourism, restaurants and shopping. The results suggests
153 market-based leisure and entertainment are more impactful than health, education, pension
154 services, etc. While the latter arguably contribute more to the common good and quality of life
155 (Stiglitz et al. 2010). Modern economies rely on stimulating the demand for market leisure and
156 entertainment due to their profitability (Debord 1994; Druckman and Jackson 2010).
157 Nevertheless, leisure could potentially be more satisfied through non-market, low-carbon, options
158 (Vita et al. 2019; Druckman and Jackson 2010).

159 ***Shelter***

160 *Renewable electricity* shows that shifting remaining fossil fuels to renewable electricity would lead
161 to increased land and water while decreasing carbon footprint by 3%. We interpret this result
162 with caution, as the scenario assumes the European renewable energy mix for 2007, where
163 hydropower held a major share, but the outcome might be different with larger contributions
164 from solar and wind. Previous findings confirm that large scale hydro-power and biofuels are
165 land and water intensive

166
167 Consequently, switching to *100% Fossil Fuel* would decrease land but increase carbon, reflecting
168 the freeing up of land currently used to supply hydropower and biofuels.

169 *Passive Housing* could potentially save 6% carbon and 5% land by reducing space heating by 40%
170 through renovating for energy efficient dwellings. The efficiency potential was estimated by
171 comparing current statistics on European space heating needs (European Energy Agency 2010)
172 to the passive house standard (15 kWh/(m²yr) passive), according to previous approaches
173 (Mosenthal and Socks 2015) (see SI).

174 *A HighTech Ecovillage* simulates self-sufficient and decentralized renewable electricity generation.
175 This scenario leads to a reduction of 7.9% of carbon and modest reductions, between 0.3-1.7%, in
176 other footprints. A *HighTech Ecovillage* fits the idea of an urban ecovillage, which reduces the
177 share of fossil fuels and the impact of grid services and transmission. *No energy Ecovillage* portrays
178 off-grid settlements with radical net reductions that eliminate all need for market energy. This

179 could reduce carbon by 14% and land by 5%, which corresponds to the baseline impact of
180 household energy. This scenario simulates pre-industrial lifestyles with respect to energy while
181 keeping other consumption constant. The proponents of this vision mentioned zero energy
182 constructions (e.g., bio-constructions, solar heaters, biogas digester, etc.) in order to maintain
183 decent living standards (SI data) (Omann et al. 2016).

184 Supplying *Water off-Grid* through natural sources offers slight impact reduction. This is due to
185 the large role of government subsidy in water infrastructure and supply. Even if eliminating
186 centralized water supply might be unrealistic today, recent studies signal the opportunity of
187 replacing engineered grey infrastructure by natural infrastructures to enhance water capture,
188 availability and quality (Palmer et al. 2015).

189

190 **4) Discussion**

191

192 The construction of scenarios is a key activity in sustainability studies and related policy
193 development (Huppmann et al. 2018; van Vuuren et al. 2017a; Grubler et al. 2018). While most
194 resource-assessment scenarios deal with hypothetical trajectories of development (O'Neill et al.
195 2017; Riahi et al. 2017), only few focus on the potential of demand-side solutions (Grubler et al.
196 2018; Creutzig et al. 2018; Schanes et al. 2016) and even fewer build on the views of non-academic
197 stakeholders (Jordan et al. 2018; Carlsson-Kanyama et al. 2008). Paradoxically, the sustainability
198 scenarios that meet a 1.5°C climate target rely heavily on mainstreaming sustainable lifestyles
199 (Grubler et al. 2018; van Vuuren et al. 2017b; Riahi et al. 2017). Hence, identifying and supporting
200 lifestyles that are environmentally sound and socially accepted is key for current mitigation and
201 adaptation challenges (Ürge-Vorsatz et al. 2018; Riahi et al. 2017).

202 In this study, we built scenarios based on stakeholders' visions of sustainable lifestyles to
203 distinguish the options with most potential from those that are seemingly fruitless or present
204 backfire risks. By simulating scenarios in an economy-wide model, we identified that the most
205 promising sufficiency scenarios (net consumption reductions) are curtailing motorized transport,
206 reducing market services via the shared economy, conserving energy, reducing food waste or
207 surplus and increasing durability of clothes and devices. Green consumption (consumption
208 changes) show most potential in shifting towards plant-based diets, sharing and repairing

209 appliances, retro-fitting insulation for passive housing and replacing market leisure and
210 entertainment for community-oriented, cultural and sports services.

211 4.1 Strengths and Limitation

212 Modelling through an EE-MRIO enables a high-throughput evaluation of different scenarios
213 under a harmonized framework, through a global life-cycle perspective, and considering multiple
214 environmental criteria. The drawback is that our results are only indicative and further scenario
215 development as well as refining modelling options within each consumption domain could yield
216 results that are more precise.

217 In most MRIOs single products or goods entail higher uncertainty, specially those with relatively
218 small values (Moran and Wood 2014). In this article we mostly model consumption goods
219 bundles (e.g. food products) and in few cases large single products (flying). EXIOBASE is one of
220 the MRIOs with higher product-resolution and its advantages have been previously discussed
221 (Wood et al. 2015). The uncertainty inherent to MRIOs is well characterized and tackling this
222 shortcoming is an effort of the wide IO community as these databases mature (Moran and Wood
223 2014; Min and Rao 2017; Rodrigues et al. 2018).

224 In our paper, the “modelling choices” derive from stakeholder interaction. As such, we do not aim
225 to “improve” their visions but to evaluate their environmental performance. The advantage of
226 striving to a faithful representation is that scenarios are traceable to the visions reported in the
227 backcasting reports. Further, assessing all visions as a whole provides a comprehensive and
228 transparent first indication of the spectrum of sustainable lifestyles and the relevant options.
229 Future applications that focus on exploring policy feasibility could refine and add complexity to
230 specific scenarios.

231 4.2 Further Work

232
233 One challenge of coupling qualitative assessments from backcasting to an MRIO framework is
234 that some envisioned lifestyles lie beyond the scope of Input-Output modelling. For example,
235 non-technical visions that encourage sharing economies, including downsizing of living space and
236 shared ownership might have significant potentials, but are better assessed through specific
237 surveys of household consumption or building types (Vásquez et al. 2016; Ivanova et al. 2018;
238 Vita et al. 2018; Daly 2017).

239 Future research on MRIO scenarios could be validated at finer geographical scales by better
240 representing the local context. In this paper, for example, we introduce physical data to model
241 energy and food to enhance the realism of EE-MRIO scenario modelling. Depending on the
242 research question, coupling to bottom-up physical data such as urban infrastructure, transport
243 fleet or household characteristics could be an asset (Ivanova et al. 2017, 2018).

244 A common limitation of economy-wide modelling in Industrial Ecology, whether Input-Output
245 or Material Stock Dynamics, is the lack of explicit consideration of in-use capital stocks, with
246 some remarkable efforts in this direction (Södersten et al. 2018; Wiedenhofer et al. 2019).
247 Construction scenarios could be enhanced by modelling in-use stocks. However, due to the long
248 lifetimes of buildings, construction materials typically represent a small share of the footprint
249 compared to yearly energy flows (Vásquez et al. 2016).

250 For household consumption, some “capital” goods are implicitly represented in MRIOs –e.g.,
251 housing is included in household demand through imputed rent. Similar, construction services,
252 office rental, machineries and other stock-like inputs are modelled as production inputs to other
253 industries, including service sectors.

254 Our EE-MRIO model represents a snapshot of the economy and disregards feedback dynamics
255 (Wood et al. 2017). In reality, we expect that scaling up alternative consumption patterns would
256 have non-linear effects due to social tipping points and learning curves (Nyborg et al. 2016a). The
257 advantage of the linear and static nature of our EE-MRIO model is that it eases the interpretation
258 of simulation results.

259 Although we focus on Europe, we expect the general direction of our results to be applicable to
260 other continents, with differences in the magnitudes and shares of foreign impacts. Still, repeating
261 the analysis for other regions and emerging economies is a topic for further research.

262 4.3 Adequacy of scenario parameters

263

264 The purpose of our assessments is not to forecast reductions but to characterize the ranges of
265 potentials and risks of materializing visions. To do so, we assume widespread adoption of
266 particular lifestyles. Nevertheless, in the SI we discuss the potential challenges of mainstreaming
267 sustainable lifestyles and compare the scenario parameters proposed by the stakeholders with
268 previous scientific literature

269 The peer-to-peer or sharing economy has been identified as a key feature of sustainable societies.
270 A recent study estimates above 70% reduction in energy intensities and yields economy-wide
271 energy reduction of 40%, due to sharing and collaborative economies as well as decentralization
272 of energy services by 2050 (Grubler et al. 2018). Here we assume that widespread sharing
273 economies, modelled in the *Non-Market Services* scenario, could reduce household demand of
274 market services by 80%. Such a reduction might seem ambitious given status-quo. However, a
275 large portion of European services represent non-basic needs, meaning that household
276 consumption of services is largely discretionary and their reduction would not drastically impact
277 quality of life (Jackson and Marks 1999; Druckman and Jackson 2010). Noteworthy that we do
278 not affect the demand of governments and non-profits serving households, which provide the
279 largest share of welfare services in Europe.

280 Most of the visions in this paper presume disruptive socio-technical changes. (Geels et al. 2017).
281 Historically, we have failed to predict the major technological and social breakthroughs of the
282 last 15 years (Rifkin 2015). However, a large share of renewables, the shared economy (transport
283 and housing), cryptocurrencies, repair cafés, cooperatives and even widespread adoption of
284 vegetarianism are increasingly enabling options for sustainable lifestyles. It is up to the wider
285 community, civil societies, firms and governments to decide and develop strategies to enhance
286 ambitious lifestyles changes.

287 4.4 Characterizing Uncertainty: The income rebound effect

288 Reducing or changing consumption can lead to savings, which consumers may spend on other
289 impactful goods, thus triggering a rebound effect which might undermine the environmental
290 benefits of lifestyles changes (Hertwich 2005a). In the SI, we repeat the scenario analysis
291 considering the potential income rebound effect by modelling savings as increased consumption,
292 according to current expenditure patterns (Wood et al. 2017). We report the rebound effect as a
293 uncertainty measure but acknowledge that voluntary lifestyle changes driven by environmental
294 values (and not economic incentives) are less subject to rebound (Hurst et al. 2013; Thøgersen
295 2013; Jackson 2005).

296 We find the largest potential rebounds for sufficiency scenarios since they entail the largest
297 savings. However, sufficiency is in line with a de-growth paradigm and which implies a steady
298 or downsized GDP, thus lowering the risk of rebound (Sekulova et al. 2017). Noteworthy, a full

299 analysis of the rebound effect would not only consider savings, but also changes in prices and
300 corresponding rules of purchasing behaviors.

301 From this uncertainty test, we conclude that policies to manage potential rebound effects are
302 recommendable. A traditional measure is to increase the prices or roll out taxes to hold energy-
303 service prices constant (Grubler et al. 2018). Such measures are more acceptable if the tax
304 addresses redistribution, social justice or a more fair access to resources, with the perk that
305 equality discourages positional consumption (Sekulova et al. 2017). More progressive measures
306 include planning saturation of service demand e.g., peak passenger-km travel, peak per capita
307 energy consumption or declining the number of emitted driver licenses (Grubler et al. 2018).

308 4.5 The challenges of green consumption 309

310 Although **sufficiency** options are generally more efficient and less risky, they are not as popular
311 as **green consumption** because of their conflict with prevailing economic growth paradigms
312 (Lorek and Spangenberg 2014; Akenji 2014; Vita 2016).

313 As expected, all sufficiency scenarios show unanimous reductions across footprints. On the other
314 hand, green consumption scenarios shift expenditure towards the goods that stakeholders
315 perceived as more “environmentally-friendly”, generally based on their lower-carbon emissions.
316 Nevertheless, while some green consumption scenarios yield reductions in carbon and toxicity,
317 these typically come at the potential risk of increasing land and water requirements. This occurs
318 specially when replacing carbon-intensive goods with land and water intensive renewable fuels,
319 materials and crops.

320

321 4.6 Lifestyle changes in the Shared Socioeconomic Pathways

322 The sufficiency and green consumption scenarios that we model here are compatible with the
323 most desirable scenario of the Shared Socioeconomic Pathways (SSP), the SSP1 “Sustainability –
324 Taking the Green Road”, which in turn is most compatible with mitigation and adaptation (Riahi
325 et al. 2017; O’Neill et al. 2017; Grubler et al. 2018). Its central feature is high environmental
326 awareness and moving towards less resource-intensive lifestyles, starting by high-income
327 countries (O’Neill et al. 2017). However, detailed lifestyle changes are not easily represented in
328 the SSP research because the demand sectors of Integrated Assessments Models (IAMs) are often

329 highly aggregated i.e., industry, energy and transportation (Riahi et al. 2017). We foresee
330 research opportunities in linking EE-MRIO with IAM-SSP research by adding heterogeneity and
331 allowing for more stylized scenarios (Rao et al. 2017; Pauliuk et al. 2017).

332 4.7 Displaced impacts and intra-generational solidarity

333 Greenhouse emissions contribute to global climate change regardless of their source location. On
334 the other hand, the negative health effects of toxicity emissions depend on the local context
335 (climate, pollution levels) and exposure to people (Johansson et al. 2017). Similarly, the
336 consequences of land-use and water are highly dependent on the local biodiversity, vegetation,
337 water availability and resource management practices (Haberl et al. 2007).

338 In terms of global justice, helping the world's poor meet their needs is an attitudinal pre-requisite
339 for sustainable lifestyles in wealthy countries (Schäpke and Rauschmayer 2014). At least half of
340 food and clothing impacts embodied in European consumption have consequences abroad (black
341 bars on Figure 2). Changes in European diets and fashion would relieve land and water resources
342 in producing countries, which are typically more climate vulnerable (Tukker et al. 2014).
343 However, reducing meat and clothing also benefits Europeans by reducing domestic carbon and
344 toxicity due to less processing, packaging and shipping. Sustainable housing mainly benefits
345 European impacts due to territorial electricity generation and local sourcing of fuels. Appliances
346 and electronics are largely produced outside Europe and thus reducing manufactured products
347 yields more benefits in foreign lands. International cooperation for sustainability could prioritize
348 the lifestyle changes that yield most bi-lateral benefits (Haberl et al. 2007; Keohane and Victor
349 2016).

350 4.8 Co-benefits and challenges of sustainable lifestyles

351 [Beyond footprint trade-offs, there are potential social trade-offs implied in the visions, discussed
352 at length in the SI.](#) Sufficiency measures could hinder economic growth and employment under
353 the current work-growth paradigm (D'Alisa et al. 2015). To prevent negative social effects, labor
354 and welfare institutions would require different practices to decouple wellbeing from paid
355 employment. Examples of new welfare practices include work-sharing or basic income schemes
356 (D'Alisa et al. 2015; Sekulova et al. 2017). Indeed, many of the backcasting visions went beyond
357 environmental concerns to include wellbeing aspects, such as working less, social connections,
358 being healthier or having more free time (Quist et al. 2016a, 2016b). Such aspects go beyond our
359 modelling scope but could be interesting leverage points for policymaking.

360 To complement the environmental analysis, in the SI we include a literature review of the
361 individual and societal benefits and challenges for quality of life associated with the modelled
362 lifestyle changes. For example, current European diets are characterized by an intake of animal
363 products above dietary recommendations for saturated fat and red meat (Westhoek et al. 2014).
364 Substitution of high saturated-fat, high-calorie meats, and processed foods with fibre rich foods,
365 fruits and vegetables has been linked to reduced risk of coronary heart disease (Dora et al. 2015).
366 Individuals with frequent walking or cycling habits show better mental and physical health than
367 their sedentary counterparts (Haines et al. 2009). At societal scale, lower environmental pollution
368 from renewable energy has proven benefits for public health (Gibon et al. 2017). Relying less on
369 market services and more on shared economy correlates with social empowerment and sense of
370 community (Frenken and Schor 2017).

371 **5) Conclusion**

372

373 The sustainability transformation requires not only innovative technologies but also innovative
374 lifestyles and engaged, well informed, citizens. In this study, we connect backcasting visions to
375 EE-MRIO to systematically assess scenarios of sustainable lifestyles and provide a scoreboard of
376 the options across consumption domains (GLAMURS et al. 2016; Quist et al. 2016a). We confirm
377 that some lifestyle changes envisioned by European citizens are promising options, with the
378 additional benefit that citizens demand such changes and that they are compatible with increased
379 quality of life. We also identify those options that are arguably fruitless or even risk backfire by
380 increasing other resources.

381 Except for switching to plant-based diets, the lifestyles with most potential generally imply
382 curbing consumption towards sufficiency levels. While we contrast sufficiency and green
383 consumption to show the independent contribution of each scenario, some scenarios are not
384 mutually exclusive and may be implemented synergistically to yield greater benefits. By studying
385 multiple environmental indicators we detect fewer trade-off risks and larger impact reduction
386 across footprints for sufficiency lifestyles, compared to green consumerism. Because European
387 lifestyles drive significant impact abroad, it is key to take responsibility by cooperating with
388 trading partners to deploy sustainable resource management, fair-trade and greener supply
389 chains.

390 This study provides an overview of the options for change and their consequences for the purpose
391 of comparison. Hence, our results are indicative of potential but not policy conclusive. In practice,
392 the outcome of the scenarios would largely depend on the implementation pathways. We rather
393 present a framework to integrate citizens' perspectives and imaginative alternatives into
394 sustainability scenarios to broaden the range of demand-side solutions.

395 Participatory modelling for sustainability can be seen as building human capital via social
396 learning or knowledge co-production (Bandura 2006). Its practice enriches scientific research, the
397 participants and, if taken to its ultimate consequences, the general public, by leading to policies
398 that truly consider the visions and needs of citizens. Understanding the global consequences of
399 local visions and actions is a pre-requisite to focus on the most promising options, and stir
400 governments, industries and communities towards them.

401

402 **Supplementary Information**

403

404 The Supplementary Information includes methodological details and data to model food and
405 energy scenarios through a physical layer. We discuss the relevant assumptions regarding the
406 adoption rates of scenarios. We present an uncertainty analysis assuming an income-rebound for
407 the scenarios that yield savings. We conduct a literature review on the co-benefits and challenges
408 for quality of life associated to the scenarios as well as critically discuss the adequacy of our
409 scalability parameters. The supplementary data file includes all the results on the environmental
410 assessments for each scenario. We include the full inventory of literal text extracts from the
411 backcasting workshops that were used to build scenarios, including the consumption implications
412 and modelling decisions.

413 **Acknowledgements**

414 This work is part of the GLAMURS project financed by the European Union's seventh
415 framework program (contract #613420). We thank all the researchers in GLAMURS who
416 organized, held and documented the backcasting workshops and those who contributed directly
417 to the text coding of reports for this paper: Vladimir. Pandur, Christin Polzin, Moritz. Petri, Ines
418 Thronicker, Malik Curuk, Helena Martínez, Alberto Díaz, Angelo Panno, Fridanna Maricchiolo,
419 Ambra Brizi, Ines Oman, Paul Lahner, Wouter Spekkink, and Eline Leising. We thank Ricardo

420 García-Mira and Adina Dumitru who coordinated the GLAMURS project. We thank Kam
421 Sripada and Angela McLean for her assistance proofreading this manuscript.

422 **References**

- 423
- 424 Ahmad, S., S. Pachauri, and F. Creutzig. 2017. Synergies and trade-offs between energy-efficient
425 urbanization and health. *Environmental Research Letters* 12(114017).
- 426 Akenji, L. 2014. Consumer scapegoatism and limits to green consumerism. *Journal of Cleaner Production*
427 63: 13–23. <http://dx.doi.org/10.1016/j.jclepro.2013.05.022>.
- 428 Akenji, L. and M. Bengtsson. 2014. Making Sustainable Consumption and Production the Core of
429 Sustainable Development Goals. *Sustainability* 6(2): 513–529. [http://www.mdpi.com/2071-](http://www.mdpi.com/2071-1050/6/2/513/)
430 [1050/6/2/513/](http://www.mdpi.com/2071-1050/6/2/513/). Accessed September 26, 2016.
- 431 Avetisyan, M., T. Hertel, and G. Sampson. 2014. Is Local Food More Environmentally Friendly? The GHG
432 Emissions Impacts of Consuming Imported versus Domestically Produced Food. *Environmental and*
433 *Resource Economics* 58(3): 415–462.
- 434 Bandura, A. 2006. Toward a Psychology of Human Agency. *Perspective on Psychological Science* 1(2):
435 164–180. <http://www.jstor.org/stable/40212163>.
- 436 Baumann, H. and G. Vita. 2015. Urban hunters and gatherers - an exploration into different varieties and
437 their relevance to industrial ecology. *ISIE 2015 Conference — Taking Stock of Industrial Ecology*.
438 [http://publications.lib.chalmers.se/publication/219636-urban-hunters-and-gatherers-an-](http://publications.lib.chalmers.se/publication/219636-urban-hunters-and-gatherers-an-exploration-into-different-varieties-and-their-relevance-to-industria)
439 [exploration-into-different-varieties-and-their-relevance-to-industria](http://publications.lib.chalmers.se/publication/219636-urban-hunters-and-gatherers-an-exploration-into-different-varieties-and-their-relevance-to-industria). Accessed May 15, 2017.
- 440 Bjørn, A., P. Kalbar, S.E. Nygaard, S. Kabins, C.L. Jensen, M. Birkved, J. Schmidt, and M.Z. Hauschild.
441 2018a. Pursuing necessary reductions in embedded GHG emissions of developed nations: Will
442 efficiency improvements and changes in consumption get us there? *Global Environmental Change*
443 52(September): 314–324.
- 444 Bjørn, A., P. Kalbar, S.E. Nygaard, S. Kabins, C.L. Jensen, M. Birkved, J. Schmidt, and M.Z. Hauschild.
445 2018b. Pursuing necessary reductions in embedded GHG emissions of developed nations: Will
446 efficiency improvements and changes in consumption get us there? *Global Environmental Change*
447 52(September): 314–324. <https://www.sciencedirect.com/science/article/pii/S0959378017304223>.
- 448 Brand-Correa, L.I., J. Martin-Ortega, and J.K. Steinberger. 2018. Human Scale Energy Services: Untangling
449 a “golden thread.” *Energy Research & Social Science* 38: 178–187.
450 <https://doi.org/10.1016/j.erss.2018.01.008>. Accessed July 5, 2018.
- 451 Brand-Correa, L.I. and J.K. Steinberger. 2017. A Framework for Decoupling Human Need Satisfaction
452 From Energy Use. *Ecological Economics* 141: 43–52.
453 <http://dx.doi.org/10.1016/j.ecolecon.2017.05.019>.
- 454 Carlsson-Kanyama, A., K.H. Dreborg, H.C. Moll, and D. Padovan. 2008. Participative backcasting: A tool
455 for involving stakeholders in local sustainability planning. *Futures* 40(1): 34–46.
- 456 CML-Leiden University. LCIA Characterisation Factors.
457 [https://www.universiteitleiden.nl/en/research/research-output/science/cml-ia-characterisation-](https://www.universiteitleiden.nl/en/research/research-output/science/cml-ia-characterisation-factors#downloads)
458 [factors#downloads](https://www.universiteitleiden.nl/en/research/research-output/science/cml-ia-characterisation-factors#downloads). Accessed March 19, 2018.

- 459 Cooper, M. 2013. The Intrinsic Foundations of Extrinsic Motivations and Goals. *Journal of Humanistic*
460 *Psychology* 53(2): 153–171. <http://journals.sagepub.com/doi/10.1177/0022167812453768>.
- 461 Creutzig, F., J. Roy, W.F. Lamb, I.M.L. Azevedo, W.B. de Bruin, H. Dalkmann, O.Y. Edelenbosch, et al.
462 2018. Towards demand-side solutions for mitigating climate change. *Nature Climate Change*.
- 463 D’Alisa, G., F. Demaria, G. Kallis, and S.K. Nelson. 2015. *Degrowth A vocabulary for a new era*. New York:
464 Routledge. [https://www.routledge.com/Degrowth-A-Vocabulary-for-a-New-Era/DAlisa-Demaria-](https://www.routledge.com/Degrowth-A-Vocabulary-for-a-New-Era/DAlisa-Demaria-Kallis/p/book/9781138000773)
465 [Kallis/p/book/9781138000773](https://www.routledge.com/Degrowth-A-Vocabulary-for-a-New-Era/DAlisa-Demaria-Kallis/p/book/9781138000773). Accessed March 18, 2018.
- 466 Daly, M. 2017. Quantifying the environmental impact of ecovillages and co-housing communities: a
467 systematic literature review. *Local Environment*. Routledge, November 2.
468 <https://www.tandfonline.com/doi/full/10.1080/13549839.2017.1348342>. Accessed July 6, 2018.
- 469 Debord, G. 1994. *The society of the spectacle*. Zone Books.
- 470 Dietz, T., G.T. Gardner, J. Gilligan, P.C. Stern, and M.P. Vandenberg. 2009. Household actions can
471 provide a behavioral wedge to rapidly reduce US carbon emissions. *Pnas* 106(4): 18452–18456.
472 <http://www.pnas.org/cgi/doi/10.1073/pnas.0908738106>.
- 473 Distelkamp, M. and M. Meyer. 2019. Pathways to a Resource-Efficient and Low-Carbon Europe.
474 *Ecological Economics* 155: 88–104.
- 475 Dora, C., A. Haines, J. Balbus, E. Fletcher, H. Adair-Rohani, G. Alabaster, R. Hossain, M. De Onis, F. Branca,
476 and M. Neira. 2015. Indicators linking health and sustainability in the post-2015 development
477 agenda. *The Lancet*. Elsevier Ltd, January.
478 <http://linkinghub.elsevier.com/retrieve/pii/S014067361460605X>.
- 479 Druckman, A. and T. Jackson. 2010. The bare necessities: How much household carbon do we really
480 need? *Ecological Economics* 69(9): 1794–1804.
- 481 Duarte, R., K. Feng, K. Hubacek, J. Sánchez-Chóliz, C. Sarasa, and L. Sun. 2016. Modeling the carbon
482 consequences of pro-environmental consumer behavior. *Applied Energy* 184: 1207–1216.
483 <http://linkinghub.elsevier.com/retrieve/pii/S0306261915012271>. Accessed February 15, 2016.
- 484 Dumitru, A., I. Anguelovski, F. Avelino, M. Bach, B. Best, C. Binder, J. Barnes, et al. 2016. Elucidating the
485 changing roles of civil society in urban sustainability transitions. *Current Opinion in Environmental*
486 *Sustainability* 22: 41–50. <https://www.sciencedirect.com/science/article/pii/S1877343517300659>.
487 Accessed April 8, 2018.
- 488 Dumitru, A., M. Ricardo García, and GLAMURS consortium. 2017. Final Report GLAMURS: Supporting
489 Green Lifestyles, Alternative Models and Upscaling Regional Sustainability. *European Commission*
490 *7th Framework Programme for Research and Technological Development. Funded under Socio-*
491 *Economic Sciences and Humanities*.
- 492 European Commission. 2014. A policy framework for climate and energy in the period from 2020 to 2030.
493 *European Parliament Communication*.
- 494 European Energy Agency. 2010. Household energy consumption for space heating per m² (2010, climate
495 corrected) — European Environment Agency. [https://www.eea.europa.eu/data-and-](https://www.eea.europa.eu/data-and-maps/figures/household-energy-consumption-for-space#tab-data-references)
496 [maps/figures/household-energy-consumption-for-space#tab-data-references](https://www.eea.europa.eu/data-and-maps/figures/household-energy-consumption-for-space#tab-data-references). Accessed July 8,
497 2017.
- 498 Eurostat. 2014. Modal split of inland passenger transport. *Statistics Explained*.

499 [http://ec.europa.eu/eurostat/statistics-](http://ec.europa.eu/eurostat/statistics-explained/index.php/File:Modal_split_of_inland_passenger_transport,_2014_(%25_of_total_inland_passenger-km)_YB17.png)
500 [explained/index.php/File:Modal_split_of_inland_passenger_transport,_2014_\(%25_of_total_inlan](http://ec.europa.eu/eurostat/statistics-explained/index.php/File:Modal_split_of_inland_passenger_transport,_2014_(%25_of_total_inland_passenger-km)_YB17.png)
501 [d_passenger-km\)_YB17.png](http://ec.europa.eu/eurostat/statistics-explained/index.php/File:Modal_split_of_inland_passenger_transport,_2014_(%25_of_total_inland_passenger-km)_YB17.png). Accessed November 29, 2017.

502 Eurostat. 2018. Housing statistics - Statistics Explained. *Eurostat- Housing Statistics*.
503 http://ec.europa.eu/eurostat/statistics-explained/index.php/Housing_statistics#Tenure_status.
504 Accessed November 15, 2017.

505 Frenken, K. and J. Schor. 2017. Putting the sharing economy into perspective. *Environmental Innovation*
506 *and Societal Transitions* 23: 3–10. [https://www.scopus.com/inward/record.uri?eid=2-s2.0-](https://www.scopus.com/inward/record.uri?eid=2-s2.0-85009887792&doi=10.1016%2Fj.eist.2017.01.003&partnerID=40&md5=24c4e6e577dea9ad2cf6d5c7f9f346a8)
507 [85009887792&doi=10.1016%2Fj.eist.2017.01.003&partnerID=40&md5=24c4e6e577dea9ad2cf6d5](https://www.scopus.com/inward/record.uri?eid=2-s2.0-85009887792&doi=10.1016%2Fj.eist.2017.01.003&partnerID=40&md5=24c4e6e577dea9ad2cf6d5c7f9f346a8)
508 [c7f9f346a8](https://www.scopus.com/inward/record.uri?eid=2-s2.0-85009887792&doi=10.1016%2Fj.eist.2017.01.003&partnerID=40&md5=24c4e6e577dea9ad2cf6d5c7f9f346a8).

509 Gardner, G.T. and P.C. Stern. 2008. The short list: The most effective actions U.S. households \tcan take
510 to curb climate change. *Environment: Science and Policy for Sustainable Development* 50(5): 12–25.

511 Geels, B.F.W., K. Benjamin, T. Schwanen, and S. Sorrell. 2017. Sociotechnical transitions for deep
512 decarbonization. *Science Policy Forum* 357(6357): 1242–1244.

513 Gibon, T., E.G. Hertwich, A. Arvesen, B. Singh, and F. Verones. 2017. Health benefits, ecological threats of
514 low-carbon electricity. *Environmental Research Letters* 12(3): 11.

515 Girod, B. and P. de Haan. 2010. More or better? A model for changes in household greenhouse gas
516 emissions due to higher income. *Journal of Industrial Ecology* 14(1): 31–49.

517 Girod, B., D.P. van Vuuren, and E.G. Hertwich. 2014. Climate policy through changing consumption
518 choices: Options and obstacles for reducing greenhouse gas emissions. *Global Environmental*
519 *Change* 25(1): 5–15. <http://www.sciencedirect.com/science/article/pii/S0959378014000077>.

520 GLAMURS, A. Dumitru, and R. et al. García Mira. 2016. Green Lifestyles, Alternative Models and
521 Upscaling Regional Sustainability(1): 1–56. www.glamurs.eu. Accessed March 12, 2018.

522 Grubler, A., C. Wilson, N. Bento, B. Boza-Kiss, V. Krey, D.L. McCollum, N.D. Rao, et al. 2018. A low energy
523 demand scenario for meeting the 1.5 °c target and sustainable development goals without negative
524 emission technologies. *Nature Energy* 3(6): 515–527. [http://dx.doi.org/10.1038/s41560-018-0172-](http://dx.doi.org/10.1038/s41560-018-0172-6)
525 [6](http://dx.doi.org/10.1038/s41560-018-0172-6).

526 Haberl, H., K.H. Erb, F. Krausmann, V. Gaube, A. Bondeau, C. Plutzer, S. Gingrich, W. Lucht, and M.
527 Fischer-Kowalski. 2007. Quantifying and mapping the human appropriation of net primary
528 production in earth’s terrestrial ecosystems. *Proceedings of the National Academy of Sciences of*
529 *the United States of America* 104(31): 12942–12947.

530 Haines, A., A.J. McMichael, K.R. Smith, I. Roberts, J. Woodcock, A. Markandya, B.G. Armstrong, et al.
531 2009. Public health benefits of strategies to reduce greenhouse-gas emissions: overview and
532 implications for policy makers. *The Lancet* 374(9707): 2104–2114.

533 Hertwich, E., N. Heeren, B. Kuczenski, G. Majeau-Bettez, R.J. Myers, S. Pauliuk, K. Stadler, and R. Lifset.
534 2018. Nullius in Verba1: Advancing Data Transparency in Industrial Ecology. *Journal of Industrial*
535 *Ecology* 22(1): 6–17. <http://doi.wiley.com/10.1111/jiec.12738>. Accessed October 31, 2018.

536 Hertwich, E. and M. Katzmayer. 2004. *Examples of Sustainable Consumption: Review, Classification and*
537 *Analysis. Industrial Ecology Programme NTNU Report 5/2004*.

538 Hertwich, E.G. 2005a. Consumption and the rebound effect: An industrial ecology perspective. *Journal of*

539 *Industrial Ecology* 9(1–2): 85–98. <http://dx.doi.org/doi:10.1162/1088198054084635>. Accessed
540 September 28, 2016.

541 Hertwich, E.G. 2005b. Consumption and Industrial Ecology. *Journal of Industrial Ecology* 9(1): 1–6.
542 <http://mitpress.mit.edu/jie>.

543 Hiç, C., P. Pradhan, D. Rybski, and J.P. Kropp. 2016. Food Surplus and Its Climate Burdens. *Environmental*
544 *Science and Technology* 50(8): 4269–4277.

545 Hoolohan, C., M. Berners-Lee, J. McKinstry-West, and C.N. Hewitt. 2013. Mitigating the greenhouse gas
546 emissions embodied in food through realistic consumer choices. *Energy Policy* 63: 1065–1074.

547 Huppmann, D., J. Rogelj, E. Kriegler, V. Krey, and K. Riahi. 2018. A new scenario resource for integrated
548 1.5 °C research. *Nature Climate Change*: 1–4. <http://www.nature.com/articles/s41558-018-0317-4>.

549 Hurst, M., H. Dittmar, R. Bond, and T. Kasser. 2013. The relationship between materialistic values and
550 environmental attitudes and behaviors: A meta-analysis. *Journal of Environmental Psychology* 36:
551 257–269. <http://dx.doi.org/10.1016/j.jenvp.2013.09.003>.

552 Illich, I. 1971. *Deschooling society*. Calder & Boyars.

553 Ivanova, D., K. Stadler, K. Steen-Olsen, R. Wood, G. Vita, A. Tukker, and E.G. Hertwich. 2016.
554 Environmental Impact Assessment of Household Consumption. *Journal of Industrial Ecology* 00(0):
555 1–11.

556 Ivanova, D., G. Vita, K. Steen-Olsen, K. Stadler, P.C.P.C. Melo, R. Wood, and E.G. Hertwich. 2017. Mapping
557 the carbon footprint of EU regions. *Environmental Research Letters* 12(5): 054013.
558 <http://iopscience.iop.org/article/10.1088/1748-9326/aa6da9>. Accessed May 2, 2017.

559 Ivanova, D., G. Vita, R. Wood, C. Lausset, A. Dumitru, K. Krause, I. Macsinga, and E.G. Hertwich. 2018.
560 Carbon mitigation in domains of high consumer lock-in. *Global Environmental Change* 52(February):
561 117–130.

562 Jackson, T. 2005. Live better by consuming less? Is there a “double dividend” in sustainable
563 consumption? *Journal of Industrial Ecology* 9(1–2): 19–36.
564 <http://dx.doi.org/doi:10.1162/1088198054084734>.

565 Jackson, T. 2011. *Prosperity without growth: economics for a finite planet*. *International Journal of*
566 *Ambient Energy*. Vol. 32. London: Routledge.
567 <http://www.tandfonline.com/doi/abs/10.1080/01430750.2011.615179>.

568 Jackson, T. and N. Marks. 1999. Consumption, sustainable welfare and human needs - With reference to
569 UK expenditure patterns between 1954 and 1994. *Ecological Economics* 28(3): 421–441.

570 Johansson, L., J.-P.P. Jalkanen, and J. Kukkonen. 2017. Global assessment of shipping emissions in 2015
571 on a high spatial and temporal resolution. *Atmospheric Environment* 167: 403–415.
572 <https://linkinghub.elsevier.com/retrieve/pii/S1352231017305563>. Accessed October 3, 2018.

573 Jordan, R., S. Gray, M. Zellner, P.D. Glynn, A. Voinov, B. Hedelin, E.J. Sterling, et al. 2018. 12 Questions for
574 the participatory modeling community. *Earth’s Future*: 1–12.

575 Katajajuuri, J.M., K. Silvennoinen, H. Hartikainen, L. Heikkilä, and A. Reinikainen. 2014. Food waste in the
576 Finnish food chain. *Journal of Cleaner Production* 73: 322–329.

- 577 Kemp-Benedict, E. 2004. From narrative to number: a role for quantitative models in scenario analysis.
578 *IEMSs 2004 International Congress: "Complexity and Integrated Resources Management."*
- 579 Keohane, R.O. and D.G. Victor. 2016. Cooperation and discord in global climate policy. *Nature Climate*
580 *Change* 6(6): 570–575. <http://www.nature.com/articles/nclimate2937>. Accessed March 15, 2018.
- 581 Lorek, S. and J.H. Spangenberg. 2014. Sustainable consumption within a sustainable economy - Beyond
582 green growth and green economies. *Journal of Cleaner Production* 63: 33–44.
583 <http://dx.doi.org/10.1016/j.jclepro.2013.08.045>.
- 584 Min, J. and N.D. Rao. 2017. Estimating Uncertainty in Household Energy Footprints. *Journal of Industrial*
585 *Ecology* 00(0): 1–11.
- 586 Moran, D. and R. Wood. 2014. Convergence Between the Eora, Wiod, Exiobase, and Openeu'S
587 Consumption-Based Carbon Accounts. *Economic Systems Research* 26(3): 245–261.
588 <http://www.tandfonline.com/doi/abs/10.1080/09535314.2014.935298>.
- 589 Moran, D., R. Wood, E. Hertwich, K. Mattson, J.F.D. Rodriguez, K. Schanes, and J. Barrett. 2018.
590 Quantifying the potential for consumer-oriented policy to reduce European and foreign carbon
591 emissions. *Climate Policy* 0(0): 1–11.
592 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1551186>.
- 593 Mosenthal, P. and M. Socks. 2015. *Potential for Energy Savings in Affordable Multifamily Housing*. Final
594 Report. Energy Efficiency For All Project. Optimal Energy, May.
- 595 Nyborg, K., J.M. Anderies, A. Dannenberg, T. Lindahl, C. Schill, M. Schlüter, W.N. Adger, et al. 2016a.
596 Social norms as solutions. *Science Magazine* 354(6308): 42–43.
- 597 Nyborg, K., J.M. Anderies, A. Dannenberg, T. Lindahl, C. Schill, M. Schlüter, W.N. Adger, et al. 2016b.
598 Social norms as solutions: Policies may influence large-scale behavioral tipping. *Science* 354: 42–43.
- 599 O'Brien, K. 2015. Political agency: The key to tackling climate change. *Science* 350(6265): 4–6.
- 600 O'Brien, M., F. Hartwig, K. Schanes, M. Kammerlander, I. Omann, H. Wilts, R. Bleischwitz, and J. Jäger.
601 2014. Living within the safe operating space: a vision for a resource efficient Europe. *European*
602 *Journal of Futures Research* 2(1): 48. [https://link.springer.com/content/pdf/10.1007%2Fs40309-](https://link.springer.com/content/pdf/10.1007%2Fs40309-014-0048-3.pdf)
603 [014-0048-3.pdf](https://link.springer.com/content/pdf/10.1007%2Fs40309-014-0048-3.pdf). Accessed October 3, 2018.
- 604 O'Neill, B.C., E. Kriegler, K.L. Ebi, E. Kemp-Benedict, K. Riahi, D.S. Rothman, B.J. van Ruijven, et al. 2017.
605 The roads ahead: Narratives for shared socioeconomic pathways describing world futures in the
606 21st century. *Global Environmental Change* 42: 169–180.
607 <http://dx.doi.org/10.1016/j.gloenvcha.2015.01.004>.
- 608 O'Neill, D.W., A.L. Fanning, W.F. Lamb, and J.K. Steinberger. 2018. A good life for all within planetary
609 boundaries. *Nature Sustainability* 1(February): 88–95. [http://dx.doi.org/10.1038/s41893-018-0021-](http://dx.doi.org/10.1038/s41893-018-0021-4)
610 [4](http://dx.doi.org/10.1038/s41893-018-0021-4). Accessed March 12, 2018.
- 611 Omann, I., M. Mock, C. Polzin, and F. Rauschmayer. 2016. *Deliverable 5.1: Report on sustainable lifestyle*
612 *initiatives in 7 case studies*. <http://glamurs.eu/downloads/deliverables/>.
- 613 Palmer, M.A., J. Liu, J.H. Matthews, M. Mumba, and P. D'Odorico. 2015. Manage water in a green way.
614 *Science* 349(6248): 584–585. <http://www.sciencemag.org/cgi/doi/10.1126/science.aac7778>.
- 615 Pauliuk, S., A. Arvesen, K. Stadler, and E.G. Hertwich. 2017. Industrial ecology in integrated assessment

- 616 models. *Nature Climate Change* 7(1): 13–20.
617 <http://www.nature.com/doi/10.1038/nclimate3148>.
- 618 Pauliuk, S., G. Majeau-Bettez, C.L. Mutel, B. Steubing, and K. Stadler. 2015. Lifting Industrial Ecology
619 Modeling to a New Level of Quality and Transparency: A Call for More Transparent Publications and
620 a Collaborative Open Source Software Framework. *Journal of Industrial Ecology* 19(6): 937–949.
621 <http://doi.wiley.com/10.1111/jiec.12316>. Accessed October 31, 2018.
- 622 PWC. 2015. *The Sharing Economy*. Delaware.
- 623 Quist, J., E. Leising, A. Blöbaum, A. Brizi, G. Carrus, A. Díaz-Ayude, A. Dumitru, et al. 2016a. *Deliverable*
624 *4.3: Report on future lifestyle scenarios and backcasting vision workshops. EU FP7 SSH Call:*
625 *2013.2.1-1- Obstacles and Prospects for Sustainable Lifestyles and Green Economy.*
626 <http://glamurs.eu/downloads/deliverables/>.
- 627 Quist, J., E. Leising, A. Brizi, G. Carrus, A. Dumitru, R.G. Mira, K. Krause, et al. 2016b. *Deliverable 5.2:*
628 *Report on future lifestyle pathways and workshops EU.* <http://glamurs.eu/downloads/deliverables/>.
- 629 Quist, J. and P. Vergragt. 2006. Past and future of backcasting: The shift to stakeholder participation and
630 a proposal for a methodological framework. *Futures* 38: 1027–1045.
- 631 Rao, N.D., J. Min, R. Defries, S. Ghosh-jerath, H. Valin, and J. Fanzo. 2018. Healthy, affordable and
632 climate-friendly diets in India. *Global Environmental Change* 49(March 2017): 154–165.
633 <https://doi.org/10.1016/j.gloenvcha.2018.02.013>.
- 634 Rao, N.D., B.J. Van Ruijven, K. Riahi, and V. Bosetti. 2017. Improving poverty and inequality modelling in
635 climate research. *Nature Climate Change* 7(12): 857–862. [http://dx.doi.org/10.1038/s41558-017-](http://dx.doi.org/10.1038/s41558-017-0004-x)
636 [0004-x](http://dx.doi.org/10.1038/s41558-017-0004-x).
- 637 Reganold, J.P. and J.M. Wachter. 2016. Organic agriculture in the twenty-first century. *Nature Plants*
638 2(February): 15221. <http://dx.doi.org/10.1038/nplants.2015.221>.
- 639 Riahi, K., D.P. van Vuuren, E. Kriegler, J. Edmonds, B.C. O'Neill, S. Fujimori, N. Bauer, et al. 2017. The
640 Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions
641 implications: An overview. *Global Environmental Change* 42: 153–168.
- 642 Rifkin, J. 2015. *The Zero Marginal Cost Society : the internet of things, the collaborative commons, and*
643 *the eclipse of capitalism*. Palgrave Macmillan.
- 644 Robinson, J.B. 1990. Futures under glass. A recipe for people who hate to predict. *Futures* 22(8): 820–
645 842.
- 646 Rodrigues, J.F.D., D. Moran, R. Wood, and P. Behrens. 2018. Uncertainty of Consumption-Based Carbon
647 Accounts. *Environmental Science and Technology* 52(13): 7577–7586.
- 648 Rogelj, J., A. Popp, K. V Calvin, G. Luderer, J. Emmerling, D. Gernaat, S. Fujimori, et al. 2018. Scenarios
649 towards limiting global mean temperature increase below 1.5 °C. *Nature Climate Change* 8(4): 325–
650 332. <https://www.nature.com/articles/s41558-018-0091-3.pdf>. Accessed March 26, 2018.
- 651 Schanes, K., S. Giljum, and E. Hertwich. 2016. Low carbon lifestyles: A framework to structure
652 consumption strategies and options to reduce carbon footprints. *Journal of Cleaner Production*
653 139(September): 1033–1043. <http://dx.doi.org/10.1016/j.jclepro.2016.08.154>.
- 654 Schanes, K., J. Jäger, and P. Drummond. 2019. Three Scenario Narratives for a Resource-Efficient and

- 655 Low-Carbon Europe in 2050. *Ecological Economics* 155(February 2017): 70–79.
656 <https://doi.org/10.1016/j.ecolecon.2018.02.009>.
- 657 Schöpke, N. and F. Rauschmayer. 2014. Going beyond efficiency: including altruistic motives in
658 behavioral models for sustainability transitions to address sufficiency. *Sustainability: Science,*
659 *Practice, and Policy* 10(1): 29–44.
- 660 Sekulova, F., G. Kallis, and F. Scheider. 2017. Climate change , happiness and income from a degrowth
661 perspective. In *Handbook on Growth and Sustainability*, ed. by Peter A. Victor and Brett Dolter,
662 160–180. Glos & Massachusetts: Edward Elgar Publishing Limited.
- 663 Södersten, C.-J., R. Wood, and E.G. Hertwich. 2018. Endogenizing capital in MRIO models: the
664 implications for consumption-based accounting. *Environ. Sci. Technol., Just Accepted Manuscript* •.
665 <http://pubs.acs.org>. Accessed September 27, 2018.
- 666 Steen-Olsen, K. and E.G. Hertwich. 2015. Life cycle assessment as a means to identify the most effective
667 action for sustainable consumption. In *Handbook of Research on Sustainable Consumption*, ed. by
668 Lucia Reisch and J. Thøgersen, 131–144. first. Glos & Massachusetts: Edward Elgar Publishing
669 Limited.
- 670 Steinberger, J.K. and J.T. Roberts. 2010. From constraint to sufficiency: The decoupling of energy and
671 carbon from human needs, 1975–2005. *Ecological Economics* 70(2): 425–433.
672 <http://linkinghub.elsevier.com/retrieve/pii/S0921800910003733>. Accessed October 23, 2014.
- 673 Stiglitz, J.E., A. Sen, and J.-P. Fitoussi. 2010. *Mismeasuring Our Lives: Why GDP Doesn't Add Up*. Vol. 1.
674 New York: New Press.
- 675 Thøgersen, J. 2005. How may consumer policy empower consumers for sustainable lifestyles? *Journal of*
676 *Consumer Policy* 28(2): 143–177.
- 677 Thøgersen, J. 2013. Psychology: Inducing green behaviour. *Nature Climate Change* 3(2): 100–101.
- 678 Tukker, A., T. Bulavskaya, S. Giljum, A. de Koning, S. Lutter, M. Simas, K. Stadler, and R. Wood. 2014. *The*
679 *Global Resource Footprint of Nations: Carbon, water, land and materials embodied in trade and*
680 *final consumption calculated with EXIOBASE 2.1. Carbon, Water, Land and Materials Embodied in*
681 *Trade and Final Consumption Calculated with EXIOBASE*. Vol. 2.
682 [http://www.researchgate.net/profile/Stefan_Giljum/publication/264080789_The_Global_Resource](http://www.researchgate.net/profile/Stefan_Giljum/publication/264080789_The_Global_Resource_Footprint_of_Nations_Carbon_water_land_and_materials_embodied_in_trade_and_final_consumption/links/02e7e53cd0969e6723000000.pdf)
683 [_Footprint_of_Nations_Carbon_water_land_and_materials_embodied_in_trade_and_final_consu](http://www.researchgate.net/profile/Stefan_Giljum/publication/264080789_The_Global_Resource_Footprint_of_Nations_Carbon_water_land_and_materials_embodied_in_trade_and_final_consumption/links/02e7e53cd0969e6723000000.pdf)
684 [mption/links/02e7e53cd0969e6723000000.pdf](http://www.researchgate.net/profile/Stefan_Giljum/publication/264080789_The_Global_Resource_Footprint_of_Nations_Carbon_water_land_and_materials_embodied_in_trade_and_final_consumption/links/02e7e53cd0969e6723000000.pdf).
- 685 Tukker, A., T. Bulavskaya, S. Giljum, A. de Koning, S. Lutter, M. Simas, K. Stadler, and R. Wood. 2016.
686 Environmental and resource footprints in a global context: Europe's structural deficit in resource
687 endowments. *Global Environmental Change* 40: 171–181.
688 <https://www.sciencedirect.com/science/article/pii/S0959378016301091>. Accessed March 12,
689 2018.
- 690 Tukker, A., M.J. Cohen, K. Hubacek, and O. Mont. 2010. The impacts of household consumption and
691 options for change. *Journal of Industrial Ecology* 14(1): 13–30.
692 <http://doi.wiley.com/10.1111/j.1530-9290.2009.00208.x>. Accessed March 12, 2018.
- 693 Ürge-Vorsatz, D., C. Rosenzweig, R.J. Dawson, R. Sanchez Rodriguez, X. Bai, A.S. Barau, K.C. Seto, and S.
694 Dhakal. 2018. Locking in positive climate responses in cities Adaptation–mitigation
695 interdependencies. *Nature Climate Change*: 1. <https://www.nature.com/articles/s41558-018-0100->

696 6.pdf. Accessed March 12, 2018.

697 Usubiaga, A., I. Butnar, and P. Schepelmann. 2018. Wasting Food, Wasting Resources: Potential
698 Environmental Savings Through Food Waste Reductions. *Journal of Industrial Ecology* 22(3): 574–
699 584.

700 Vanham, D., F. Bouraoui, A. Leip, B. Grizzetti, and G. Bidoglio. 2015. Lost water and nitrogen resources
701 due to EU consumer food waste. *Environmental Research Letters* 10(8): 084008.
702 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/10/i=8/a=084008?key=crossref.dc6f64f92c31f0190936f0f4219a3404)
703 [9326/10/i=8/a=084008?key=crossref.dc6f64f92c31f0190936f0f4219a3404](http://stacks.iop.org/1748-9326/10/i=8/a=084008?key=crossref.dc6f64f92c31f0190936f0f4219a3404).

704 Vásquez, F., A.N. Løvik, N.H. Sandberg, and D.B. Müller. 2016. Dynamic type-cohort-time approach for
705 the analysis of energy reductions strategies in the building stock. *Energy and Buildings* 111: 37–55.
706 <http://www.sciencedirect.com/science/article/pii/S0378778815303832>.

707 Vásquez, F., G. Vita, and D. Müller. 2018a. Food Security for an Aging and Heavier Population.
708 *Sustainability* 10(10): 3683. <http://www.mdpi.com/2071-1050/10/10/3683>.

709 Vásquez, F., G. Vita, and D. Müller. 2018b. Food Security for an Aging and Heavier Population.
710 *Sustainability* 10(10): 3683.

711 Vergragt, P.J. and J. Quist. 2011. Backcasting for sustainability: Introduction to the special issue.
712 *Technological Forecasting and Social Change* 78(5): 747–755.
713 <http://dx.doi.org/10.1016/j.techfore.2011.03.010>.

714 Vieux, F., N. Darmon, D. Touazi, and L.G. Soler. 2012. Greenhouse gas emissions of self-selected
715 individual diets in France: Changing the diet structure or consuming less? *Ecological Economics* 75:
716 91–101.

717 Vita, G. 2016. Smart City or Ecovillage? An Industrial Ecology approach. *Fox & Hedgehog: The Current*
718 *Global Affairs Review*. [http://www.foxhedgehog.com/2016/09/smart-city-or-ecovillage-an-](http://www.foxhedgehog.com/2016/09/smart-city-or-ecovillage-an-industrial-ecology-approach/)
719 [industrial-ecology-approach/](http://www.foxhedgehog.com/2016/09/smart-city-or-ecovillage-an-industrial-ecology-approach/).

720 Vita, G., E.G. Hertwich, K. Stadler, and R. Wood. 2019. Connecting global emissions to fundamental
721 human needs and their satisfaction. *Environmental Research Letters* 14(1): 014002.
722 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/14/i=1/a=014002?key=crossref.7ebdb6d35d4aeace64743d94e5915a3c)
723 [9326/14/i=1/a=014002?key=crossref.7ebdb6d35d4aeace64743d94e5915a3c](http://stacks.iop.org/1748-9326/14/i=1/a=014002?key=crossref.7ebdb6d35d4aeace64743d94e5915a3c). Accessed February
724 14, 2019.

725 Vita, G., D. Ivanova, A. Dumitru, R. García-mira, G. Carrus, K. Stadler, K. Krause, R. Wood, and E.G.
726 Hertwich. 2018. Members of environmental grassroots initiatives reconcile lower carbon emissions
727 with higher well-being. *Environmental Research Letters*(Forthcoming).

728 Vuuren, D.P. van, K. Riahi, K. Calvin, R. Dellink, J. Emmerling, S. Fujimori, S. KC, E. Kriegler, and B. O'Neill.
729 2017a. The Shared Socio-economic Pathways: Trajectories for human development and global
730 environmental change. *Global Environmental Change* 42: 148–152.

731 Vuuren, D.P. van, K. Riahi, K. Calvin, R. Dellink, J. Emmerling, S. Fujimori, S. KC, E. Kriegler, and B. O'Neill.
732 2017b. The Shared Socio-economic Pathways: Trajectories for human development and global
733 environmental change. *Global Environmental Change*.

734 Westhoek, H., J.P. Lesschen, T. Rood, S. Wagner, A. De Marco, D. Murphy-Bokern, A. Leip, H. van
735 Grinsven, M.A. Sutton, and O. Oenema. 2014. Food choices, health and environment: Effects of

- 736 cutting Europe's meat and dairy intake. *Global Environmental Change* 26: 196–205.
- 737 Wiebe, K.S. 2016. The impact of renewable energy diffusion on European consumption-based emissions.
738 *Economic Systems Research* 28(2): 133–150.
739 <http://www.tandfonline.com/doi/full/10.1080/09535314.2015.1113936>. Accessed May 29, 2018.
- 740 Wiedenhofer, D., T. Fishman, C. Lauk, W. Haas, and F. Krausmann. 2019. Integrating Material Stock
741 Dynamics Into Economy-Wide Material Flow Accounting: Concepts, Modelling, and Global
742 Application for 1900–2050. *Ecological Economics* 156(September 2018): 121–133.
743 <https://doi.org/10.1016/j.ecolecon.2018.09.010>.
- 744 Wiedenhofer, D., B. Smetschka, L. Akenji, M. Jalas, and H. Haberl. 2018. Household time use, carbon
745 footprints, and urban form: a review of the potential contributions of everyday living to the 1.5 °C
746 climate target. *Current Opinion in Environmental Sustainability* 30: 7–17.
747 <http://linkinghub.elsevier.com/retrieve/pii/S1877343517301318>. Accessed March 16, 2018.
- 748 Wood, R., D. Moran, K. Stadler, D. Ivanova, K. Steen-Olsen, A. Tisserant, and E.G. Hertwich. 2017.
749 Prioritizing Consumption-Based Carbon Policy Based on the Evaluation of Mitigation Potential Using
750 Input-Output Methods. *Journal of Industrial Ecology* 0(3): 540–552.
751 <http://doi.wiley.com/10.1111/jiec.12702>. Accessed June 13, 2018.
- 752 Wood, R., K. Stadler, T. Bulavskaya, S. Lutter, S. Giljum, A. de Koning, J. Kuenen, et al. 2015. Global
753 Sustainability Accounting—Developing EXIOBASE for Multi-Regional Footprint Analysis.
754 *Sustainability* 7(1): 138–163. <http://www.mdpi.com/2071-1050/7/1/138/>.
- 755 Wynes, S. and K.A. Nicholas. 2017. The climate mitigation gap: education and government
756 recommendations miss the most effective individual actions. *Environmental Research Letters* 12(7):
757 074024. [http://stacks.iop.org/1748-
758 9326/12/i=7/a=074024?key=crossref.03823b1b77b6f51ed344568b22e48bad](http://stacks.iop.org/1748-9326/12/i=7/a=074024?key=crossref.03823b1b77b6f51ed344568b22e48bad).
- 759