

1 **Modelling of environmental flow options for optimal Atlantic salmon (*Salmo salar*)**
2 **embryo survival during hydropeaking**

3

4 **Abstract**

5 Recent findings on causes of Atlantic salmon embryo mortality during winter in a
6 hydropeaked river suggest that long duration drawdowns during very cold periods are the
7 most likely cause of mortality in the ramping zone areas. This paper presents a framework in
8 which thresholds for optimal embryo survival at the micro-scale are linked to physical habitat
9 requirements at the meso-scale and integrated into alternative hydropower operations at the
10 catchment scale. The connections within this framework are executed with a one-dimensional
11 hydraulic model at the meso-scale and a hydropower simulation program at the catchment
12 scale. The economic costs and feasibility of several alternative options for hydropeaking
13 operation that would comply with ecological requirements for the optimal survival of
14 embryos were evaluated. This paper presents a method to assess a wide range of alternative
15 hydropower options that consider key factors to mitigate the conflicting process requirements
16 of ecological targets, technical feasibility, and economics. Targeted alternative environmental
17 flow releases to meet specific ecological objectives are often more effective than general
18 operational rules to comply with legislation. The development of well-informed and targeted
19 mitigation strategies is important for future environmental hydropower management.

20

21 **Key words:** Atlantic salmon; embryo survival, hydropeaking management, modelling tools,
22 upscaling

23

24

25

26

27

28 **Introduction**

29 Hydropeaking in regulated rivers is predicted to be more frequent in the near future given the
30 increasing demand for renewable energy in Europe. Norway, through its hydropower systems,
31 has >50% of the total European water storage potential with possibilities for increase
32 (Catrinu-Renström and Knudsen, 2011), providing opportunities for balancing the energy load
33 with other renewables such as wind power. Load balancing using hydropower implies more
34 frequent fluctuations of water levels in rivers due to hydropeaking operations.

35 Atlantic salmon (*Salmo salar*) spawns typically in the autumn by depositing their eggs in
36 redds in the river bed gravels (de Gaudemar *et al.*, 2000, Mills, 1989). The most restrictive
37 niche of all Atlantic salmon life stages is egg and embryo development, which occurs during
38 winter (Cunjak *et al.*, 1998, Cunjak and Therrien, 1998). Although survival rates of embryo
39 and alevins in natural conditions can be very high (Elliott, 1984) mainly due to the protection
40 offered in the gravels, eggs have no capacity to move from malign abiotic factors. This makes
41 their survival dependent on subsurface and surface water exchange (Schmidt and Hahn,
42 2012), hyporheic water quality, water delivery rate, temperature and gravel composition and
43 the complex interaction between these (Gibbins *et al.*, 2008, Malcolm *et al.*, 2008, Malcolm *et*
44 *al.*, 2003).

45 River regulation may reduce the amount and quality of suitable spawning habitat, which is a
46 prime necessity for recruitment and population sustainability. In particular, the dewatering of
47 salmonid redds is of great concern for water resource management in regulated rivers
48 (Malcolm *et al.*, 2012). If spawning occurs during high flows, nest sites may be dewatered as
49 a result of hydropeaking operations (McMichael *et al.*, 2005; Vollset *et al.*, unpublished data)
50 or when water levels are not maintained (Bauersfeld, 1978; Skoglund *et al.*, 2012). Eggs of
51 fall-spawners can freeze and die in cold areas during low flow periods in late winter under
52 natural conditions, in regulated rivers with large annual variations, or in rivers subject to
53 hydropeaking if flow is reduced after spawning (Skoglund *et al.*, 2012). The extent of egg
54 mortality due to freezing in regulated rivers can be influenced by subsurface water inputs in
55 the catchment (Saltveit and Brabrand, 2013), but the exposure to dry and frost conditions will
56 remain a main driver for the survival of Atlantic salmon eggs in rivers with long drawdowns
57 due to hydropeaking regulations (Casas-Mulet *et al.*, submitted).

58 Harby *et al.* (2001) introduces some general advice on best-practices for environmental
59 management of Norwegian power plants with hydropeaking operations, but more emphasis is
60 needed to develop targeted mitigation strategies on an individual-case basis, particularly for
61 salmonid populations. Some examples of research-based targeted mitigation measures for
62 salmonids are presented in the literature for regulated rivers (Fjeldstad *et al.*, 2014, Gibbins
63 and Acornley, 2000). However, for the environmental management of hydropeaking further
64 research is needed to fully understand the issues on an individual-case basis in order to
65 develop targeted mitigation strategies at the catchment scale.

66 With detailed investigations of small-scale physical and biological processes, and a good
67 understanding of the links with processes operating at larger scales, it is possible to reflect
68 how management decisions at larger scales can affect such small-scale processes. In order to
69 transfer the information from small to large scales, spatial upscaling methods are needed, to
70 overcome the validity problems that micro-scale habitat analysis might carry (Borsányi,
71 2005).

72 Hydropower operational strategies at the catchment scale are a major influence on what
73 occurs at the meso and micro-scales. Therefore, ecological and physical processes occurring
74 at the smaller scales should be considered when making managerial decisions at the
75 catchment scale. Studies at the micro-scale enable an accurate and detailed representation of
76 ecological processes. The connections between hydraulic processes, geomorphology and river
77 ecology have been studied in recent years (Maddock, 1999, Padmore, 1998, Petts *et al.*,
78 2006). Several studies have proven relevant links between physical habitat and ecology at the
79 meso-scale (Kemp *et al.*, 1999, Moir and Pasternack, 2008, Padmore, 1997, Parasiewicz,
80 2007), suggesting the meso-scale to be an adequate dimension to study relevant ecological
81 processes and a feasible scale for river management decisions (Newson and Newson, 2000).

82 The challenges rely on integrating different scales as a first step to link research findings with
83 hydropower operation alternatives at the catchment scale for hydropeaking environmental
84 impact assessment. A decision support tool that is able to evaluate ecological needs and to
85 calculate power production in different peaking scenarios is needed for large-scale
86 management decisions. By linking the findings at the micro-scale with meso and macro-scale
87 processes, the resulting system should be a tool to assess the cost of production while also
88 fulfilling ecological requirements. In order to establish real links between findings at the
89 micro-scale to the hydropower simulation level, connections have to be done through

90 modelling tools. The use of predictive modeling tools and their validation will be important to
91 understand processes and to assess future changes in the regulation of the hydro systems, and
92 they will be key to define when, where and how hydropower plants can use peaking
93 operations without causing ecological damage.

94 This paper presents a framework in which research on Atlantic salmon egg survival at the
95 small scale and links to physical habitat at the meso-scale are integrated to assess alternative
96 environmental flow options to reduce egg mortality. The connections in this framework are
97 executed by combining hydraulic and hydropower operation models with egg survival data for
98 a spawning site. This paper presents a tool to assess the integration of hydropower operation
99 at the catchment scale to mitigate and/or avoid potential impacts on small-scale fish
100 recruitment processes in the river.

101 The main objective is to assess the possibilities of alternative hydropower operations for the
102 significant improvement of egg survival potential using data from the Lundesokna river as a
103 test case. Several scenarios will be assessed according to duration, volume of extra water
104 release, and their economic cost. In addition, the strategy presented could be seen as a general
105 methodology for assessing how changes in hydropower operation might influence in-stream
106 processes.

107

108 **Materials and methods**

109 *Study site*

110 The Lundesokna hydropower system in central Norway (Figure 1) covers a total catchment
111 area of 395 km² with an average annual runoff of 381 Mm³y⁻¹. It comprises the whole
112 Lundesokna catchment (264 km², 41.2 km river length and 247 Mm³ average annual runoff)
113 and part of the Burusjøen (16.8 km²) and Holta (114.4 km²) catchments, which are connected
114 via three main transfers. The Lundesokna hydropower system consists of three reservoirs
115 (Håen, Samsjøen and Holtsjøen) with a combined 145 Mm³ in water volume, and three power
116 plants (Sokna, Håen and Sama) with a total installed capacity of 61 MW and average annual
117 production of 278 GWh. Sokna power plant operates according to market price fluctuations
118 occurring daily and weekly and to available water in reservoirs. This results in the lower 2.5

119 km of the Lundesokna river being subject to regular and abrupt flow fluctuations with a
120 typical flow range varying from $0.3 \text{ m}^3\text{s}^{-1}$ (no power production) to $20 \text{ m}^3\text{s}^{-1}$ (full production).

121 Sokna power plant has an optimal intake capacity of $20 \text{ m}^3\text{s}^{-1}$ and the minimum production
122 flow is $8 \text{ m}^3\text{s}^{-1}$, below of which the system might experience poor efficiency and cavitation
123 problems (Viggo Finset, pers.comm.). There are no compulsory minimum flow requirements
124 in the Lundesokna river, but a voluntary release of $0.3 \text{ m}^3\text{s}^{-1}$ through a manually operated gate
125 occurs between May and September when needed for the operation of the hatchery in the
126 catchment. The study focuses on Lundesokna below the Sokna power plant outlet.

127 *Durations of critical periods*

128 Environmental requirements for the optimal survival of Atlantic salmon embryos were
129 established from the findings in, where survival rates were lower in the ramping zone (76%)
130 compared to the permanently wet areas (99%). Exposure to dry conditions due to production
131 stops combined with air temperatures below $0 \text{ }^\circ\text{C}$ was the main factor explaining egg
132 mortality. The maximum duration exposure to dry and freezing conditions with no egg
133 mortality was about three hours. Highest survival rates ($>86\%$) showed maximum durations
134 of production stops with air temperatures below zero up to six hours. Some of Casas-Mulet, *et*
135 *al.* (submitted) findings recommended to minimize the duration of hydropower production
136 stops during extremely cold air temperatures in hydropeaking rivers.

137 In order to translate such recommendations into practical management measures, both the
138 three and six hour maximum cessation of hydropower production in combination with air
139 temperatures below zero were considered as potential duration thresholds to establish
140 alternative management options for hydropower production that would be compatible with
141 optimal embryo survival, using the Lundesokna river as an example.

142 *Minimum flow thresholds*

143 The upper boundary of potentially suitable spawning areas in the Lundesokna river was
144 estimated at 30.6 m through field observations in a representative transect in the study area in
145 Casas-Mulet *et al.* (submitted). Such elevation was used as a threshold to establish minimum
146 flows needs, assuming that water at this elevation would cover 100% of the potential
147 spawning grounds (based on field observation on substrate suitability) in the ramping zone.

148 In order to translate the obtained water elevations to discharge in the river, hydraulic
149 simulations had been carried out for the whole Lundesokna river length with the aid of the
150 one-dimensional hydraulic simulation model HEC-RAS (US Army Corps of Engineers,
151 2012). More details on the field data collection, model set-up and calibration are described in
152 (Casas-Mulet *et al.*, 2014). The resulting minimum discharge at which 100% of the eggs were
153 covered by water was $3.5 \text{ m}^3\text{s}^{-1}$. Minimum production flow of $8 \text{ m}^3\text{s}^{-1}$ (minimum turbine capacity)
154 was also modelled in HEC-RAS and considered as the minimum flow threshold for some of
155 the options for alternative hydropeaking management. In addition, the increase of wet area
156 due to such additional flow increases (3.5 and $8 \text{ m}^3\text{s}^{-1}$) was calculated at the transect level.

157 *Options for alternative hydropeaking management*

158 Based on the above information, several options for additional water release were established
159 according to: (i) type of additional water release (bypass or production); (ii) intervals of
160 additional water release (permanently, every 3 hours or every 6 hours) and (iii) duration of the
161 release (non-stop, 2 hours or 1 hour). Such alternative hydropower management options were
162 to be implemented during critical periods with low flows (mainly due to production stops)
163 coinciding with air temperatures below $0 \text{ }^\circ\text{C}$. They ranged from a permanent minimum bypass
164 release of $3.5 \text{ m}^3\text{s}^{-1}$ (options 1) or the minimum possible production flow of $8 \text{ m}^3\text{s}^{-1}$ (options
165 2) to an alternated or flexible minimum bypass release or production conditional to air
166 temperature (options 3 and 4). All options are illustrated in Figure 2 and summarized in Table
167 1.

168 Critical periods were first identified from available data for the hydrological years (1
169 September to 31 August) 2002-2003 to 2012-2013. Hourly air temperature was combined
170 with hourly production and spill data from Sokna power plant to identify the critical periods
171 with flows below $3.5 \text{ m}^3\text{s}^{-1}$. Hourly air temperature data (2002-2013) was obtained from the
172 nearby Voll climatic station and it was used to identify periods with air temperature below 0
173 $^\circ\text{C}$. A correction factor to estimate the actual temperature in the field site during the studied
174 period (2002-2013) was applied to the data from Voll. The correction factor was obtained by
175 comparison of air temperature data obtained from the field site during the period December
176 2011 to May 2013 with data from Voll. All data was obtained from Norwegian Water
177 Resources and Energy Directorate (NVE) and the Norwegian Meteorological Institute.

178

179 *Simulation of additional water volume used*

180 The additional water required for additional bypass or production release in each of the
181 options was calculated by adding up the hourly additional release for each of the periods
182 (2002-2003 to 2012-2013).

183 An assessment on whether the hydropower system was able to handle the additional volume
184 of water needed for each of the options was carried out by using the nMAG hydropower
185 simulation program (Killingveit & Sælthun, 1995). A model of the river Lundesokna
186 hydropower production system was established. It involved a total of 11 modules consisting
187 of the described reservoirs, power plants, interbasin transfers and control points (Figure 3).
188 The simulation used a reservoir guide curve for all reservoirs assuming they start at 70% of
189 the volume on 1 January, emptied on 1 May, filled up again to 90% on 1 July and kept at that
190 level until 1 October, then gradually decreased back to 70% on 31 December. Model inputs
191 consisted of existing runoff data from the nearby Hugdal catchment, with similar physical
192 characteristics to Lundesokna. Runoff data were obtained from the period 1986 to 2013.
193 Calibration was based on historical production data and reservoir water levels for the same
194 periods. All data were obtained from The Norwegian Water Resources and Energy
195 Directorate (NVE) and TrønderEnergi.

196 The capacity of the hydropower system to provide the additional water needed for each of the
197 designed options was assessed by comparing it to the nMAG simulated maximum volume of
198 water available in the reservoirs for each of the simulated years.

199 *Economics*

200 An economic assessment was carried out for each of the options by calculating the revenue of
201 additional production, the costs of additional water release and the costs of additional starts in
202 Sokna power plant. The revenue of additional production was obtained by adding up the
203 additional hourly production (MWh) multiplied by the real hourly energy price (euro per
204 MWh) for each of the options. This only applied to options 2 and 4. Energy price data were
205 obtained from TrønderEnergi in norwegian crowns (NOK) per MWh that was converted into
206 euro per MWh according to daily exchange rates from The National Bank of Norway. A
207 monetary cost due to additional water used was assumed on the basis that such extra use of
208 water could have been kept and used at times with a higher hourly price. This assumption was

209 applied to all the options irrespective of bypass release or production. Therefore, the final
210 annual production or bypass release flow (m^3s^{-1}) used was added up, converted to production
211 (MWh) and multiplied by a high price assumed to be a 0.99 percentile of the annual price
212 range for the periods 2002-2003 to 2012-2013. The monetary cost of the additional turbine
213 starts due to additional production (only options 2 and 4) in Sokna power plant was
214 considered for each of the periods. The cost of starting the hydropower plant is at present 200
215 euros. Inflation values were obtained for each of the periods and applied to the present cost to
216 obtain the actual cost for each of the simulated periods. The final economic balance for each
217 of the designed options was compared to the actual annual revenue made for each of the
218 periods.

219 A final qualitative and quantitative assessment was made for each of the options regarding (i)
220 the likelihood to meet optimal survival of eggs, (ii) the feasibility of the hydropower system
221 to carry out the additional production or bypass releases in terms of water usage,
222 economically and technically, and (iii) the additional usage of the power plant. The likelihood
223 of meeting the optimal survival for eggs was assessed as *very likely* when potential spawning
224 areas were wet during air temperatures below zero with 100% of certainty or with safe
225 intervals of 3 hours, and *likely* when potential spawning areas were wet during air
226 temperatures below zero but with <100% of certainty or with intervals of 6 hours. Feasibility
227 in terms of water usage was ranked according to the average additional percentage of water
228 used in relation to the maximum volume of water in the system. Economic feasibility was
229 assessed relative to the average percentage of monetary loss in comparison to the annual
230 revenue. Technical feasibility was assessed as not feasible and feasible according to the
231 present technical characteristics of the system. The additional usage of the power plant was
232 ranked according to the average percentage of additional hours of production in relation to
233 average actual production.

234

235 **Results**

236 *Linking scales and models: The linking methodology*

237 Findings from Casas-Mulet *et al.* (submitted) helped establish certain bottleneck periods for
238 the survival of Atlantic salmon embryo in a study site in the Lundesokna river. A certain

239 minimum water elevation (30.6 m) should be maintained during periods with air temperature
240 below 0 °C to avoid mortality in potential spawning areas. Links between the micro and
241 meso-scales were made by translating water elevations into minimum flows ($3.5 \text{ m}^3\text{s}^{-1}$)
242 through a HEC-RAS simulation. Such minimum discharge and critical conditions were then
243 translated into several alternative hydropower operation options encompassing bypass release
244 or production flow to be carried out at certain periods and time intervals. The method of
245 linking scales is illustrated in Figure 4 as part of the assessment process explained below.

246 *Modelling calibration and outputs*

247 The nMAG simulation with the current operational strategy resulted in a total annual
248 production of 305.1 GWh, showing a difference of <0.001 % in comparison to the actual
249 production of the three power plants for the period 1986-2009. Average reservoir water levels
250 differed between 0.01 and -0.41% when compared to the actual reservoir data for the studied
251 period (Table 2).

252 The calibrated HEC-RAS model for the Lundesokna river provided that the increase of wet
253 area from the minimum recorded flow ($0.3 \text{ m}^3\text{s}^{-1}$) to the minimum bypass release ($3.5 \text{ m}^3\text{s}^{-1}$)
254 was 55%, while from the minimum recorded flow to the production flow ($8 \text{ m}^3\text{s}^{-1}$) the
255 increase was 190% at the transect level.

256 *Water use and system availability assessment*

257 The additional volume of water needed for each of the options and for each of the simulated
258 years is illustrated in Figure 5. Scenarios with permanent or partially permanent production or
259 bypass release (options 1 and 2) required volumes of additional water ranging from 77.5 to
260 15.1 million of m^3 per year. Option 4.1 (permanent production during periods with air
261 temperature below zero) also had a high requirement for additional water (20 mill. m^3). Those
262 were very high in comparison to those required in options 3 and 4 (ranging from 0.7 to 7 mill.
263 m^3).

264 The percentage of additional volume of water needed in comparison to the maximum water
265 volume in the system (including Håen, Samsjøen and Holtsjøen reservoirs) for each of the
266 simulated years is illustrated in Figure 6 for each of the bypass and production options.
267 Option 2.1 would require between 90% and 27% of the maximum water volume in the
268 system, showing great variability between individual years. Options with targeted releases

269 through bypass and production (3.2.1 to 3.3.2 and 4.2.1 to 4.3.2) would require much less
270 percentage of the total volume, varying between 9% and 0.02% of such volume, showing as
271 well some variability between individual years.

272 Figure 7 illustrates the percentage of additional hours of production for each of the simulated
273 options in comparison to the actual hours of annual production. As expected, options 2.1 and
274 2.2, with a permanent or partially permanent production, were the options with a higher
275 proportion of additional hours to be released (44.4% and 21.4% on average respectively).
276 Options 4.2.1 to 4.3.2, with an alternative bypass or production of 1 hour every 6 hours, were
277 the options with the lowest additional production hours, ranging from 10% to 1% additional
278 production hours (see also Table 3).

279 *Economics*

280 Revenue from additional production and costs from additional starts in the hydropower plant
281 did not have any effect on the economic balance made for the bypass releases options (Figure
282 8). The cost of bypass release was therefore related only to the potential economic loss due to
283 volume of water used, with the higher costs found in options 1, for the permanent bypass
284 release and the lower cost found in options 3, with the least hours of required release (3.3.2).
285 Alternative production options accounted for some revenue due to the additional production.
286 It was significant in comparison to the additional cost of extra starts in the power plant, but
287 could not compensate for the loss of volume of water used that could have been used in
288 periods with higher market prices. The final balance therefore was also an economic loss,
289 which was greater in options 2 and was reduced progressively in options 4 as less volume of
290 water was required. The annual balance for each of the alternative options is illustrated in
291 Figure 9, showing the high variability in economic loss from year to year within the same
292 option.

293 Table 3 summarizes the above results in ecological, economic and technical terms for the
294 assessment of each of the options and Figure 4 illustrates a simplified decision tree including
295 all terms of assessment and the inclusion of the linking method.

296

297 **Discussion**

298 The present study takes a multi-scale approach, from both an ecological and physical habitat
299 point of view to a hydropower operation application. In order to establish connections
300 between the ecological and physical processes occurring at the micro-scale and the potential
301 hydropower operation alternatives at the macro-scale, connections were established through
302 the use of existing modelling tools. The connection between scales was done by linking
303 detailed findings on ecological processes at the micro-scale with the 1D hydraulic model
304 (HEC-RAS) at the meso-scale, representing physical habitat changes, and later this was linked
305 to the nMAG hydropower operation model at the macro-scale by assessing the additional
306 volume of water required. Ecologically relevant findings at the micro-scale can be translated
307 to the meso-scale with the use of HEC-RAS (water elevation and wet area changes) and that
308 in turn provides input to the hydropower simulator nMAG. Results show this methodology is
309 a valid way to establish links between scales and models in order to assess potential
310 alternative hydropower operations for the management of hydropeaking rivers.

311 The use of modelling tools and their integration for the prediction of alternative hydropeaking
312 management to meet environmental requirements was also illustrated in Borsányi *et al.*
313 (2001), where links between existing programs including a one-dimensional hydraulic model
314 and nMAG were assessed in detail for the quantification of habitat use during normal and
315 habitat friendly hydropeaking strategies. Fjeldstad *et al.* (2014), for example used nMAG
316 together with smolt models for the analysis of smolt migration in order to explore possible
317 mitigation scenarios for increased smolt survival. The framework presented in the present
318 paper can be used for any environmentally relevant findings at the micro or meso-scale and
319 potential alternative hydropower operations at the macro-scale. Prediction of future changes
320 in hydropower management can be done and help define when, where and how hydropower
321 plants can use peaking operations from an environmental point of view.

322 The total water volume available in the Lundesokna hydropower system (from the three
323 reservoirs) would barely allow carrying some of the alternative hydropower management
324 options in wet years. The amount of water required in options 1 and 2 and 4.1 would make the
325 hydropower operation unfeasible both in terms of water availability and economics. Therefore
326 only the conditional bypass or production release in options 3 and 4 could be considered.

327 Technically, the mechanism used to release the additional volume of water needed (bypass or
328 production) is an important consideration. Sokna power plant at present does not have any
329 established system to automatically control bypass release, and $8 \text{ m}^3\text{s}^{-1}$ is the minimum flow

330 that the power plant is capable of producing. Therefore, currently options including
331 production (options 2 and 4) would be the most feasible to undertake from a technical point of
332 view.

333 When taking into account the amount of extra hours of production, options 2 resulted in a
334 very high additional use of the power plant in comparison to options 4. Options 1 and 3
335 including bypass release, would not impose any additional production hours. This is an
336 important factor to consider from the increased use of the turbine that might reduce its life
337 span and impose significant additional costs in the long term.

338 From an ecological point of view, an increase in wet area is likely to provide more available
339 habitat for salmonids in early stages and can be used as a measure of reduced risk of
340 mortality. Specifically for the purpose of this study, an increase of 55% in wet area as a result
341 of a bypass release of $3.5 \text{ m}^3\text{s}^{-1}$ (options 1 and 3) would be considered sufficient to ensure egg
342 survival in the areas where spawning is more likely to occur. Equal option alternatives but
343 with production flow to $8 \text{ m}^3\text{s}^{-1}$ would not provide any additional benefit for the specific
344 objective of salmon embryo survival optimization. Regardless their non-feasibility at the
345 hydropower system level, options 1.1 and 2.1 (producing or bypassing extra water
346 permanently) would ensure that the potential spawning areas are wet at all times, but in non-
347 critical periods they would most likely not add any additional benefit for the optimal embryo
348 survival. A targeted bypass release or production flow during specific periods with air
349 temperatures below zero would be sufficient to allow optimal embryo survival. Such options
350 would probably even be more efficient than the partial permanent release for the period 1
351 October to 31 March suggested in options 1.2 and 2.2. In some of the considered years, air
352 temperatures coinciding with production stops were found to occur outside of such
353 established period.

354 Economically, revenue lost in the final assessment is very similar between comparable bypass
355 release and production options, suggesting that although production might increase some
356 revenue, this is mainly revenue resulting from selling at low prices, which is very low in
357 comparison of the potential loss of water volume that could had been sold at higher prices.
358 This also indicates how the variability of the power prices is a main driver in deciding when
359 and how much to produce and how it influences the calculation of the costs in the present
360 study.

361

362 Overall, this paper illustrates that in certain cases, with research-based knowledge, it can be
363 more cost-effective to employ a targeted change in production patterns aiming at a specific
364 ecological objective than to establish a permanent release to meet general legislative rules,
365 such as the establishment of constant or periodically constant minimum flows. Integrating
366 knowledge of ecology and environmental processes with operational planning allows for
367 further design of environmental flow regimes that fit the production schedule of the
368 hydropower plant and meet ecological requirements, particularly in existing power plants.

369 The development of well-informed and targeted mitigation strategies is important for future
370 environmental hydropower management. The presented method considers key factors when
371 evaluating different hydropeaking scenarios in order to develop a technically feasible
372 operation strategy that optimizes both economic and ecological performance. It provides a
373 tool for setting flows for the particular case of egg survival; however the focus of this study is
374 to provide a general methodology that is also applicable to other components of the aquatic
375 ecosystem, and ultimately a tool that can be used in the decision-making process for the
376 establishment of true environmental flow regimes rather than minimum flow regimes
377 currently employed in many regulated rivers.

378

379 **References**

380 Bauersfeld K. (1978) *Stranding of juvenile salmon by flow reductions at Mayfield Dam on the*
381 *Cowlitz River, 1976*. Washington State Department of Fisheries, Olympia. 36 pp.

382 Borsányi P. (2005) *A classification method for scaling river biotopes for assessing*
383 *hydropower regulation impacts*. PhD Thesis, Norwegian University of Science and
384 Technology, 255 pp.

385 Borsányi P., Killingveit Å. & Alfredsen K. (2001) A Decision Support System for
386 Hydropower Peaking Operation. *Hydropower in the New Millennium: Proceedings of the 4th*
387 *International Conference on Hydropower Development*. 20-22 June 2001. Bergen, Norway,
388 pp. 191-196.

389 Casas-Mulet R., Saltveit S.J., Alfredsen K. Submitted manuscript. The survival of Atlantic
390 salmon (*Salmo salar*) eggs during dewatering in a river subject to hydropeaking.

391 Casas-Mulet R., Alfredsen K., Boissy T., Sundt H. & Rütther N. (2014) Performance of a on-
392 dimensional hydraulic model for the calculation of stranding areas in hydropeaking rivers.
393 *River Research and Applications*. Early view.

- 394 Catrinu-Renström M.D. & Knudsen J.K. (2011) *Perspectives on hydropower's role to*
 395 *balance non-regulated renewable power production in Northern Europe*. Sintef Report,
 396 Trondheim.
- 397 Cunjak R.A., Prowse T.D. & Parrish D.L. (1998) Atlantic salmon (*Salmo salar*) in winter:
 398 "the season of parr discontent"? *Canadian Journal of Fisheries and Aquatic Sciences* **55**
 399 (Suppl.1), 161-180.
- 400 Cunjak R.A. & Therrien J. (1998) Inter-stage survival of wild juvenile Atlantic salmon, *Salmo*
 401 *salar* L. *Fisheries Management and Ecology* **5** (3), 209-223.
- 402 De Gaudemar B., Schroder S. & Beall E. (2000) Nest Placement and Egg Distribution in
 403 Atlantic Salmon Redds. *Environmental Biology of Fishes* **57** (1), 37-47.
- 404 Elliott J.M. (1984) Numerical Changes and Population Regulation in Young Migratory Trout
 405 *Salmo trutta* in a Lake District Stream, 1966-83. *Journal of Animal Ecology* **53** (1), 327-350.
- 406 Fjeldstad H.P., Alfredsen K. & Boissy T. (2014) Optimising Atlantic salmon smolt survival
 407 by use of hydropower simulation modelling in a regulated river. *Fisheries Management and*
 408 *Ecology* **21** (1), 22-31.
- 409 Gibbins C., Shellberg J., Moir H. & Soulsby C. (2008) Hydrological influences on adult
 410 salmonid migration, spawning, and embryo survival. In: Sear, D. & P. Devries (eds) *Salmon*
 411 *Spawning Habitat in Rivers: Physical Controls, Biological Responses and Approaches to*
 412 *Remediation*, American Fisheries Society.
- 413 Gibbins C.N. & Acornley R.M. (2000) Salmonid habitat modelling studies and their
 414 contribution to the development of an ecologically acceptable release policy for Kielder
 415 Reservoir, North-east England. *Regulated Rivers: Research & Management* **16** (3), 203-224.
- 416 Harby A., Alfredsen K., Fjeldstad H.P., Halleraker J.H., Arnekleiv J.V., Borsányi P.,
 417 Flodmark L.E.W., Saltveit S.J., Johansen S.W., Vehanen T., Huusko A., Clarke K. & Scruton
 418 D.A. (2001) Ecological Impacts of hydro peaking in rivers. *Proceedings of the 4th*
 419 *International Conference on Hydropower Development*. Lisse, Netherlands.
- 420 Kemp J.L., Harper D.M. & Crosa G.A. (1999) Use of 'functional habitats' to link ecology
 421 with morphology and hydrology in river rehabilitation. *Aquatic Conservation: Marine and*
 422 *Freshwater Ecosystems* **9** (1), 159-178.
- 423 Killingveit Å. & Sælthun N.R. (1995) Hydrology. *Volume 7 in the Hydropower Development*
 424 *series*. Norwegian Institute of Technology, Trondheim, 213 pp.
- 425 Maddock I. (1999) The importance of physical habitat assessment for evaluating river health.
 426 *Freshwater Biology* **41** (2), 373-391.
- 427 Malcolm I.A., Gibbins C.N., Soulsby C., Tetzlaff D. & Moir H.J. (2012) The influence of
 428 hydrology and hydraulics on salmonids between spawning and emergence: implications for
 429 the management of flows in regulated rivers. *Fisheries Management and Ecology* **19** (6), 464-
 430 474.

- 431 Malcolm I.A., Youngson A.F., Greig S. & Soulsby C. (2008) Hyporheic influences on
 432 spawning success. In: Sear, D. & Devries, P. (eds) *Salmon Spawning Habitat in Rivers:
 433 Physical Controls, Biological Responses and Approaches to Remediation*, American Fisheries
 434 Society.
- 435 Malcolm I.A., Youngson A.F. & Soulsby C. (2003) Survival of salmonid eggs in a degraded
 436 gravel-bed stream: effects of groundwater–surface water interactions. *River Research and
 437 Applications* **19** (4), 303-316.
- 438 McMichael G. A., Rakowski C. L., James B. B., & Lukas J. A. (2005) Estimated Fall
 439 Chinook Salmon Survival to Emergence in Dewatered Redds in a Shallow Side Channel of
 440 the Columbia River. *North American Journal of Fisheries Management* **25** (3), 876-884.
- 441 Mills D.H. (1989) *Ecology and Management of Atlantic Salmon*. London: Chapman & Hall.
- 442 Moir H.J. & Pasternack G.B. (2008) Relationships between mesoscale morphological units,
 443 stream hydraulics and Chinook salmon (*Oncorhynchus tshawytscha*) spawning habitat on the
 444 Lower Yuba River, California. *Geomorphology* **100** (3–4), 527-548.
- 445 Newson M.D. & Newson C.L. (2000) Geomorphology, ecology and river channel habitat:
 446 mesoscale approaches to basin-scale challenges. *Progress in Physical Geography* **24** (2), 195-
 447 217.
- 448 Padmore C.L. (1997) Biotopes and their hydraulics: a method for defining the physical
 449 component of freshwater quality. In: Boon, P.J. & Howell, D.L. (eds) *Freshwater Quality:
 450 Defining the Indefinable*. Edinburgh, Scottish Natural Heritage.
- 451 Padmore C.L. (1998) The role of physical biotopes in determining the conservation status and
 452 flow requirements of British rivers. *Aquatic Ecosystem Health & Management* **1** (1), 25-35.
- 453 Parasiewicz P. (2007) The MesoHABSIM model revisited. *River Research and Applications*
 454 **23** (8), 893-903.
- 455 Petts G., Morales Y. & Sadler J. (2006) Linking hydrology and biology to assess the water
 456 needs of river ecosystems. *Hydrological Processes* **20** (10), 2247-2251.
- 457 Saltveit S.J. & Brabrand Å. (2013) Incubation, hatching and survival of eggs of Atlantic
 458 salmon (*Salmo salar*) in spawning redds influenced by groundwater. *Limnologica - Ecology
 459 and Management of Inland Waters* **43** (5), 325-331.
- 460 Schmidt S.I. & Hahn H.J. (2012) What is groundwater and what does this mean to fauna? –
 461 An opinion. *Limnologica - Ecology and Management of Inland Waters* **42** (1), 1-6.
- 462 Skoglund H., Barlaup B., Gabrielsen S., Lehmann G., Halvorsen G., Wiers T., Skår B., Pulg
 463 U., & Vollset K. (2012) Fiskebiologiske undersøkelser i Eidfjordvassdraget – sluttrapport for
 464 perioden 2004-2012. uniMiljø Report. 108 pp
- 465 US Army Corps of Engineers (2012) HEC-RAS. 4.1 ed.

466 Vollset K., Barlaup B., Skoglund H., Gabrielsen S. & Wiers T. Unpublished data. Effects of
467 hydropeaking on the spawning behaviour of Atlantic salmon (*Salmo salar*) and brown trout
468 (*Salmo trutta*).