

1 **No evidence that seed predators constrain pollinator-mediated trait**  
2 **evolution in a tropical vine**

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16 Running head: Weak effects of seed predators on floral evolution

17 **Abstract**

18 **Premise of The Study:** Turnover in biotic communities across heterogeneous landscapes is  
19 expected to lead to variation in interactions among plants, their mutualists, and their  
20 antagonists. Across a fragmented landscape in northern Costa Rica, populations of the  
21 euphorb vine *Dalechampia scandens* vary widely in mating systems and associated blossom  
22 traits. Previous work suggests that populations are well adapted to the local reliability of  
23 pollination by apid and megachilid bees. Here, we test whether variation in the intensity of  
24 predispersal seed predation by *Nanobaris* seed weevils also contributes to the observed  
25 variation in blossom traits.

26 **Methods:** We studied spatio-temporal variation in the relationships between floral  
27 advertisement and the probability of seed predation within three focal populations. Then, we  
28 assessed among-population covariation of predation rate, pollination reliability, mating  
29 system, and blossom traits across 20 populations.

30 **Key Results:** The probability of seed predation was largely unrelated to variation in floral  
31 advertisement both within focal populations, and among the larger sample of populations. The  
32 rate of seed predation was only weakly associated with the rate of cross-pollination  
33 (allogamy) in each population but tended to be proportionally greater in populations  
34 experiencing less reliable pollination.

35 **Conclusions:** These results suggest that geographic variation in the intensity of antagonistic  
36 interactions have had only minor modifying effects on the evolutionary trajectories of floral  
37 advertisement in plant populations in this system. Thus, pollinator-driven floral trait evolution  
38 in *Dalechampia scandens* in the study area appears not to be influenced by conflicting seed-  
39 predator-mediated selection.

40 **Key words:** conflicting selection; *Dalechampia*; Euphorbiaceae; herkogamy; interaction  
41 turnover; phenotypic selection; plant mating systems; seed predation

## 42 INTRODUCTION

43 Changes in biotic assemblages and disruption of species interactions are important biological  
44 consequences of global climate change, habitat destruction, and other anthropogenic  
45 disruptions of the environment (Magurran, 2016; Urban et al., 2016). Plant species occurring  
46 across fragmented or otherwise heterogeneous landscapes often experience variation in  
47 communities of competitors, mutualists, and antagonists, and provide excellent opportunities  
48 for understanding plant responses to turnover in interactor communities. For example, the  
49 ongoing decline of pollinators is expected to reduce the reliability of pollination in plant  
50 populations worldwide (Aguilar et al., 2006; Eckert et al., 2010; Potts et al., 2010; Thomann  
51 et al., 2013). The most commonly observed plant evolutionary response to pollinator declines  
52 is the evolution of greater autonomous selfing rates as a mechanism of reproductive assurance  
53 (Moeller, 2006; Eckert et al., 2010; Brys and Jacquemyn, 2012; Opedal et al., 2016a; but see  
54 Koski et al., 2017). However, the evolution of floral traits and mating systems in response to  
55 changing pollinator communities may not necessarily occur independently from other biotic  
56 interactions, such as herbivory and seed predation. Indeed, many studies have demonstrated  
57 antagonist-mediated selection on floral traits, and that this selection can sometimes run  
58 counter to pollinator-mediated selection (e.g. Strauss and Armbruster, 1997; Gómez and  
59 Zamora, 2000; Adler and Bronstein, 2004; Cariveau et al., 2004; Rey et al., 2006; Strauss and  
60 Whittall, 2006; Gómez et al., 2009; Kolb and Ehrlen, 2010; Pérez-Barrales et al., 2013; Sun et  
61 al., 2016). Therefore, the outcome of plant adaptation to changes in the abundance of one  
62 interacting species may be modified by conflicting selective pressures generated by other  
63 interactors.

64         The net strength and direction of selection on floral traits is expected to depend on the  
65 relative intensities of mutualistic and antagonistic interactions (Benkman, 2013;  
66 Vanhoenacker et al., 2013). All else being equal, the opportunity for mutualist-mediated  
67 selection is expected to decrease with increasing interaction intensity, while the opportunity  
68 for antagonist-mediated selection is expected to increase with increasing interaction intensity.  
69 Therefore, much of the observed variation in selection acting on plant phenotypes may result  
70 from spatial and temporal variation in the intensities of species interactions (Thompson,  
71 2005). Spatial turnover in species interactions may be particularly common in  
72 anthropogenically disturbed landscapes, because pollinators and other interactors often  
73 respond differently to habitat destruction and fragmentation (Cunningham, 2000; Steffan-  
74 Dewenter et al., 2001; Garcia and Chacoff, 2007; Magrach et al., 2014; Brudvig et al., 2015).

75 For example, hawthorn trees in northern Spain occurring in more fragmented habitats  
76 experienced less reliable pollination by bees and flies, and less frugivory by birds, but more  
77 intense seed predation by mice (Garcia and Chacoff, 2007). Similarly, both the bee pollinators  
78 and lepidopteran and dipteran seed predators of *Centaurea jacea* in Germany were less  
79 abundant in experimental populations located at sites containing less semi-natural habitat, but  
80 the effect of landscape structure on interactions were species-specific and depended on the  
81 spatial scale analyzed (Steffan-Dewenter et al., 2001). We may therefore expect both the  
82 absolute and relative intensities of mutualistic and antagonistic interactions to vary across  
83 heterogeneous landscapes, but whether this leads to systematic differences in selection  
84 remains an empirical question.

85         The neotropical vine *Dalechampia scandens* L. (Euphorbiaceae) is pollinated by  
86 female apid and megachilid bees (Armbruster and Webster, 1982; Armbruster, 1985). A  
87 recent study in Costa Rica showed that the floral traits and mating systems of populations  
88 varied predictably along a gradient of pollination reliability (Opedal et al., 2016a):  
89 populations at pollinator-poor sites had evolved blossom traits associated with greater rates of  
90 autonomous selfing, including smaller involucral bracts (a floral-advertisement trait;  
91 Armbruster et al., 2005) and reduced herkogamy. While these results suggest that population-  
92 mean floral trait values have tracked variation in the reliability of pollination, pollinators may  
93 not be the only biotic interactors affecting the evolution of *D. scandens* blossom traits. This  
94 was demonstrated by a study in a Mexican population, where the net selection on blossom  
95 traits was determined by interactions with both pollinators and predispersal seed predators  
96 (Pérez-Barrales et al., 2013). While pollinators selected for larger floral bracts, seed predators  
97 apparently responded to the same cues, generating conflicting selection. If seed predators  
98 consistently select for reduced floral advertisement, seed-predator-mediated selection might  
99 shift trait values away from the mating-system-related optimum determined by the local  
100 reliability of pollination. Specifically, if the strength of selection increases with the intensity  
101 of predation (Vanhoenacker et al., 2013), more intensively predated populations would be  
102 expected to have smaller floral bracts than expected from their historical pollination  
103 environment.

104         To assess whether predispersal seed predators modify pollinator-mediated floral trait  
105 evolution in *D. scandens*, we first studied spatio-temporal variation in seed-predator-mediated  
106 selection, i.e. the relationships between floral advertisement and the probability of seed  
107 predation within populations. Second, we assessed the joint influence of mating system and

108 seed predation rate on among-population divergence in floral advertisement. Specifically, we  
109 asked (1) whether seed predators exert selection on floral advertisement, and (2) whether  
110 selection varies in time and space depending on the rate of seed predation. Using data from a  
111 larger number of populations, we tested (3) whether predation rates covary with pollination  
112 reliability among populations, and (4) whether the rate of attack by seed predators predicts  
113 population-mean floral trait values after controlling for the effect of mating systems.

## 114 **MATERIALS AND METHODS**

### 115 *Study system*

116 *Dalechampia scandens* L. (s.l.) (Euphorbiaceae) is a species complex of perennial woody  
117 vines native to the lowland Neotropics (Armbruster, 1985). It occurs in naturally open areas  
118 and anthropogenic disturbances, including limestone outcrops, open shrublands, light gaps,  
119 and roadsides. Male and female flowers are aggregated into bisexual blossom inflorescences  
120 (pseudanthia – ‘false flowers’), which function as pollination units. Blossoms are  
121 protogynous, with a female phase of c. 2-3 days followed by a bisexual phase. The blossoms  
122 are visited and pollinated by female apid and megachilid bees, which collect floral resin from  
123 a gland-like structure associated with the male flowers (Armbruster, 1985). Outcrossing rates  
124 in four natural populations in Costa Rica ranged from 0.16 to 0.49, indicating a mixed mating  
125 system with a tendency towards selfing (Opedal et al., 2016a). The male and female  
126 subinflorescences are together subtended by a pair of involucre bracts that open during the  
127 day to allow pollination, and close at night to protect the floral tissues (Armbruster, 1985;  
128 Armbruster et al., 1997). The bracts are normally creamy white during anthesis, when they  
129 function as an advertisement towards pollinators. During fruit maturation, the bracts change  
130 color to dark green and increase their rates of photosynthesis, becoming more cryptic and  
131 providing carbon for the developing seeds (Pélabon et al., 2015b). A maximum of nine seeds  
132 per blossom is dispersed by explosive dehiscence of capsules (Armbruster, 1982). Developing  
133 seeds are subject to predation by seed weevils (Curculionidae: Baridinae). The weevils  
134 oviposit on blossoms prior to seed development, and may thus use the bracts to find receptive  
135 inflorescences and/or the bract size as a cue to determine the future availability of resources  
136 for their larvae (Pérez-Barrales et al., 2013).

137 During the peak blooming seasons (Oct.-Dec.) of 2014 and 2015, we studied 20  
138 populations in north-western Costa Rica (see Appendix S1 in the Supplementary Data with  
139 this article for exact locations), belonging to the ‘large-glanded’ taxon of the *D. scandens*  
140 complex (Bolstad et al., 2014). Neighboring populations are separated by 1.8 – 36.5 km, and

141 contemporary gene flow between populations is uncommon or absent (Opedal et al., 2017b).  
142 Further details about the populations are provided in Opedal et al. (2016).

143 *Effects of floral advertisement on seed predation within populations*

144 We studied the effect of a floral-advertisement trait on the probability of seed predation in  
145 each of three populations, one of which was studied in two consecutive years, as part of a  
146 long-term study of spatio-temporal variation in selection on *Dalechampia scandens* blossom  
147 traits (see also Pérez-Barrales et al., 2013). During each study, we marked distinct patches  
148 comprising one or sometimes several intertwined individuals. In each patch, we recorded  
149 daily the number of pollen grains deposited onto the stigmas of individually-marked  
150 blossoms. On the first day of the bisexual phase (the day the first male flower opened), we  
151 measured a set of blossom traits involved in interactions with pollinators and predispersal  
152 seed predators. We also measured the height of each blossom above ground. In this study, we  
153 focused on the area of the upper floral bract, computed as the product of bract length and  
154 width. Bract area is positively correlated with the size of the resin gland and thus represents  
155 an honest signal of the quantity of reward offered to pollinators (Armbruster et al., 2005;  
156 Pélabon et al., 2012). Previous work suggests that bees preferentially visit blossoms with  
157 larger bracts (Armbruster et al., 2005; Pérez-Barrales et al., 2013). However, bract area may  
158 also be shaped by interactions with seed predators (Pérez-Barrales et al., 2013). We collected  
159 developing infructescences approximately four weeks after measurements were made and  
160 recorded the number of viable and predated seeds. Predated seeds are easily identified as  
161 empty seed coats, often with exit holes of adult weevils. Undehisced capsules were dissected  
162 to ascertain whether the seeds had been eaten.

163 *Population-level comparative study: relationships among mating system, seed-predation rate,*  
164 *and upper bract area*

165 Our previous work has shown that blossom traits vary predictably along a gradient of  
166 pollination reliability and mating systems. Here, we focus on the potential of seed predators to  
167 modify the outcome of pollinator-mediated floral-trait evolution. To assess among-population  
168 covariation of predation rate, pollination reliability, and blossom traits, we recorded all or a  
169 subset of these variables in 20 populations (Appendix S1). To quantify realized pollination  
170 reliability, we recorded allogamous pollen loads on the stigmas of female-phase blossoms ( $n$   
171 = 14 – 101 blossoms, mean = 43.2, median = 36), when autogamous selfing is not yet  
172 possible. We measured upper bract area and anther-stigma distance (herkogamy), a key floral  
173 trait mediating mating-system variation among populations and species (Opedal et al., 2017a;

174 Opedal, 2018), on randomly selected blossoms in early bisexual condition ( $n = 4 - 55$   
175 blossoms, mean = 23.9, median = 24). Trait differences among populations are largely  
176 genetically determined, as demonstrated by the correlation of phenotypic traits among  
177 populations measured in the wild and under common-environment greenhouse conditions  
178 (Opedal et al., 2016a). We collected developing infructescences in 20 populations ( $n = 3 -$   
179 101 infructescences, mean = 22.9, median = 13). Following explosive dehiscence of capsules,  
180 we recorded the number of seeds produced, and whether each seed had been eaten. Adult  
181 weevils emerging from predated infructescences were identified as *Nanobaris plumbata*  
182 (Curculionidae: Baridinae), a species distributed apparently from southern Mexico to Panama  
183 (Champion, 1909; J. Prena, pers. com.).

#### 184 Analyses

##### 185 *Effects of floral advertisement on seed predation within populations*

186 We modelled the effect of upper bract area on the probability of seed predation at the blossom  
187 level by fitting a generalized linear mixed-effect model with binomial error distribution and  
188 logit link function (glmmADMB; Fournier et al., 2012). We included only those blossoms  
189 that set seeds in the analysis ( $n = 155$ ). The response variable included the number of seeds  
190 eaten and the number of seeds surviving for each blossom, thus weighing the probability of  
191 predation by the total number of seeds produced. We also included blossom height above  
192 ground as a covariate, and patch as a random effect. This approach treats blossoms nested  
193 within patches as the unit of study, which is justified by the fact that seed predators are  
194 unlikely to differentiate between blossoms on individual vines when these grow intertwined in  
195 a patch (Bolstad et al., 2010; Pérez-Barrales et al., 2013). Thus, we interpret any significant  
196 relationship between floral advertisement and the probability of seed predation at the blossom  
197 level as evidence for seed-predator-mediated phenotypic selection. To test for differences in  
198 seed-predator oviposition patterns among populations, we compared a full model including  
199 interactions between population and population mean-centered upper bract area and blossom  
200 height to a simpler model excluding the bract area  $\times$  population interaction using AICc  
201 (Burnham and Anderson, 2002). Mean-centering was done by subtracting the population  
202 mean from individual trait values. Because environmental factors vary between years and the  
203 insect seed predators in different years are different individuals, we treated the observations of  
204 the Palo Verde population in two consecutive years as different populations.

##### 205 *Population-level comparative analysis*

206 We computed population-level predation rates as the percentage of infructescences with one  
207 or more seeds eaten, and cross-pollination (allogamy) rates as the percentage of blossoms  
208 receiving pollen during the female phase. These measures estimate the intensity of  
209 interactions at the population level and thus the potential for selection (Vanhoenacker et al.,  
210 2013), rather than the absolute abundances of pollinators and seed predators. Substituting  
211 these measures with the average proportion of seeds predated per infructescence and average  
212 stigmatic pollen loads yielded qualitatively identical results (not shown). Because this  
213 analysis concerned long-term evolutionary trends, we pooled data across years to obtain the  
214 best possible population-level estimate.

215 We used path analysis (Shipley, 2016) to assess the independent effects of mating  
216 system and seed predation rate on the evolutionary divergence of upper bract area among  
217 populations. In our study system, female-phase stigmatic pollen loads is a strong predictor of  
218 current outcrossing rate (Opedal et al., 2016a). However, because the current analysis is at the  
219 level of evolved relationships among populations, we used population-mean herkogamy as a  
220 proxy of the long-term mating system of each population, assuming that this would average  
221 out annual fluctuations in outcrossing rates (Opedal, 2018). This approach is justified by the  
222 observation that, across the study populations, herkogamy is positively correlated with  
223 pollination reliability (stigmatic pollen loads), bee abundance on perfume baits, outcrossing  
224 rate, and allelic diversity at microsatellite loci (Opedal et al., 2016a). By treating herkogamy  
225 as a proxy of the mating history of each population, we were able to ‘remove’ the mating-  
226 system related variation in upper bract area when testing the effect of predation rate on upper  
227 bract area. In other words, we tested whether predation rates explained variation in the  
228 residuals of the previously observed relationship between pollination reliability and floral  
229 traits. To achieve this, we obtained path coefficients from a multiple-regression model with  
230 population-mean upper bract area as response variable, and herkogamy and predation rate as  
231 explanatory variables. All variables were standardized to zero mean and unit variance in order  
232 to obtain standardized regression coefficients interpretable as effect sizes in units of standard  
233 deviations. Statistical analyses were conducted in R 3.3.1 (R Core Team, 2018).

## 234 **RESULTS**

### 235 *Effects of floral advertisement on seed predation within populations*

236 Both the absolute and relative intensities of mutualistic and antagonistic interactions differed  
237 among the three focal populations, and between years at Palo Verde (Table 1). At Horizontes  
238 in 2015 and at Palo Verde in 2014, most blossoms were visited by pollinators, and seed



239 predators attacked a substantial proportion of infructescences. At Puente la Amistad in 2014  
240 predation was of comparable magnitude, but pollination was unreliable. Conversely, at Palo  
241 Verde in 2015, pollination was reliable but seed predation was rare.

242 The effect of upper bract area on the probability of seed predation (Fig. 1), after  
243 controlling for effects of blossom height, was population specific (the full model was  
244 supported over the simpler model with no trait  $\times$  population interaction,  $\Delta\text{AICc} = 4.67$ , and  
245 over an intercept-only model,  $\Delta\text{AICc} = 3.90$ ). At Palo Verde in 2015, blossoms with smaller  
246 bracts were more likely to suffer seed predation. At Palo Verde and Puente la Amistad in  
247 2014, and at Horizontes in 2015, the probability of seed predation was independent of upper  
248 bract area. Thus, seed-predator-mediated selection was detected only in the least heavily  
249 predated population (Fig. 1b).

250 *Population-level comparative study: relationships among mating system, seed-predation rate,*  
251 *and upper bract area*

252 Across 20 populations, 409 (13.9%) of the 2933 seeds scored were eaten by seed weevils. The  
253 percentage of predated infructescences at the population level ranged from 0% to 75% (mean  
254 = 34.5%, median = 32.5%,  $n = 20$  populations, Appendix S1). Populations suffering greater  
255 predation rates produced fewer viable seeds ( $r = -0.71$ , 95% CI = -0.88, -0.39). Predation rates  
256 were similar between years across populations (30.6% of infructescences were predated in  
257 2014, and 27.9% in 2015, respectively).

258 Predation rates covaried positively yet non-significantly with rates of allogamous  
259 pollination across populations ( $r = 0.48$ , 95% CI = -0.10, 0.81, Fig. 2). The regression slope of  
260 predation rate on allogamous pollination rate ( $\beta = 0.38 \pm 0.21$ , Fig. 2) was less than one,  
261 corresponding to a tendency for seed predation to be relatively more intense in populations  
262 experiencing less reliable pollination.

263 Herkogamy and upper bract area covaried positively across populations, placing  
264 populations along an axis of increasing trait values (Fig. 3b). After controlling for assumed  
265 mating system (herkogamy), the relationship between predation rate and population-mean  
266 upper bract area was negative yet non-significant (Fig. 3). Hence, there was a weakly  
267 supported tendency for more intensively predated populations to have smaller floral bracts  
268 than expected from their mating-system history.

269 **DISCUSSION**

270 Variation in interactor communities across heterogeneous, fragmented landscapes may lead to  
271 variation in the selective pressures acting on plant populations, i.e. creating geographic  
272 selection mosaics (Thompson, 2005; Gómez et al., 2009; Sun et al., 2016). Two important  
273 parameters needed for predicting the long-term consequences of anthropogenic habitat  
274 destruction and fragmentation for plant populations are therefore i) whether different  
275 interactors (e.g. pollinators vs. antagonists) differ in their response to environmental change  
276 and ii) the degree to which spatial and temporal variation in interaction intensities lead to  
277 differences in selection (Benkman, 2013; Vanhoenacker et al., 2013). Across a fragmented  
278 landscape in north-western Costa Rica, *D. scandens* populations experience contrasting levels  
279 of pollination reliability, and have apparently adapted to the resulting reproductive  
280 environment (Opedal et al., 2016a). Here, we showed that those populations also experience  
281 contrasting intensities of seed predation, with a tendency for greater predation intensities in  
282 populations with more reliable pollination. We also observed a tendency for the intensity of  
283 seed predation relative to pollination to be higher at sites experiencing unreliable pollination  
284 (Fig. 2). For example, the pollinator-poor Puente la Amistad population experienced seed  
285 predation at a rate comparable to or even greater than that observed in the more pollinator-rich  
286 Palo Verde and Horizontes populations (Table 1). However, variation in the intensity of seed  
287 predation has not detectably influenced the evolution of floral advertisement (involucral-bract  
288 size) in *D. scandens* populations.

289         Predispersal seed predation by *Nanobaris* seed weevils occurred independently of  
290 variation in floral advertisement in all populations except Palo Verde in 2015, where the  
291 probability of predation decreased with increasing bract area (Fig. 1). These results are  
292 inconsistent with the expectations that seed predators use increasing floral advertisement as a  
293 cue indicating greater availability of resources for their offspring in the future (Brody, 1992;  
294 Strauss and Irwin, 2004; Strauss and Whittall, 2006; Parachnowitsch and Caruso, 2008;  
295 Pérez-Barrales et al., 2013; Sun et al., 2016), and that the strength of antagonist-mediated  
296 selection increases with increasing interaction intensity (Benkman, 2013; Vanhoenacker et al.,  
297 2013). Although the negative effect detected in the Palo Verde 2015 study was statistically  
298 significant, the low number of infructescences attacked in that study suggests that this effect  
299 could be a false positive (Type I error). We can think of no obvious direct mechanism  
300 explaining weevil preference for smaller floral bracts, although one possibility would be that  
301 bract area correlates negatively with some other trait attractive to weevils, such as color  
302 (Carlson and Holsinger, 2010) or fragrance (Theis and Adler, 2012). These results also

303 contrast with a study conducted in a Mexican *D. scandens* population, where 30% of  
304 infructescences were affected by seed predation, and seed weevils were more likely to  
305 oviposit on blossoms with larger bracts (Pérez-Barrales et al., 2013). This pattern was not  
306 detected in the same population in the following year, however, despite a predation rate of  
307 18% (R. Pérez-Barrales, unpublished data). Overall, in six studies conducted in four different  
308 *D. scandens* populations in two different regions, relationships between floral advertisement  
309 and the probability of seed predation have been detected only twice (including the Palo Verde  
310 2015 study), and in opposite directions. These observations lead us to conclude that, while  
311 seed predators may occasionally mediate phenotypic selection on *D. scandens* blossom traits,  
312 it occurs infrequently and is not predictable from population-mean seed predation rates within  
313 the range observed in the *D. scandens* study populations.

314         The hypothesis that joint attraction of pollinators and seed predators generate  
315 conflicting selection on floral advertisement has received reasonably strong empirical support  
316 (Brody, 1992; Brody and Waser, 1995; Brody and Mitchell, 1997; Cariveau et al., 2004;  
317 Pérez-Barrales et al., 2013; Sun et al., 2016), yet the likelihood of detecting seed-predator-  
318 mediated selection may depend on several factors. Working in a large set of *Primula farinosa*  
319 populations, Vanhoenacker et al. (2013) demonstrated stronger seed-predator-mediated  
320 selection at greater intensities of predation, yet the relationship tended to be non-linear and  
321 accelerating at greater predation intensities. Furthermore, the strength of seed-predator-  
322 mediated selection may depend not only on the intensity of seed predation, but also on the  
323 reliability of pollination (Brody, 1992; Vanhoenacker et al., 2013). This effect arises in part  
324 because the reliability of pollination may affect the degree to which seed set differs between  
325 attractive vs. non-attractive phenotypes, and hence the reliability of floral advertisements as a  
326 cue indicating seed production. While pollen limitation on seed set has been demonstrated in  
327 at least one study detecting predator-mediated selection (Brody, 1992), others have detected  
328 selection in the absence of apparent pollen limitation (Cariveau et al., 2004; Parachnowitsch  
329 and Caruso, 2008). Furthermore, Bartkowska and Johnston (2012) found that pollinators, but  
330 not seed predators, mediate selection in a pollen-limited population of *Lobelia cardinalis*. In  
331 the current study, we failed to detect seed-predator-mediated selection across focal  
332 populations that differed both in the rate of predation and in the reliability of pollination.  
333 Thus, while further work is needed to resolve these contrasting results, one possible  
334 explanation for the lack of weevil choosiness within *D. scandens* populations is that variation  
335 in seed set is too limited to generate strong preferences for floral phenotypes associated with

336 larger seed sets (see Brody, 1992). Such effects could perhaps be expected when pollination is  
337 reliable, or when a self-compatible species is capable of effective autonomous self-  
338 pollination.

339         The general lack of detectable effects of upper bract area on seed predation within  
340 populations was mirrored in the patterns observed among populations. If seed predators  
341 preferentially oviposit on large-bracted blossoms, we would expect a negative relationship  
342 between predation rate and the size of advertisement traits among populations. After  
343 controlling for mating-system-related variation in upper bract area by including herkogamy in  
344 the path analysis, the relationship between predation rate and upper bract area was indeed  
345 negative, but statistically non-significant (Fig. 3). This result argues against our causal  
346 hypothesis, that the intensity of seed predation determines the strength of predator-mediated  
347 selection and thus contributes to the evolutionary divergence of floral traits. Weak  
348 relationships between current predation intensity, strength of selection, and floral traits also  
349 argue against the alternative hypothesis that seed predators are differentially attracted to  
350 populations with different mean trait values (Dart and Eckert, 2015). There are at least two  
351 non-mutually-exclusive explanations for this finding. First, the weak relationships between  
352 the phenotypic traits and current predation intensity may be due to recent changes in predation  
353 intensity. Indeed, while variation in herkogamy and upper bract area represents the outcome  
354 of long-term interactions with pollinators, antagonists, and other selective factors, our data on  
355 predation intensity were collected over only two years. Second, the current and/or long-term  
356 relative abundance of seed weevils across populations may depend on other factors not  
357 included in our analysis.

358         Dart and Eckert (2015) have suggested that florivores are attracted to large-flowered  
359 populations of *Camissoniopsis cheiranthifolia* due to the greater quantity of resources  
360 available. Similarly, seed weevils benefit from laying eggs on blossoms producing many  
361 seeds, and seed predation rates might therefore depend on the average seed set in each  
362 population. While the average number of outcrossed seeds produced in *D. scandens*  
363 populations is likely to increase with the rate of allogamous pollination, and indirectly with  
364 blossom size, the opposite may be true for the number of selfed seeds. In *D. scandens*, seeds  
365 resulting from selfing are the same size as those resulting from outcrossing (Opedal et al.,  
366 2015; Pélabon et al., 2015a), suggesting they are equally valuable as resources for seed  
367 predators. Indeed, mating-system-related local adaptation in mixed-mating plant species may  
368 tend to reduce among-population variation in seed set, and hence the amount of resources

369 available for seed predators. If seed predators respond to mean resource availability at the  
370 population level, seed predation rates would then be expected to vary independently from  
371 pollination environments and floral traits, as observed in *D. scandens*. We lack data on  
372 average open-pollinated seed set for most of our study populations, precluding a strong test of  
373 this hypothesis. However, the range observed across our focal populations was indeed limited  
374 (Table 1), and across all populations the range of seed sets of blossoms developing fruits (and  
375 hence collected for the purpose of this study) was also relatively limited (mean = 6.45 seeds,  
376 s.d. = 0.82, range = 4.23 - 8.25, Appendix S1).

377 Judging from their rates of interaction with *D. scandens* populations, the relative  
378 abundances of bee pollinators and coleopteran seed predators are largely decoupled across our  
379 study area, suggesting that these species groups respond differently to landscape-scale  
380 environmental heterogeneity. While pollination tended to be more reliable in populations  
381 occurring in less-disturbed habitats (along gravel roads through forested areas), predation  
382 tended to be more intense in heavily disturbed habitats along highways (Appendix S1). The  
383 primary pollinators of *D. scandens* in the study area are female euglossine bees (Table 1, Fig.  
384 3). These forest-associated bees appear largely to avoid highway roadsides and other heavily  
385 disturbed habitats (Brosi, 2009; Briggs et al., 2013; Opedal et al., 2017b), suggesting that  
386 habitat destruction may be an indirect driver of mating-system and floral-trait evolution in this  
387 system. In contrast, *Nanobaris* seed weevils (Fig. 3) seem less affected by habitat type in our  
388 study area and were relatively abundant at several pollinator-poor sites. If low population  
389 densities of euglossine bees at heavily disturbed sites is indeed caused by habitat destruction,  
390 this finding adds to previous studies suggesting stronger effects of habitat destruction and  
391 fragmentation on mutualistic than antagonistic interactions (Magrach et al., 2014; Brudvig et  
392 al., 2015).

393 The main conclusion of this study is that, although predispersal seed predation may  
394 reduce the average fitness of *D. scandens* populations and thus be ecologically important  
395 (Kolb et al., 2007), seed-predator-mediated selection is unlikely to have been an important  
396 driver of floral evolution, at least for the traits we measured. By excluding a possible  
397 confounding effect, this observation strengthens our previous conclusion that pollination-  
398 related selection drives the evolutionary divergence of blossom traits among *D. scandens*  
399 populations (Opedal et al., 2016a). Specifically, the study populations appear to have tracked  
400 variation in adaptive landscapes generated by pollinator communities, but seed predators had  
401 only minor modifying effects on the outcome of pollinator-mediated evolution of blossom

402 traits. Hence, seed predators are unlikely to constrain pollinator-mediated floral-trait evolution  
403 in this system.

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#### 412 **Author contributions**

413 ØHO, EA, RPB, and WSA conducted field work. ØHO conducted lab work, analyzed data,  
414 and wrote the manuscript with contributions from all authors.

#### 415 **Literature Cited**

- 416 Adler, L. S., and J. L. Bronstein. 2004. Attracting antagonists: does floral nectar increase leaf  
417 herbivory? *Ecology* 85: 1519-1526.
- 418 Aguilar, R., L. Ashworth, L. Galetto, and M. A. Aizen. 2006. Plant reproductive susceptibility to habitat  
419 fragmentation: review and synthesis through a meta-analysis. *Ecology Letters* 9: 968-980.
- 420 Armbruster, W. S. 1982. Seed production and dispersal in *Dalechampia* (Euphorbiaceae) - divergent  
421 patterns and ecological consequences. *American Journal of Botany* 69: 1429-1440.
- 422 Armbruster, W. S. 1985. Patterns of character divergence and the evolution of reproductive ecotypes  
423 of *Dalechampia scandens* (Euphorbiaceae). *Evolution* 39: 733-752.
- 424 Armbruster, W. S., L. Antonsen, and C. Pélabon. 2005. Phenotypic selection on *Dalechampia*  
425 blossoms: honest signaling affects pollination success. *Ecology* 86: 3323-3333.
- 426 Armbruster, W. S., J. J. Howard, T. P. Clausen, E. M. Debevec, J. C. Loquvam, M. Matsuki, B.  
427 Cerendolo, and F. Andel. 1997. Do biochemical exaptations link evolution of plant defense  
428 and pollination systems? Historical hypotheses and experimental tests with *Dalechampia*  
429 vines. *American Naturalist* 149: 461-484.
- 430 Armbruster, W. S., and G. L. Webster. 1982. Divergent pollination systems in sympatric species of  
431 South American *Dalechampia* (Euphorbiaceae). *American Midland Naturalist* 108: 325-337.
- 432 Bartkowska, M. P., and M. O. Johnston. 2012. Pollinators cause stronger selection than herbivores  
433 on floral traits in *Lobelia cardinalis* (Lobeliaceae). *New Phytologist* 193: 1039-1048.
- 434 Benkman, C. W. 2013. Biotic interaction strength and the intensity of selection. *Ecology Letters* 16:  
435 1054-1060.
- 436 Bolstad, G. H., W. S. Armbruster, C. Pélabon, R. Pérez-Barrales, and T. F. Hansen. 2010. Direct  
437 selection at the blossom level on floral reward by pollinators in a natural population of  
438 *Dalechampia schottii*: full-disclosure honesty? *New Phytologist* 188: 370-384.
- 439 Bolstad, G. H., T. F. Hansen, C. Pélabon, M. Falahati-Anbaran, R. Pérez-Barrales, and W. S.  
440 Armbruster. 2014. Genetic constraints predict evolutionary divergence in *Dalechampia*  
441 blossoms. *Philos Trans R Soc Lond B Biol Sci* 369: 20130255.

442 Briggs, H. M., I. Perfecto, and B. J. Brosi. 2013. The role of the agricultural matrix: coffee  
443 management and euglossine bee (Hymenoptera: Apidae: Euglossini) communities in  
444 southern Mexico. *Environmental Entomology* 42: 1210-1217.

445 Brody, A. K. 1992. Oviposition choices by a pre-dispersal seed predator (*Hylemya* sp.) I.  
446 Correspondence with hummingbird pollinators, and the role of plant size, density and floral  
447 morphology. *Oecologia* 91: 56-62.

448 Brody, A. K., and R. J. Mitchell. 1997. Effects of experimental manipulation of inflorescence size on  
449 pollination and pre-dispersal seed predation in the hummingbird-pollinated plant *Ipomopsis*  
450 *aggregata*. *Oecologia* 110: 86-93.

451 Brody, A. K., and N. M. Waser. 1995. Oviposition patterns and larval success of a pre-dispersal seed  
452 predator attacking two confamilial host plants. *Oikos* 74: 447-452.

453 Brosi, B. J. 2009. The effects of forest fragmentation on euglossine bee communities (Hymenoptera:  
454 Apidae: Euglossini). *Biological Conservation* 142: 414-423.

455 Brudvig, L. A., E. I. Damschen, N. M. Haddad, D. J. Levey, and J. J. Tewksbury. 2015. The influence of  
456 habitat fragmentation on multiple plant–animal interactions and plant reproduction. *Ecology*  
457 96: 2669-2678.

458 Brys, R., and H. Jacquemyn. 2012. Effects of human-mediated pollinator impoverishment on floral  
459 traits and mating patterns in a short-lived herb: an experimental approach. *Functional*  
460 *Ecology* 26: 189-197.

461 Burnham, K. P., and D. R. Anderson. 2002. Model selection and multi-model inference: a practical  
462 information-theoretic approach. 2nd ed. Springer, New York, NY, US.

463 Cariveau, D., R. E. Irwin, A. K. Brody, L. S. Garcia-Mayeya, and A. Von Der Ohe. 2004. Direct and  
464 indirect effects of pollinators and seed predators to selection on plant and floral traits. *Oikos*  
465 104: 15-26.

466 Carlson, J. E., and K. E. Holsinger. 2010. Natural selection on inflorescence color polymorphisms in  
467 wild *Protea* populations: The role of pollinators, seed predators, and intertrait correlations.  
468 *American Journal of Botany* 97: 934-944.

469 Champion, G. C. 1909. *Biologia Centrali-Americana*. Insecta. Coleoptera. Vol. IV, pt. 5, p. 475.

470 Cunningham, S. A. 2000. Effects of habitat fragmentation on the reproductive ecology of four plant  
471 species in Mallee woodland. *Conservation Biology* 14: 758-768.

472 Dart, S., and C. G. Eckert. 2015. Variation in pollen limitation and floral parasitism across a mating  
473 system transition in a Pacific coastal dune plant: evolutionary causes or ecological  
474 consequences? *Annals of Botany* 115: 315-326.

475 Eckert, C. G., S. Kalisz, M. A. Geber, R. Sargent, E. Elle, P. O. Cheptou, C. Goodwillie, et al. 2010. Plant  
476 mating systems in a changing world. *Trends in Ecology and Evolution* 25: 35-43.

477 Fournier, D. A., H. J. Skaug, J. Ancheta, J. Ianelli, A. Magnusson, M. N. Maunder, A. Nielsen, and J.  
478 Sibert. 2012. AD Model Builder: using automatic differentiation for statistical inference of  
479 highly parameterized complex nonlinear models. *Optimization Methods & Software* 27: 233-  
480 249.

481 Garcia, D., and N. P. Chacoff. 2007. Scale-dependent effects of habitat fragmentation on hawthorn  
482 pollination, frugivory, and seed predation. *Conservation Biology* 21: 400-411.

483 Gómez, J. M., F. Perfectti, J. Bosch, and J. P. M. Camacho. 2009. A geographic selection mosaic in a  
484 generalized plant-pollinator-herbivore system. *Ecological Monographs* 79: 245-263.

485 Gómez, J. M., and R. Zamora. 2000. Spatial variation in the selective scenarios of *Hormathophylla*  
486 *spinosa* (Cruciferae). *American Naturalist* 155: 657-668.

487 Kolb, A., and J. Ehrlen. 2010. Environmental context drives seed predator-mediated selection on a  
488 floral display trait. *Evolutionary Ecology* 24: 433-445.

489 Kolb, A., J. Ehrlen, and O. Eriksson. 2007. Ecological and evolutionary consequences of spatial and  
490 temporal variation in pre-dispersal seed predation. *Perspectives in Plant Ecology, Evolution*  
491 *and Systematics* 9: 79-100.

492 Koski, M. H., D. Grossenbacher, J. W. Busch, and L. F. Galloway. 2017. A geographic cline in the  
493 ability to self-fertilize is unrelated to the pollination environment. *Ecology* 98: 2930-2939.

494 Magrath, A., W. F. Laurance, A. R. Larrinaga, and L. Santamaria. 2014. Meta-analysis of the effects of  
495 forest fragmentation on interspecific interactions. *Conservation Biology* 28: 1342-1348.

496 Magurran, A. E. 2016. How ecosystems change. *Science* 351: 448-449.

497 Moeller, D. A. 2006. Geographic structure of pollinator communities, reproductive assurance, and  
498 the evolution of self-pollination. *Ecology* 87: 1510-1522.

499 Opedal, Ø. H. 2018. Herkogamy, a principal functional trait of plant reproductive biology.  
500 *International Journal of Plant Sciences* DOI: 10.1086/700314.

501 Opedal, Ø. H., E. Albertsen, W. S. Armbruster, R. Pérez-Barrales, M. Falahati-Anbaran, and C.  
502 Pélabon. 2016a. Evolutionary consequences of ecological factors: pollinator reliability  
503 predicts mating-system traits of a perennial plant. *Ecology Letters* 19: 1486-1495.

504 Opedal, Ø. H., E. Albertsen, W. S. Armbruster, R. Pérez-Barrales, M. Falahati-Anbaran, and C.  
505 Pélabon. 2016b. Data from: Evolutionary consequences of ecological factors: pollinator  
506 reliability predicts mating-system traits of a perennial plant. Dryad Digital Repository. DOI:  
507 <https://doi.org/10.5061/dryad.8ph35>.

508 Opedal, Ø. H., W. S. Armbruster, and C. Pélabon. 2015. Inbreeding effects in a mixed-mating vine:  
509 effects of mating history, pollen competition and stress on the cost of inbreeding. *AoB*  
510 *PLANTS* 7: plv133.

511 Opedal, Ø. H., G. H. Bolstad, T. F. Hansen, W. S. Armbruster, and C. Pélabon. 2017a. The evolvability  
512 of herkogamy: quantifying the evolutionary potential of a composite trait. *Evolution* 71:  
513 1572-1586.

514 Opedal, Ø. H., M. Falahati-Anbaran, E. Albertsen, W. S. Armbruster, R. Pérez-Barrales, H. K. Stenøien,  
515 and C. Pélabon. 2017b. Euglossine bees mediate only limited long-distance gene flow in a  
516 tropical vine. *New Phytologist* 213: 1898-1908.

517 Parachnowitsch, A. L., and C. M. Caruso. 2008. Predispersal seed herbivores, not pollinators, exert  
518 selection on floral traits via female fitness. *Ecology* 89: 1802-1810.

519 Pélabon, C., E. Albertsen, M. Falahati-Anbaran, J. Wright, and W. S. Armbruster. 2015a. Does  
520 multiple paternity affect seed mass in angiosperms? An experimental test in *Dalechampia*  
521 *scandens*. *Journal of Evolutionary Biology* 28: 1719-1733.

522 Pélabon, C., L. Hennet, R. Strimbeck, H. Johnson, and W. S. Armbruster. 2015b. Blossom colour  
523 change after pollination provides carbon for developing seeds. *Functional Ecology* 29: 1137-  
524 1143.

525 Pélabon, C., P. Thöne, T. F. Hansen, and W. S. Armbruster. 2012. Signal honesty and cost of pollinator  
526 rewards in *Dalechampia scandens* (Euphorbiaceae). *Annals of Botany* 109: 1331-1339.

527 Pérez-Barrales, R., G. H. Bolstad, C. Pélabon, T. F. Hansen, and W. S. Armbruster. 2013. Pollinators  
528 and seed predators generate conflicting selection on *Dalechampia* blossoms. *Oikos* 122:  
529 1411-1428.

530 Potts, S. G., J. C. Biesmeijer, C. Kremen, P. Neumann, O. Schweiger, and W. E. Kunin. 2010. Global  
531 pollinator declines: trends, impacts and drivers. *Trends in Ecology & Evolution* 25: 345-353.

532 R Core Team. 2018. R: A language and environment for statistical computing. website: [http://www.R-](http://www.R-project.org/)  
533 [project.org/](http://www.R-project.org/).

534 Rey, P. J., C. M. Herrera, J. Guitian, X. Cerda, A. M. Sanchez-Lafuente, M. Medrano, and J. L. Garrido.  
535 2006. The geographic mosaic in predispersal interactions and selection on *Helleborus*  
536 *foetidus* (Ranunculaceae). *Journal of Evolutionary Biology* 19: 21-34.

537 Shipley, B. 2016. Cause and correlation in biology. A user's guide to path analysis, structural  
538 equations and causal inference with R. Cambridge University Press, Cambridge.

539 Steffan-Dewenter, I., U. Münzenberg, and T. Tschardt. 2001. Pollination, seed set and seed  
540 predation on a landscape scale. *Proceedings of the Royal Society B-Biological Sciences* 268:  
541 1685-1690.

542 Strauss, S. Y., and W. S. Armbruster. 1997. Linking herbivory and pollination - New perspectives on  
543 plant and animal ecology and evolution. *Ecology* 78: 1617-1618.

544 Strauss, S. Y., and R. E. Irwin. 2004. Ecological and evolutionary consequences of multispecies plant-  
545 animal interactions. *Annual Review of Ecology Evolution and Systematics* 35: 435-466.



- 546 Strauss, S. Y., and J. B. Whittall. 2006. Non-pollinator agents of selection on floral traits. *In* L. D.  
547 Harder AND S. C. Barrett [eds.], *Ecology and Evolution of Flowers*. Oxford University Press,  
548 Oxford, UK.
- 549 Sun, S. G., W. S. Armbruster, and S. Q. Huang. 2016. Geographic consistency and variation in  
550 conflicting selection generated by pollinators and seed predators. *Annals of Botany* 118: 227-  
551 237.
- 552 Theis, N., and L. S. Adler. 2012. Advertising to the enemy: enhanced floral fragrance increases beetle  
553 attraction and reduces plant reproduction. *Ecology* 93: 430-435.
- 554 Thomann, M., E. Imbert, C. Devaux, and P.-O. Cheptou. 2013. Flowering plants under global  
555 pollinator decline. *Trends in Plant Science* 18: 353-359.
- 556 Thompson, J. N. 2005. *The geographic mosaic of coevolution*. The University of Chicago Press,  
557 Chicago, US.
- 558 Urban, M. C., G. Bocedi, A. P. Hendry, J. B. Mihoub, G. Pe'er, A. Singer, J. R. Bridle, et al. 2016.  
559 Improving the forecast for biodiversity under climate change. *Science* 353.
- 560 Vanhoenacker, D., J. Ågren, and J. Ehrlen. 2013. Non-linear relationship between intensity of plant-  
561 animal interactions and selection strength. *Ecology Letters* 16: 198-205.

562

### 563 **Data Accessibility**

564 Population-level predation and pollination data are available in Appendix S1 in the  
565 Supplementary Data with this article. Trait measurements are available in the Dryad Digital  
566 Repository: <https://doi.org/10.5061/dryad.8ph35> (Opedal et al., 2016b).

567

568 **Figure legends**

569 Fig. 1. (a) Effects of population-mean centered upper bract area (UBA) on the probability of  
570 seed predation within *Dalechampia scandens* populations at Puente la Amistad (yellow  
571 curve), Palo Verde (blue curve, 2014; green curve, 2015) and Horizontes (red curve). (b)  
572 Relationship between population-level predation rate and  $\beta_{\text{UBA}}$ , the slope of the logistic  
573 regression of predation probability (P) on upper bract area. Circle sizes in (a) are proportional  
574 to blossom seed set, and error bars in (b) indicate standard errors.

575 Fig. 2. Relationship between allogamy rate (percentage of inflorescences receiving  
576 allogamous pollen) and seed predation rate (percentage of inflorescences suffering seed  
577 predation) across 13 *D. scandens* populations in Costa Rica. The solid line indicates the 1:1  
578 relationship, where mutualistic and antagonistic interactions are equally intense, and the  
579 dashed line indicates the estimated regression slope.

580 Fig. 3. Path diagram and scatterplots showing effects of mating system (population-mean  
581 herkogamy) and predation rate (proportion of infructescences attacked by *Nanobaris* seed  
582 weevils) on population-mean upper bract area (UBA, a floral advertisement trait). The  
583 double-headed curved arrow indicates the correlation (a) between herkogamy and predation  
584 rate. Single-headed arrows indicate direct effects of mating system (b) and predation rate (c)  
585 on upper bract area and are given with standardized regression coefficients interpretable as  
586 effect size in units of standard deviations.  $U = \text{unexplained variation, computed as } \sqrt{1 - r^2}$ .  
587 Circle sizes in panels (a) and (c) are proportional to the square root of the sample size for  
588 predation rates, and error bars in (b) indicate standard errors. Photographs by Ø. H. Opedal.

Table 1. Summary of biotic interactions in *Dalechampia scandens* populations in Costa Rica. Pollinator visitation rates are the percentages of observed pollinator visits made by members of each pollinator genus, pollination rate is the percentage of blossoms receiving allogamous pollen, predation rate is the percentage of infructescences that developed seeds and were then predated, and seed set is population-mean open-pollinated seed number (out of a maximum of 9 seeds), including those seeds scored as predated.

Population: Year	Coordinates	Pollinator visitation rates (%)			Pollination rate (%)	Predation rate (%)	Seed set (s.d)
		<i>Hypanthidium</i>	<i>Euglossa</i>	<i>Eufriesea</i>			
Puente la Amistad: 2014	10°14' N, 85°15' W	13.0	26.1	60.9	30.3	31.4	2.32 (3.27)
Palo Verde: 2014	10°23' N, 85°19' W	22.7	77.3		88.1	37.5	3.58 (3.46)
Palo Verde: 2015	10°23' N, 85°19' W	19.1	30.9	50.0	78.6	9.1	2.32 (3.30)
Horizontes: 2015	10°42' N, 85°36' W	43.2	56.8		82.8	40.4	2.75 (3.68)