

1 White-tailed eagle (*Haliaeetus albicilla*) feathers from Norway are suitable for  
2 monitoring of legacy, but not emerging contaminants

3 Mari E. Løseth<sup>a,\*</sup>, Nathalie Briels<sup>a</sup>, Jørgen Flo<sup>a</sup>, Govindan Malarvannan<sup>b</sup>, Giulia Poma<sup>b</sup>, Adrian Covaci<sup>b</sup>,  
4 Dorte Herzke<sup>c</sup>, Torgeir Nygård<sup>d</sup>, Jan O. Bustnes<sup>e</sup>, Bjørn M. Jenssen<sup>a</sup>, Veerle L. B. Jaspers<sup>a</sup>

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6 **Affiliations:**

7 <sup>a</sup>Department of Biology, Norwegian University of Science and Technology (NTNU),  
8 Høgskoleringen 5, 7491 Trondheim, Norway

9 <sup>b</sup>Toxicological Center, University of Antwerp, Universiteitsplein 1, 2610 Wilrijk, Belgium

10 <sup>c</sup>Norwegian Institute for Air Research (NILU), FRAM - High North Research Centre on  
11 Climate and the Environment, 9007 Tromsø, Norway

12 <sup>d</sup>Norwegian Institute for Nature Research (NINA), Høgskoleringen 9, 7034 Trondheim,  
13 Norway

14 <sup>e</sup>Norwegian Institute for Nature Research (NINA), FRAM - High North Research Centre on  
15 Climate and the Environment, 9007 Tromsø, Norway

16

17

18 **\*Corresponding author:**

19 Mari Engvig Løseth: mari.loseth@ntnu.no

20

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## 25 **Abstract**

26 While feathers have been successfully validated for monitoring of internal concentrations of heavy  
27 metals and legacy persistent organic pollutants (POPs), less is known about their suitability for  
28 monitoring of emerging contaminants (ECs). Our study presents a broad investigation of both legacy  
29 POPs and ECs in non-destructive matrices from a bird of prey. Plasma and feathers were sampled in  
30 2015 and 2016 from 70 white-tailed eagle (*Haliaeetus albicilla*) nestlings from two archipelagos in  
31 Norway. Preen oil was also sampled in 2016. Samples were analysed for POPs (polychlorinated  
32 biphenyls (PCBs), polybrominated diphenyl ethers (PBDEs) and organochlorinated pesticides (OCPs))  
33 and ECs (per- and polyfluoroalkyl substances (PFASs), dechlorane plus (DPs), phosphate and novel  
34 brominated flame retardants (PFRs and NBFRs)). A total of nine PCBs, three OCPs, one PBDE and  
35 one PFAS were detected in over 50 % of the plasma and feather samples within each sampling year and  
36 location. Significant and positive correlations were found between plasma, feathers and preen oil  
37 concentrations of legacy POPs and confirm the findings of previous research on the usefulness of these  
38 matrices for non-destructive monitoring. In contrast, the suitability of feathers for ECs seems to be  
39 limited. Detection frequencies (DF) of PFASs were higher in plasma (mean DF: 78 %) than in feathers  
40 (mean DF: 38 %). Only perfluoroundecanoic acid could be quantified in over 50 % of both plasma and  
41 feather samples, yet their correlation was poor and not significant. The detection frequencies of PFRs,  
42 NBFRs and DPs were very low in plasma (mean DF: 1 - 13 %), compared to feathers (mean DF: 10 -  
43 57 %). This may suggest external atmospheric deposition, rapid internal biotransformation or excretion  
44 of these compounds. Accordingly, we suggest prioritising plasma for PFASs analyses, while the sources  
45 of PFRs, NBFRs and DPs in feathers and plasma need further investigation.

46

## 47 **1. Introduction**

48 Polychlorinated biphenyls (PCBs), organochlorinated pesticides (OCPs) and polybrominated  
49 diphenyl ethers (PBDEs) are compounds previously used in industrial applications, agriculture  
50 and consumer products (Mackay et al., 2006). Classified as persistent organic pollutants  
51 (POPs), these compounds are generally lipophilic, semi-volatile and resistant to chemical and

52 biological degradation (Buccini, 2003; Mackay et al., 2006). Consequently, POPs persist in the  
53 environment (Letcher et al., 2010; Mackay et al., 2006) and may result in high uptake in biota,  
54 followed by bioaccumulation and biomagnification, especially in long and lipid-rich food webs  
55 (Borgå et al., 2004; Jones and de Voogt, 1999). As replacements for the legacy POPs regulated  
56 by the Stockholm Convention (UNEP, 2009), new and (re-) emerging contaminants (ECs) have  
57 entered the market. Those include phosphorus flame retardants (PFRs; van der Veen and de  
58 Boer, 2012), “novel” brominated flame retardants (NBFRs; Covaci et al., 2011), dechlorane  
59 plus (DPs; Sverko et al., 2011) and certain per- and polyfluoroalkyl substances (PFASs; Lau et  
60 al., 2007). These ECs exhibit different physicochemical properties than the legacy POPs and  
61 may accumulate in other matrices, such as protein-rich tissues (Lau et al., 2007), or become  
62 rapidly metabolised and excreted (Briels et al., 2018; Covaci et al., 2011; van der Veen and de  
63 Boer, 2012).

64  
65 Wild birds are important biomonitors for numerous environmental contaminants (Burger and  
66 Gochfeld, 2004; Furness, 1993). Due to ethical and species conservational aspects, non-  
67 destructive sampling methods such as the collection of blood or addled eggs are often applied  
68 in environmental monitoring programs of wild birds (Espín et al., 2016). The contaminant  
69 concentrations detected in blood plasma provide a snapshot of recent exposure through diet  
70 (Henriksen et al., 1998), but during periods of low food availability or starvation concentrations  
71 can also originate from internal fat reserves (re-exposure) (Fenstad et al., 2014). Egg  
72 concentrations on the other hand reflect maternal concentrations deposited during the egg  
73 formation (Becker and Sperveslage, 1989). Feathers, either plucked or moulted, present another  
74 non-destructive sampling matrix. Feathers are connected to the blood circulation during  
75 formation and growth, and during this period the internal contaminant concentrations may

76 thereby be transferred and deposited into the feather (Jaspers et al. 2006; García-Fernández et  
77 al., 2013).

78

79 The use of feathers as a non-destructive matrix for biomonitoring is increasing (García-  
80 Fernández et al., 2013; Gómez-Ramírez et al., 2014). While feathers have been used for  
81 decades as a matrix for monitoring environmental concentrations of metal (Burger, 1993), it  
82 was only in the early 2000s that feathers were proposed for legacy POP analyses (Dauwe et al.,  
83 2005; Jaspers et al., 2006). Recently, feathers have also been investigated as a matrix for  
84 analysing and monitoring PFASs (Gómez-Ramírez et al., 2017; Jaspers et al., 2013; Li et al.,  
85 2017; Meyer et al., 2009), and only a few studies published to date have investigated the  
86 suitability of NBFs and PFRs monitoring in feathers (Eulaers et al., 2014; Svendsen et al.,  
87 2018). Consequently, little is known about the exposure to and deposition of these ECs into  
88 feathers. Preen oil has also been proposed as a non-destructive matrix for monitoring PCBs,  
89 PBDEs and OCPs (Eulaers et al., 2011b; Van den Brink, 1997), but few studies have collected  
90 preen oil for contaminant analyses (Eulaers et al., 2011a, 2011b; Van den Brink, 1997).

91

92 Studies investigating non-destructive sampling matrices in birds have been conducted on a  
93 wide variety of bird species (García-Fernández et al., 2013). However, there is a general lack  
94 of studies with larger sample sizes that have investigated both legacy POPs and ECs in several  
95 non-destructive matrices (Espín et al., 2016; García-Fernández et al., 2013). This may improve  
96 the evaluation of the suitability of these matrices for monitoring purposes. An overview of  
97 contaminant monitoring activities in Europe revealed that 100 monitoring programs from 28  
98 countries have included feathers samples from birds of prey (Espín et al., 2016).

99

100 Due to their apex trophic position, large body size and long lifespan, birds of prey such as the  
101 white-tailed eagle (*Haliaeetus albicilla*), are good sentinel species for monitoring the presence  
102 of contaminants in the environment (Burger and Gochfeld, 2004). White-tailed eagle nestlings  
103 are stationary in their nests and therefore good indicators of local exposure to a wide range of  
104 environmental contaminants (Olsson et al., 2000). They are also relatively easy to sample while  
105 still in the nest (Espín et al., 2016; Eulaers et al., 2011b). The white-tailed eagle was listed as  
106 threatened by the International Union for Conservation of Nature in 1988, but today it is listed  
107 as of least concern (Birdlife Int., 2016).

108

109 In this study, we aimed to evaluate if body feathers and preen oil from white-tailed eagle  
110 nestlings present a good non-destructive matrix to monitor internal concentrations of both  
111 legacy POPs and ECs. Consequently, we investigated concentrations of legacy POPs and ECs  
112 in plasma, feathers and preen oil from 70 white-tailed eagle nestlings. Furthermore, we  
113 investigated correlations of POP and EC concentrations in these matrices and evaluated the  
114 consistency of these results by including samples from two field locations during two  
115 consecutive years. As the sampled feathers were still growing and connected to the blood  
116 circulation, we expected to find strong correlations between feathers and plasma concentrations  
117 of POPs and ECs. We also expected to find strong correlations between plasma and preen oil,  
118 as the oil is produced by an internal gland which is connected to the blood circulation.

119

## 120 **2. Materials and methods**

### 121 **2.1. Field sampling**

122 The study was conducted on 70 white-tailed eagle nestlings from two archipelagos in Norway,  
123 Smøla (63.35°N; 8.03°E) and Steigen (67.93°N; 14.98°E), during the breeding seasons of 2015  
124 and 2016. We sampled 13 nestlings in Smøla in 2015 and 22 nestlings in 2016. In Steigen, 14

125 nestlings were sampled in 2015 and 21 nestlings in 2016. All nestlings, aged from 8-12 weeks  
126 old, were caught at the nest site and handled for approximately 15 min. Body feathers were  
127 gently pulled from the dorsal region, approximately 10 per individual, and stored in  
128 polyethylene zipper bags (VWR, USA) at -20°C. A blood sample of 8 mL was collected in  
129 heparinised vacutainers through brachial venepuncture. The blood samples were centrifuged  
130 (860 g), after which plasma was transferred to cryogenic tubes (Nalgene®, USA) and stored at  
131 -20 °C. Preen oil could only be collected in a sufficient amount in 2016. It was collected in a  
132 1.5 mL Eppendorf tube (VWR, USA) by massaging the preen gland using disposable gloves  
133 and avoiding traces of feathers in the sample. The sampling was approved by the Norwegian  
134 Food Safety Authority (Mattilsynet; 2015/6432 and 2016/8709) and the handling of the birds  
135 were in accordance with the regulations of the Norwegian Animal Welfare Act.

136

## 137 **2.2. Chemical analyses**

### 138 ***2.2.1. Feather pre-treatment***

139 Clean stainless steel and glass tools were used to wash and cut the feathers. Tools were  
140 thoroughly rinsed between individual samples with acetone for POP and EC analyses and  
141 methanol for PFASs analyses. The feather quills (calamus) were removed and remaining  
142 feathers were washed in MilliQ-water to remove dust and particles from the feathers prior to  
143 analysis (Jaspers et al., 2007a, 2007b, 2008). For a thorough wash, two pairs of tweezers were  
144 used to separate the barbs by pulling the barbs downwards and away from each other. Feathers  
145 were placed on clean lab paper, covered with tissue paper (Facial tissues, VWR) and dried  
146 overnight at room temperature. Finally, the feathers were cut into approximately 1–2 mm  
147 pieces and homogenates were accurately weighed prior to analyses (range: 0.10 – 0.40 grams).  
148 The feather pre-treatment was conducted on the bench in a clean lab (not used for chemical  
149 analyses).

150 **2.2.2. Legacy POPs and ECs**

151 Chemical analyses of legacy POPs and ECs in feathers, plasma and preen oil were performed  
152 at the Toxicological Centre of the University of Antwerp, Belgium. The targeted compounds  
153 for the analyses were 23 PCBs, 10 OCPs, seven PBDEs, eight PFRs, three NBFRs and two  
154 DPs. The full compound list can be found in the Supplementary information (SI), Tables S1  
155 and S2. Contents of the internal standards (IS1 (POPs), IS2 (ECs) and IS3 (DPs)) can be found  
156 in Table S3.

157

158 **Plasma extraction:** One mL of plasma was spiked with 100  $\mu$ L IS1 and with 40  $\mu$ L IS2. To  
159 this, 1 mL of Milli-Q water, 200  $\mu$ L of formic acid (98 %) and 4 mL of the extraction solvent  
160 *n*-hexane/dichloromethane (DCM) mixture (4:1, v/v) were added before 1 min of vortexing.  
161 This mixture was then centrifuged for 5 min (2200 g) before the organic layer was transferred  
162 to a clean glass tube. This extraction was repeated before the extracts were evaporated to near  
163 dryness and resolubilised in 0.50 mL *n*-hexane followed by 1 min vortexing.

164

165 **Feather extraction:** To approximately 200 mg feathers, 100  $\mu$ L of IS1, 40  $\mu$ L of IS2, 5 mL of  
166 hydrochloric acid (HCl, 4M) and 5 mL of *n*-hexane/DCM mixture (4:1, v/v) were added before  
167 the samples were incubated at 45 °C overnight. The incubated sample solutions were vortexed  
168 thoroughly for 1 min and the organic layer was retrieved. This liquid-liquid extraction was  
169 repeated with 5 mL of *n*-hexane/DCM mixture (4:1, v/v). Extracts were then evaporated to near  
170 dryness (~200  $\mu$ L) by a gentle nitrogen stream and resolubilised in 0.50 mL *n*-hexane followed  
171 by 1 min vortexing.

172

173 **Preen oil extraction:** Between 13-40 mg of preen oil was transferred to clean glass tubes using  
174 a spatula and the accurate weight was recorded. The spatula was thoroughly cleaned with

175 acetone between samples. Prior to extraction, the samples were spiked with 100  $\mu\text{L}$  IS1 and 20  
176  $\mu\text{L}$  of IS3. Subsequently, 2 mL of *n*-hexane was added to the spiked sample, which was then  
177 vortexed for 1 min.

178

179 Further clean-up and fractionation of all sample extracts were performed according to Eulaers  
180 et al. (2011b) and Poma et al. (2017), with slight modifications. Detailed descriptions of these  
181 modifications are available in the SI. The preen oil samples could not be analysed for PFRs  
182 due to the high lipid content of the oil which made it difficult to use the PFR clean-up procedure  
183 to get a lipid-free extract. After clean-up, all the extracts were concentrated to near dryness  
184 under a gentle nitrogen stream and resolubilised in 100  $\mu\text{L}$  of iso-octane. For each batch of 24  
185 samples, 100  $\mu\text{L}$  of recovery standard (CB 207, 50  $\text{pg}/\mu\text{L}$  in iso-octane/toluene 9:1, v/v) was  
186 added to five of the samples and vortexed for 30 s. Extracts were transferred to injection vials  
187 and analysed by gas chromatography with electron capture negative ionization and mass  
188 spectrometry (GC-ECNI/MS) according to Eulaers et al. (2011b) for legacy POPs and Poma et  
189 al. (2017) for ECs (details in SI).

### 190 **2.2.3. *Per- and polyfluoroalkyl substances***

191 The analysis of PFASs in feathers and plasma was performed at the Norwegian Institute of Air  
192 Research in Tromsø, Norway. The targeted PFASs were one perfluorinated sulfonamide, seven  
193 perfluorinated sulfonates and 11 perfluorinated acids. See Table S4 for the full list of targeted  
194 compounds. The preen oil samples were not analysed for PFASs due to their high lipid content  
195 and small sample volumes. The contents of the internal standard for PFASs (IS4) can be found  
196 in Table S3.

197

198 **Plasma extraction:** Plasma samples were extracted and analysed according to Herzke et al.  
199 (2009). Aliquots of 200 and 300  $\mu\text{L}$  of plasma were thawed and homogenised, then spiked with



200 20  $\mu\text{L}$  of IS4. One mL methanol (MeOH) was added to the samples and the solutions were  
201 mixed by shaking and vortexing for 1 min. The samples were ultrasonicated three times for 10  
202 min, with intermittent vortexing. To enhance phase separation and sedimentation, the samples  
203 were centrifuged for 10 min (1500 g). The supernatant (methanol phase) was then purified in  
204 1.70 mL Eppendorf tubes (VWR, USA) containing 25 mg Supelclean<sup>TM</sup> ENVI-Carb<sup>TM</sup>  
205 graphitised carbon absorbent (Sigma-Aldrich, USA) and 50  $\mu\text{L}$  glacial acetic acid. After  
206 centrifuging for 10 min (1500 g), an exact volume of 0.50 mL supernatant was transferred to  
207 glass vials and added 20  $\mu\text{L}$  of recovery standard solution (3,7-diMe-PFOA, 0.102 ng/ $\mu\text{L}$ ).

208

209 **Feather extraction:** Feather samples were extracted and analysed according to Jaspers et al.  
210 (2013). Pre-cleaned and homogenised feathers were transferred to sterile polypropylene tubes  
211 (VWR, USA). For preen oil removal, feather homogenates were immersed in 20 mL of *n*-  
212 hexane and ultra-sonicated for 10 min. The *n*-hexane was decanted after centrifugation and the  
213 tubes with the homogenates were dried overnight. Previous tests have shown no removal of  
214 PFASs from feathers by *n*-hexane washes (pers. comm. Dorte Herzke). When dry, samples  
215 were spiked with 20  $\mu\text{L}$  of IS4. To resolve the PFASs bound to proteins in the feather  
216 homogenate, we added 2 mL 200 mM NaOH in MeOH. The homogenate was then vortexed  
217 for 1 min and set to soak for 60 min. Then, we added 10 mL of MeOH and the homogenate  
218 was mixed, ultra-sonicated for 3 x 10 min and let to soak overnight. The next day, PFASs were  
219 further extracted from the samples by adding 200  $\mu\text{L}$  of 2M HCl in MeOH. Extracts were then  
220 centrifuged for 5 min at 1500 g, transferred to new polypropylene tubes and evaporated to 1  
221 mL with RapidVap (Labconco, USA). The 1 mL extracts were then cleaned up with carbon  
222 and recovery standard was added similar as to the plasma samples.

223

224 Prior to quantification analysis, extract aliquots of 100  $\mu$ L were transferred to autosampler vials  
225 with insert and an equal amount of 2 mM aqueous ammonium acetate was added. The extracts  
226 were then refrigerated until quantification. Quantification was performed according to Hanssen  
227 et al. (2013), using ultrahigh pressure liquid chromatography and triple–quadrupole mass-  
228 spectrometry (UHPLC-MS/MS). All labelled and internal standards were provided by NILU  
229 (IRMM-427, ID 0119) and all solvents were purchased from Merck (Darmstadt, Germany).

230

#### 231 ***2.2.4 Quality assurance and quality control***

232 Quality assurance of the analytical method was carried out by measurements of procedural  
233 blanks and standard reference material (SRM). For POPs extractions from plasma, the SRM  
234 was human plasma from the AMAP (Arctic Monitoring and Assessment Program)  
235 interlaboratory exercise. For PFASs extractions from plasma, the SRM was a commercially  
236 available human plasma sample (NIST SRM 1957, USA). For POPs extraction from preen oil,  
237 the SRM was whale blubber (NIST, SRM 1945, USA). These SRMs were used to control the  
238 performance of the analytical method for every 10<sup>th</sup> sample, together with a procedural blank.  
239 No SRM was available for feather samples. However, a procedural blank was analysed for  
240 every 10<sup>th</sup> sample and recoveries of internal standards calculated for every sample as well as  
241 for the blanks. For legacy POPs, PFRs, NBFRs and DPs, the limits of quantification (LOQs)  
242 were calculated as three times the standard deviation of the procedural blanks for each  
243 compound and sample type. For PFAS, the LOQs were calculated as three times the limit of  
244 detection (LOD), which again was calculated as the sum of the average of the procedural blanks  
245 and three times the signal-to-noise ratio for each compound and sample type. The LOQs for all  
246 compounds are available in the SI (Table S1, S2 and S4). For analytes that were not detected  
247 in the blanks, LOQs were set to ten times the signal-to-noise ratio of the sample runs.  
248 Recoveries of internal standards can be found in Table S5 and S6. No contamination was

249 observed in the feather blanks. For plasma, only perfluorohexane sulfonate (PFHxS) was  
250 observed in 33 % of the blanks at average concentration of 0.15 ng/mL. No blank corrections  
251 were carried out for any of the investigated compounds. All PFAS samples fulfilled the  
252 requirements for QA/QC except for PFDoA where recoveries were less than 50 % in some  
253 occasions, and lower than 35 % for one sample. Even so, the low standard deviations give good  
254 confidence in the robustness of the applied method.

255

### 256 **2.3. Statistical analyses**

257 Statistical analyses were performed using R version 3.4.2. Descriptive statistics of all the  
258 investigated compounds are available in the SI (Table S7 – S13). Concentrations of the  
259 compounds are expressed in ng/mL wet weight (ww) for plasma, ng/g ww for feathers and ng/g  
260 ww for preen oil. Data were treated in the same way as in most previous studies on bird feathers  
261 and preen oil to allow for direct comparison (i.e. Eulaers et al., 2011a, 2011b, Gómez-Ramírez  
262 et al., 2017). Thus, compounds quantified in over 50 % of samples and detected in both feathers  
263 and plasma samples within each location and each year were included in the statistical analyses.  
264 Data below the LOQ were substituted with LOQ \* detection frequency (DF) within the year  
265 and location of each matrix (Voorspoels et al., 2002). Results from Shapiro-Wilk's test for  
266 normality and visual inspection of normal quantile-quantile plots showed that the  
267 concentrations of the compounds were not normally distributed, some not even after log<sub>e</sub>  
268 transformation. All statistical analyses were therefore performed using non-parametric tests on  
269 untransformed data. Significance levels were set to  $\alpha = 0.05$ .

270

271 Concentration differences between years and locations of  $\Sigma_9$ PCBs,  $\Sigma_3$ OCPs, 2,2',4,4'-  
272 tetrabromodiphenyl ether (BDE 47) and perfluoroundecanoic acid (PFUnA) were investigated  
273 by Kruskal-Wallis analyses for each matrix separately. Significant Kruskal-Wallis analyses

274 were further investigated by Dunn's test of multiple comparisons with Bonferroni correction.  
275 Correlations between concentrations in the three matrices of each selected compound were  
276 investigated by Spearman's rank correlation ( $r_s$ ) for each year and location separately, as well  
277 as combined. Concentrations from both years ( $n = 70$ ) were included for feather and plasma  
278 correlations. For preen oil, only data from 2016 ( $n = 43$ ) could be included. PFUnA  
279 concentrations in feathers and plasma from 2015 and 2016 were not in a monotonic relationship  
280 due to the large differences between the years and thus could not be analysed with Spearman's  
281 rank correlation. To investigate if a larger sample size could create a monotonic relationship,  
282 we included PFUnA concentrations reported in feathers and plasma from white-tailed eagle  
283 nestlings sampled in 2014 in Steigen (data from Gómez-Ramírez et al. 2017). By adding the  
284 latter samples and pooling samples from 2014 – 2016, a monotonic relationship was established  
285 between feather and plasma concentrations, and the relationship was analysed by Spearman's  
286 rank correlation.

287

### 288 **3. Results**

#### 289 **3.1. Detection frequencies and concentrations of legacy POPs and ECs**

290 The compounds that were quantified in over 50 % of both plasma and feather samples from all  
291 white-tailed eagle nestlings within each year and location ( $n = 70$ ) included nine PCBs  
292 ( $\Sigma_9$ PCBs: CB 99, 101, 105, 118, 138, 153, 170, 180 and 187), three OCPs ( $\Sigma_3$ OCPs: *oxy-*  
293 *chlordane* (OxC), *dichlorodiphenyldichloroethylene* (*p,p'*-DDE) and *dichlorodiphenyl-*  
294 *trichloroethane* (*p,p'*-DDT)), BDE 47 and PFUnA. Table 1 presents the median (min-max)  
295 concentrations of these compounds, while the same info for all 54 targeted compounds is listed  
296 in the SI (Table S7 – S13). The concentrations of  $\Sigma_9$ PCBs,  $\Sigma_3$ OCPs, BDE 47 and PFUnA  
297 differed between the three matrices, and the general concentration pattern on a wet weight basis  
298 was preen oil > feathers > plasma. The most abundant compounds (highest concentrations)

299 among the legacy POPs were CB 153, *p,p'*-DDE and BDE 47 in all matrices (Table 1). For  
300 PFASs, linear PFOS was found in the highest concentrations in both plasma (2.3 – 31.9 ng/mL,  
301 mean DF: 100 %) and feathers (< 0.03 – 90.2 ng/g, mean DF: 48.9 %) (Table S12). However,  
302 the average detection frequencies of linear PFOS in feathers over the two years were strongly  
303 influenced by the low detection frequencies in 2015 of only 8 and 29 % in Smøla and Steigen,  
304 respectively. Detection frequencies of all targeted PFASs averaged at 78 % in plasma and 38  
305 % in feathers, with higher concentrations detected in plasma than in feathers. On the contrary,  
306 perfluorooctanesulfonamide (PFOSA) was only detected in feathers, at detection frequencies  
307 between 46 – 100 % in both years and locations. Of the PFRs, the most abundant compounds  
308 were tris(1,3-dichloro-2-propyl) phosphate (TDCIPP) in plasma (< 0.2 – 1.4 ng/mL, mean DF:  
309 17.8 %) and triphenyl phosphate (TPhP) in feathers (< 1.0 – 1229.3 ng/g, mean DF: 94 %)  
310 (Table S11), however at detection frequencies lower than 50 % in feathers and plasma,  
311 respectively. Contrary to PFASs, the concentrations of PFRs detected in feathers exceeded  
312 those in plasma and the average detection frequencies of the targeted PFRs were 13 % in plasma  
313 and 80 % in feathers. The most abundant NBFR was bis(2-ethylhexyl)-3,4,5,6-  
314 tetrabromophthalate (TBPH) in both plasma (0.08 ng/mL) and feathers (< 0.40 – 1.03 ng/g)  
315 (Table S11), but both at low detection frequencies (< 5 % and < 36 %, respectively). No NBFRs  
316 were detected in the preen oil. Of the DPs, the most dominating isomer was *anti*-DP in both  
317 plasma (< 0.002 – 0.03 ng/mL, mean DF: 12 %) and feathers (< 0.10 – 1.14 ng/g, mean DF: 21  
318 %) (Table S11). *Anti*-DP was also detected in one of the preen oil samples, at 0.45 ng/g.

319

### 320 ***3.1.1 Differences between locations and years***

321 We detected significant concentration differences for  $\Sigma_9$ PCBs,  $\Sigma_3$ OCPs, BDE 47 and PFUnA  
322 in plasma ( $\chi^2_{(70,3)} = 9.04 - 51.2$ ,  $p < 0.05$ ) and feathers ( $\chi^2_{(70,3)} = 28.8 - 34.0$ ,  $p < 0.05$ ) between  
323 the two years and between locations (Table S7 – S12). The median concentrations of these

324 contaminant groups were generally higher in feathers and plasma samples from Steigen than  
325 Smøla. For preen oil, the median concentrations were also slightly higher in Steigen than in  
326 Smøla, although not significantly ( $\chi^2_{(43,1)} = 0.3 - 0.7, p > 0.05$ ).

327

## 328 **3.2. Correlations between matrices**

### 329 **3.2.1 Plasma and feather correlations**

330 Strong and significant positive correlations between plasma and feather concentrations were  
331 found for all PCBs, OCPs (except for *p,p'*-DDT) and BDE 47 ( $r_S: 0.33 - 0.95, p < 0.02$ ), when  
332 both years and locations were combined (Table 2, Figure 1). When years and locations were  
333 investigated separately, we detected significant positive correlations between plasma and  
334 feathers for all compounds ( $r_S = 0.43 - 0.93, p < 0.05$ ), except for CB 101, 105 and 180.  
335 Contrary to the POPs, a correlation analysis of PFUnA concentrations in feathers and plasma  
336 on all samples combined was not possible in the current study, as the relationship was non-  
337 monotonic. However, when the years and locations were analysed separately, positive and  
338 significant correlations within Steigen were detected for 2015, as well as for 2016 ( $r_S = 0.69$   
339 and 0.56, respectively,  $p < 0.01$ ). A study from Gómez-Ramírez et al. (2017) has investigated  
340 PFASs in plasma and feathers from white-tailed eagle nestlings from Steigen, sampled in 2014.  
341 Since data from Gómez-Ramírez et al. (2017) were produced in the same lab, using the same  
342 methodology, we combined the raw data from their study with our data and performed new  
343 statistical analysis (since a monotonic relationship was achieved). However, no correlation was  
344 detected between PFUnA concentrations in plasma and feathers on the combined data ( $r_S =$   
345  $0.001, p = 0.99$ ; Figure 1). Figure 1 illustrates a highly scattered distribution, indicating that  
346 the concentrations of PFUnA in these two matrices are highly variable and poorly correlated,  
347 both within locations and years.

348

349 **Table 1:** Summary statistics [median (min – max)] of contaminants quantified in over 50 % of plasma and body  
 350 feathers samples within each year and location in white-tailed eagle nestlings from Smøla and Steigen (Norway).  
 351 Concentrations of other PFASs, PFRs and NBRs were below LOQ in > 50% of the samples and can be found in  
 352 Supplementary information. The preen oil concentrations (ng/g ww) were only available from 2016. Units are  
 353 ng/ml ww for plasma and ng/g ww for feathers. Samples not available for analyses are marked with “n.a”.

	Smøla			Steigen		
		2015 (n = 13)	2016 (n = 22)	2015 (n = 14)	2016 (n = 21)	
	Matrix	median (min - max)	median (min - max)	median (min - max)	median (min - max)	
<b>CB 99</b>	Plasma	0.16 (0.08 – 0.59)	0.18 (0.06 – 1.47)	0.50 (0.18 – 4.61)	0.23 (0.06 – 0.98)	
	Feathers	1.10 (0.18 – 3.85)	0.92 (0.41 – 7.89)	7.78 (2.71 – 31.05)	1.08 (0.21 – 3.61)	
	Preen oil	n.a	28.00 (12.34 – 198.68)	n.a	33.59 (0.95 – 161.58)	
<b>CB 101</b>	Plasma	0.21 (0.09 – 0.31)	0.14 (0.01 – 0.56)	0.16 (0.07 – 0.56)	0.12 (0.02 – 0.25)	
	Feathers	0.72 (0.18 – 1.82)	0.55 (0.31 – 1.57)	1.50 (0.80 – 1.81)	0.44 (0.19 – 0.99)	
	Preen oil	n.a	16.68 (9.99 – 49.55)	n.a	19.86 (0.95 – 32.62)	
<b>CB 105</b>	Plasma	0.08 (0.04 – 0.30)	0.11 (0.04 – 0.79)	0.26 (0.10 – 2.59)	0.14 (0.04 – 0.66)	
	Feathers	0.23 (0.11 – 0.57)	0.24 (0.11 – 0.94)	1.37 (0.52 – 3.85)	0.31 (0.12 – 1.05)	
	Preen oil	n.a	17.94 (7.99 – 91.04)	n.a	22.33 (7.99 – 91.04)	
<b>CB 118</b>	Plasma	0.23 (0.11 – 0.81)	0.41 (0.17 – 2.92)	0.70 (0.28 – 7.30)	0.50 (0.14 – 2.30)	
	Feathers	0.72 (0.23 – 2.1)	0.90 (0.45 – 4.12)	5.12 (2.18 – 15.03)	1.07 (0.4 – 3.37)	
	Preen oil	n.a	51.84 (19.68 – 240.09)	n.a	66.46 (23.95 – 289.96)	
<b>CB 138</b>	Plasma	0.27 (0.11 – 1.25)	1.10 (0.40 – 10.55)	0.66 (0.29 – 5.63)	1.26 (0.28 – 8.88)	
	Feathers	0.42 (0.17 – 1.51)	1.86 (0.75 – 11.78)	1.34 (0.60 – 5.89)	2.64 (0.62 – 6.26)	
	Preen oil	n.a	129.63 (39.37 – 720.65)	n.a	168.07 (40.26 – 602.9)	
<b>CB 153</b>	Plasma	0.74 (0.21 – 3.06)	1.44 (0.55 – 9.48)	2.05 (1.12 – 26.27)	1.75 (0.43 – 10.16)	
	Feathers	1.77 (0.63 – 6.77)	3.13 (1.22 – 17.07)	12.86 (5.48 – 38.64)	4.15 (0.92 – 9.73)	
	Preen oil	n.a	259.08 (81.81 – 1420.57)	n.a	327.62 (94.03 – 1164.43)	
<b>CB 170</b>	Plasma	0.07 (0.02 – 0.36)	0.22 (0.07 – 1.30)	0.18 (0.06 – 2.16)	0.23 (0.07 – 1.98)	
	Feathers	0.14 (0.07 – 0.48)	0.34 (0.13 – 1.45)	0.52 (0.27 – 1.73)	0.42 (0.11 – 1.23)	
	Preen oil	n.a	37.74 (10.16 – 180.79)	n.a	50.51 (12.85 – 184.32)	
<b>CB 180</b>	Plasma	0.17 (0.04 – 0.84)	0.70 (0.20 – 3.55)	0.45 (0.13 – 5.29)	0.65 (0.19 – 5.89)	
	Feathers	0.23 (0.10 – 1.03)	0.74 (0.26 – 3.06)	0.96 (0.56 – 3.30)	0.92 (0.24 – 2.73)	
	Preen oil	n.a	117.86 (29.89 – 489.81)	n.a	139.58 (40.9 – 579.26)	
<b>CB 187</b>	Plasma	0.11 (0.03 – 0.43)	0.32 (0.13 – 2.55)	0.22 (0.07 – 1.81)	0.36 (0.10 – 1.95)	
	Feathers	0.17 (0.08 – 0.78)	0.48 (0.17 – 3.21)	0.76 (0.26 – 1.89)	0.45 (0.14 – 1.08)	
	Preen oil	n.a	55.94 (16.74 – 434.19)	n.a	59.11 (20.75 – 186.85)	
<b>Σ<sub>9</sub> PCB</b>	Plasma	1.90 (0.78 – 7.97)	4.44 (1.70 – 32.37)	4.87 (2.81 – 56.17)	5.34 (1.44 – 32.98)	
	Feathers	5.56 (1.86 – 18.39)	9.20 (3.86 – 51.10)	34.10 (14.89 – 101.29)	11.94 (2.96 – 27.61)	
	Preen oil	n.a	723.56 (231.26- 3804.89)	n.a	863.79 (269.37 – 3196.99)	
<b>OxC</b>	Plasma	0.04 (0.02 – 0.14)	0.08 (0.02 – 0.53)	0.24 (0.05 – 2.16)	0.13 (0.04 – 0.6)	
	Feathers	0.11 (0.07 – 0.32)	0.26 (0.13 – 1.28)	1.36 (0.36 – 6.89)	0.48 (0.08 – 1.47)	
	Preen oil	n.a	9.52 (3.36 – 30.37)	n.a	13.54 (4.80 – 53.73)	
<b>p,p'-DDE</b>	Plasma	1.21 (0.56 – 5.23)	1.20 (0.56 – 9.47)	3.95 (2.18 – 47.61)	1.45 (0.48 – 8.64)	
	Feathers	4.96 (2.35 – 22.1)	3.23 (1.48 – 17.03)	26.59 (12.38 – 94.38)	3.74 (1.14 – 8.81)	
	Preen oil	n.a	298.56 (170.99 -1447.32)	n.a	364.88 (185.38 – 932.96)	
<b>p,p'-DDT</b>	Plasma	0.20 (0.08 – 0.30)	0.15 (0.06 – 0.63)	0.12 (0.02 – 0.31)	0.27 (0.02 – 0.38)	
	Feathers	0.50 (0.14 – 1.14)	0.24 (0.15 – 1.03)	0.89 (0.17 – 1.92)	0.24 (0.14 – 0.47)	
	Preen oil	n.a	7.07 (3.85 – 15.09)	n.a	7.65 (5.06 – 46.63)	
<b>Σ<sub>3</sub> OCP</b>	Plasma	1.54 (0.67 – 5.62)	1.47 (0.64 – 10.64)	4.37 (2.40 – 49.83)	3.48 (1.31 – 12.96)	
	Feathers	5.68 (2.95 – 23.56)	4.12 (2.03 – 18.63)	28.62 (13.98 – 103.19)	4.36 (1.37 – 10.08)	
	Preen oil	n.a	319.63 (180.46 – 1491.16)	n.a	384.06 (198.87 – 1024.27)	
<b>BDE 47</b>	Plasma	0.06 (0.03 – 0.28)	0.08 (0.01 – 0.81)	0.19 (0.06 – 1.82)	0.09 (0.01 – 0.36)	
	Feathers	0.56 (0.28 – 1.58)	0.30 (0.16 – 2.43)	1.73 (0.63 – 3.59)	0.45 (0.14 – 1.06)	
	Preen oil	n.a	9.19 (4.25 – 96.35)	n.a	10.47 (4.45 – 60.28)	
<b>PFUnA</b>	Plasma	3.59 (2.43 – 4.36)	1.15 (0.68 – 2.05)	3.36 (2.3 – 5.08)	1.40 (0.94 – 2.15)	
	Feathers	0.15 (0.05 – 0.6)	0.58 (0.07 – 0.95)	0.38 (0.13 – 0.81)	0.82 (0.26 – 1.07)*	
	Preen oil	n.a	n.a	n.a	n.a	

\* Feathers for PFUnA Steigen 2016: n = 19

355

356

357 **3.2.2 Plasma and preen oil correlations**

358 Significant positive correlations were found for all compounds that were detected > 50 % in  
 359 preen oil and plasma sampled in 2016, when both locations were combined ( $r_s > 0.35$ ,  $p < 0.02$ ,  
 360 Table 2). When samples from Steigen and Smøla were analysed for correlations separately,  
 361 significant positive correlations were found for all compounds except CB 101 from Smøla ( $r_s$   
 362 = 0.21,  $p < 0.35$ ). Unfortunately, the preen oil samples could not be analysed for PFRs and  
 363 PFAS due to their high lipid content and small sample volume.

364

365 **3.2.3 Feathers and preen oil correlations**

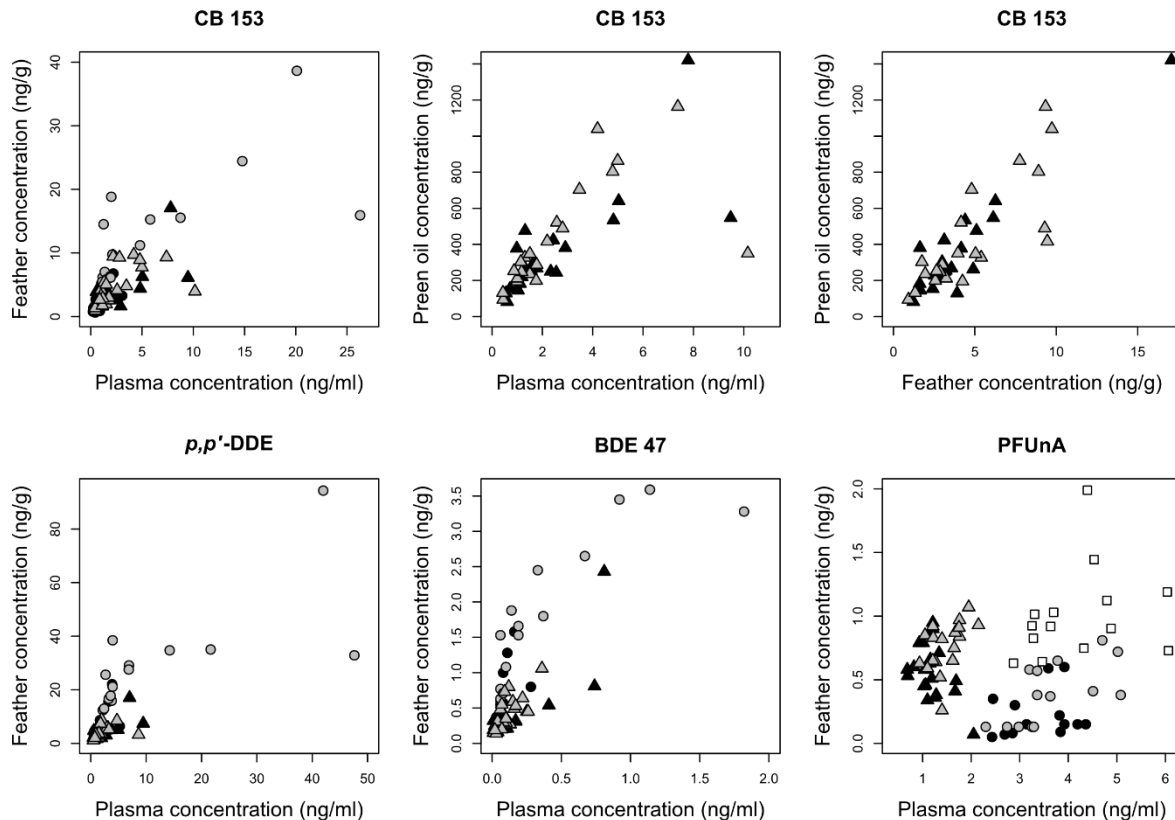
366 Similar to plasma and feathers, strong and significant correlations were found between feathers  
 367 and preen oil concentrations from 2016 for all compounds detected > 50 % ( $r_s > 0.46$ ,  $p < 0.01$ ),  
 368 except for *p,p'*-DDT (Table 2), when both locations were combined. When samples from  
 369 Smøla and Steigen were analysed for correlations separately, the relationships between feathers  
 370 and preen oil from Smøla were weak and not significant for CB 101, *p,p'*-DDT and BDE 47  
 371 ( $r_s = 0.18 - 0.25$ ,  $p > 0.3$ ).

372 **Table 2:** Spearman's correlation coefficients ( $r_s$ ) and significance values ( $p$ ) between contaminant concentrations  
 373 in blood plasma, body feathers and preen oil from white-tailed eagle nestlings from Smøla and Steigen (Norway).  
 374 Correlation coefficients are not available for PFUnA due to the non-monotonic relationship between feathers and  
 375 plasma. Correlations for compounds not present in both matrices could not be calculated ("n.a."). Significant  $p$ -  
 376 values are marked with \*.

	Blood plasma ~ body feathers		Blood plasma ~ preen oil		Body feathers ~ preen oil	
	$n = 70$		$n = 43$		$n = 43$	
	$r_s$	$p$ -value	$r_s$	$p$ -value	$r_s$	$p$ -value
<b>CB 99</b>	0.73	< 0.001*	0.78	< 0.001*	0.81	< 0.001*
<b>CB 101</b>	0.33	0.006*	0.35	0.02*	0.46	0.002*
<b>CB 105</b>	0.76	< 0.001*	0.86	< 0.001*	0.74	< 0.001*
<b>CB 118</b>	0.74	< 0.001*	0.86	< 0.001*	0.77	< 0.001*
<b>CB 138</b>	0.78	< 0.001*	0.83	< 0.001*	0.80	< 0.001*
<b>CB 153</b>	0.72	< 0.001*	0.83	< 0.001*	0.79	< 0.001*
<b>CB 170</b>	0.73	< 0.001*	0.77	< 0.001*	0.73	< 0.001*
<b>CB 180</b>	0.72	< 0.001*	0.76	< 0.001*	0.73	< 0.001*
<b>CB 187</b>	0.72	< 0.001*	0.78	< 0.001*	0.79	< 0.001*
<b>OxC</b>	0.83	< 0.001*	0.95	< 0.001*	0.75	< 0.001*
<b><i>p,p'</i>-DDE</b>	0.73	< 0.001*	0.84	< 0.001*	0.71	< 0.001*
<b><i>p,p'</i>-DDT</b>	0.12	0.32	0.52	< 0.001*	0.24	0.116
<b>BDE 47</b>	0.67	< 0.001*	0.86	< 0.001*	0.73	< 0.001*
<b>PFUnA</b>	-	-	n.a.	n.a.	n.a.	n.a.

377





378

379 **Figure 1:** Correlation plots of concentrations in plasma (ng/ml ww), feathers (ng/g ww) and preen oil (ng/g ww)  
 380 of CB 153, *p,p'*-DDE, BDE 47 and PFUnA from white-tailed eagle nestlings from Steigen in grey and from Smøla  
 381 in black. Samples from 2015 are in circles (○), while samples from 2016 are in triangles (Δ). Preen oil was only  
 382 sampled in 2016. Samples from Gómez-Ramírez et al. 2017 are in open squares (□).  
 383

#### 384 **4. Discussion**

##### 385 **4.1 Detection frequencies and concentrations of legacy POPs and ECs**

386 We expected to find that the concentrations and detection frequencies of legacy POPs and ECs  
 387 in plasma would also be reflected in the feathers and preen oil. This was true for legacy POPs,  
 388 as we found high detection frequencies for PCBs, OCPs and BDE 47 in plasma, feathers and  
 389 preen oil. The concentration profile with high concentrations in the preen oil was expected due  
 390 to the high lipid content of the oil and the lipophilic nature of these compounds (Eulaers et al.,  
 391 2011b). The concentrations of the main contaminant contributors (CB 153, *p,p'*-DDE, BDE 47  
 392 and PFUnA, Table 1) in plasma, feathers and preen oil were slightly lower in the current study  
 393 than previously reported in white-tailed eagle nestlings from Norway (Eulaers et al., 2011a,  
 394 2011b, 2013, 2014; Gómez-Ramírez et al., 2017). The samples from the previous studies were

395 collected in 2008, 2009, 2011 and 2014, and some also at other locations than Smøla and  
396 Steigen. Hence, some of this variation may be due to temporal, spatial, biological or dietary  
397 differences (Eulaers et al., 2013, Løseth et al., in preparation).

398

399 The detection frequencies of the analysed ECs differed between plasma and feathers and may  
400 suggest different exposure routes or different toxicokinetics in the two matrices. The higher  
401 detection frequencies of PFASs in plasma than in feathers, is contrary to a study where PFOS  
402 concentrations were compared between keratinous tissues (hair and nails) and serum in humans  
403 (Li et al., 2013). Although the detection frequencies of PFASs were low in feathers in the  
404 current study, the concentrations correspond to those reported in a previous study on feathers  
405 from white-tailed eagle nestlings (Gómez-Ramírez et al., 2017).

406

407 The higher concentrations of PFOSA, PFRs, NBRs and DPs in feathers than in plasma,  
408 suggest that the feathers may not only reflect the internal contamination burden. Some of the  
409 concentrations may potentially originate from external contamination (Eulaers et al., 2014;  
410 Jaspers et al., 2008). Possible sources of external contamination can come from outdoor  
411 environments, field accommodations or other indoor environments (Cequier et al., 2014; Green  
412 et al., 2008, Möller et al., 2011, Tollbäck et al. 2006). The field accommodation at Smøla 2015  
413 was a newly built house and even though the feathers were thoroughly rinsed before  
414 contaminant extraction, their PFR profile of TPhP > tris(2-chloroisopropyl) phosphate (TCPP)  
415 > tris(chloroethyl) phosphate (TCEP) show similarities to profiles reported in indoor air and  
416 dust (Cequier et al., 2014; Green et al., 2008, Tollbäck et al. 2006). The general PFR profile  
417 detected in feathers from the other location and years was TCPP > TCEP > TPhP. This same  
418 profile has been reported in a study of atmospheric air from the North Sea (Möller et al., 2011)  
419 and at a remote Arctic location (Green et al., 2008). *Anti-DP* and TBPH were also detected in

420 higher concentrations in feathers than in plasma. These compounds have also been detected in  
421 indoor air and dust from Norway (Cequier et al., 2014). The similarity of PFR profiles and the  
422 occurrence of *anti*-DP and TBPH in feathers and in air further suggests that feathers may act  
423 as air- and dust samplers. Some of the detected concentrations may therefore originate from  
424 external contamination, which may not have been removed by the washing procedure.

425

426 The lower detection frequencies and concentrations of PFRs and TBPH in plasma, compared  
427 to feathers, may also result from rapid metabolism and excretion of these compounds from  
428 internal tissues as reported in other studies (Barr et al., 2012; Briels et al., 2018; Covaci et al.,  
429 2011; Hou et al., 2016). The PFRs and NBFRs detected in the current study have previously  
430 been detected in white-tailed eagle samples, primarily in feathers, from Trøndelag and Troms,  
431 Norway (Eulaers et al., 2014). As in Eulaers et al. (2014), our study found low concentrations  
432 and detection frequencies of PFRs and NBFRs in plasma, further suggesting high excretion  
433 rates, low absorption or low exposure of these compounds. Dechloranes, on the contrary, may  
434 not biotransform (Briels et al., 2018) and can accumulate in biota (Feo et al., 2012). To our  
435 knowledge, this is the first study detecting DPs in feathers and preen oil, and further studies  
436 are therefore needed to investigate if the concentrations of PFRs, NBFRs and DPs in feathers  
437 are of external and/or internal origin. Although the present study documents that white-tailed  
438 eagle nestlings are exposed to PFRs, NBFRs and DPs, we did not investigate the possible  
439 correlations between these compounds in plasma, feathers and preen oil due to the < 50 %  
440 detection frequencies in some of these matrices.

441

#### 442 **4.2. Correlation between matrices**

443 In general, the relatively low concentrations of PCBs, OCPs, BDE 47 and PFUnA quantified  
444 in plasma reflected the nestlings' recent exposure through diet (Henriksen et al., 1998) and

445 remains from maternal transfer to the eggs (Bourgeon et al., 2013). The concentrations of these  
446 compounds quantified in the feathers were incorporated into the feathers some weeks prior to  
447 the sampling, and the concentrations therefore reflected blood concentrations at that time  
448 (García-Fernández et al., 2013; Jaspers et al., 2006). Feather concentrations of POPs and ECs  
449 can also be affected by preening activity and external contamination from air and dust. The  
450 preen oil is lipid rich and may function as a passive excretion route for lipophilic compounds  
451 onto the feathers (Eulaers et al., 2011b; Jaspers et al., 2008). However, in nestlings this activity  
452 is considered to be of minor influence on feather concentrations (Eulaers et al., 2011b; Jaspers  
453 et al., 2011). Although the concentrations of the quantified POPs and ECs seem to be higher in  
454 feathers than in blood (on a ww basis), the pattern may vary depending on the structure and  
455 toxicokinetics of the compound, as we generally found higher concentrations of PFASs in  
456 plasma than in feathers (Table 1).

457

458 The high detection frequencies of legacy POPs in plasma and feathers also resulted in  
459 significant correlations between these matrices. The strong correlations of PCBs between  
460 plasma and feathers are in accordance with previous studies on white-tailed eagle nestlings  
461 (Eulaers et al., 2011a, 2011b). The correlations also correspond with results from earlier studies  
462 on PCBs correlations in internal tissues and feathers (Dauwe et al., 2005; Van den Steen et al.,  
463 2007). Nevertheless, when the two locations were analysed separately, no significant  
464 correlations were detected for CB 101, 105 and 180 between plasma and feathers. This lack of  
465 correlation corresponds to previous studies with small sample sizes (Eulaers et al., 2011a,  
466 2011b), and may reflect temporal, spatial or biological variation. The differences between the  
467 two locations, regarding the mentioned variables, will be further investigated in another study  
468 (Løseth et al., in preparation).

469

470 Our significant correlations for POP concentrations in plasma and feathers are contrary to a  
471 study on adult black-legged kittiwakes (*Rissa tridactyla*) from Svalbard, where the authors  
472 investigated several PCBs, OCPs and PFRs in plasma and feathers (Svendsen et al., 2018). In  
473 that study, a significant positive correlation was only identified for CB 153 (Svendesen et al.,  
474 2018). The authors argued that the absence of correlations between plasma and feathers  
475 concentration may be linked to the migratory behaviour of the adult kittiwakes. The sampled  
476 primary feathers were grown when the birds were at their wintering areas, and plasma and  
477 feathers were collected during summer (Svendsen et al., 2018). In our study, the nestlings were  
478 sampled when they were stationary in their nests, and concentrations detected in their growing  
479 feathers are therefore more likely to correlate with plasma concentrations.

480

481 The high detection frequencies and concentrations of legacy POPs in the preen oil also resulted  
482 in significant correlations between plasma, feathers and preen oil. These significant  
483 correlations were in accordance with a previous study on white-tailed eagle nestlings (Eulaers  
484 et al., 2011b). This is the fourth study, to our knowledge, where plasma and preen oil  
485 concentrations of legacy POPs have been compared (Eulaers et al., 2011b; Van den Brink,  
486 1997; Yamashita et al., 2007). Therefore, our study further adds to the evidence of preen oil as  
487 a suitable matrix for biomonitoring of legacy POPs as it strongly reflects internal  
488 concentrations.

489

490 Of the analysed ECs, PFUnA was the only compound which could be investigated for  
491 correlations between plasma and feathers. The two significant correlations detected between  
492 plasma and feathers for PFUnA in Steigen 2015 and 2016 contrasts with reports from a study  
493 on white-tailed eagle nestlings at the same location in 2014 (Gomez-Ramirez et al., 2017).  
494 Their study detected significant correlations between plasma and feathers for other PFASs, but

495 not for PFUnA (Gomez-Ramirez et al., 2017). It should, however, be noted that in the present  
496 study, no correlations were detected between plasma and feather concentrations of PFUnA in  
497 the Smøla population, in either 2015 or 2016. Also, no significant correlation was detected  
498 when our data were combined with data from Gomez-Ramirez et al. (2017). The variability of  
499 the plasma and feather correlation suggests that feathers may not be a suitable matrix for  
500 investigating internal concentrations of PFUnA (Gómez-Ramírez et al., 2017). The large  
501 variation observed between years and locations in detection frequencies of several PFASs in  
502 feathers also leads us to question the general suitability of feathers for monitoring internal  
503 PFASs concentrations. As there is little knowledge on the deposition of PFASs into feathers,  
504 we suggest prioritising the use of plasma samples to investigate internal PFASs concentrations  
505 in birds.

506

## 507 **5. Conclusions**

508 This is the first study to present a wide investigation of feathers and preen oil, in relation to  
509 plasma, for monitoring of both legacy and emerging compounds in white-tailed eagle nestlings  
510 from Norway. Our results propose both feathers and preen oil as suitable matrices for legacy  
511 POP analyses as the concentrations were significantly and positively correlated with plasma  
512 concentrations. This was also the first study to investigate non-destructive sampling methods  
513 from one species at different locations and years. Despite inter-annual and spatial variation of  
514 POPs, our large sample size allowed strong and robust statistical analyses providing further  
515 support for the strong and significant correlations between the three matrices for legacy POPs  
516 found in previous studies. For PFASs on the other hand, the inter-annual and spatial variation  
517 as well as the low detection frequencies in feathers compared to plasma resulted in poor and  
518 non-significant correlations between feathers and plasma. Because of the generally high  
519 detection frequencies of PFASs in plasma despite inter-annual and spatial variation, we suggest

520 prioritising the use of plasma for PFAS analyses. Correlations could not be investigated for  
521 PFRs, NBFRs and DPs due to low detection frequencies in plasma. The higher detection  
522 frequencies and concentrations of these emerging contaminants in feathers compared to plasma  
523 may suggest that feathers are prone to external contamination and/or that these compounds are  
524 rapidly metabolised and excreted. Further studies are needed to investigate if PFRs, NBFRs  
525 and DPs detected in feathers are from external or internal origin.

526

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535 additional data on PFUnA concentrations in white-tailed eagle nestlings from Steigen (2014).  
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537 Flagship (the Raptor project) at the Fram Centre in Tromsø.

538

## 539 **Supplementary information:**

540 Table S1: List of targeted organochlorinated compounds for analyses

541 Table S2: List of targeted flame retardant compounds for analyses

542 Table S3: Contents of internal standards

543 Table S4: List of targeted per- and polyfluoroalkyl substances for analyses

544 Table S5: Recoveries of internal standards in plasma

545 Table S6: Recoveries of internal standards in feathers and preen oil

546 Table S7-S13: Detection frequencies and descriptive statistics for all analysed compounds in

547 plasma, feathers and preen oil

548 Additional analytical details of POPs and PFR analyses

549



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748 White-tailed eagle (*Haliaeetus albicilla*) feathers from Norway are suitable for  
749 monitoring of legacy, but not emerging contaminants

750 Mari E. Løseth<sup>a\*</sup>, Nathalie Briels<sup>a</sup>, Jørgen Flo<sup>a</sup>, Govindan Malarvannan<sup>b</sup>, Giulia Poma<sup>b</sup>, Adrian Covaci<sup>b</sup>,  
751 Dorte Herzke<sup>c</sup>, Torgeir Nygård<sup>d</sup>, Jan O. Bustnes<sup>e</sup>, Bjørn M. Jenssen<sup>a</sup> and Veerle L. B. Jaspers<sup>a</sup>

752

753

754 **Affiliations:**

755 <sup>a</sup>Department of Biology, Norwegian University of Science and Technology (NTNU),  
756 Høgskoleringen 5, 7491 Trondheim, Norway

757 <sup>b</sup>Toxicological Center, University of Antwerp, Universiteitsplein 1, 2610 Wilrijk, Belgium

758 <sup>c</sup>Norwegian Institute for Air Research (NILU), FRAM- High North Research Centre on  
759 Climate and the Environment, 9007 Tromsø, Norway

760 <sup>d</sup>Norwegian Institute for Nature Research (NINA), Høgskoleringen 9, 7034 Trondheim,  
761 Norway

762 <sup>e</sup>Norwegian Institute for Nature Research (NINA), FRAM- High North Research Centre on  
763 Climate and the Environment, 9007 Tromsø, Norway

764

765 **\*Corresponding author:**

766 Mari Engvig Løseth: mari.loseth@ntnu.no

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778 plasma, feathers and preen oil

779 Additional analytical details of POPs and PFR analyses

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782 **Table S3:** Targeted chlorinated compounds analysed in plasma, feather and preen oil samples from white-tailed  
783 eagle nestlings sampled at Steigen and Smøla (Norway) in 2015 and 2016. Only samples from 2016 were analyzed  
784 for *p,p'*-DDD. PCB congeners are numbered by the IUPAC system (International Union of Pure and Applied  
785 Chemistry). Limit of quantification (LOQ) for the compounds are the same for 2015 and 2016 and are presented  
786 as ng/mL for plasma and ng/g for feathers and preen oil.

Organochlorinated compounds					
Group	Abbreviations	Compounds	LOQ plasma	LOQ feathers	LOQ preen oil
Organo-Chlorinated Pesticides (OCPs)	OxC	<i>oxy</i> -chlordane	0.01	0.10	1.0
	TN	<i>trans</i> -nonachlor	0.01	0.10	1.0
	CN	<i>cis</i> -nonachlor	0.01	0.10	1.0
	HCB	hexachlorobenzene	0.01	0.10	1.0
	<i>a</i> -HCH	1 $\alpha$ ,2 $\alpha$ ,3 $\beta$ ,4 $\alpha$ ,5 $\beta$ ,6 $\beta$ -hexachlorocyclohexane	0.01	0.10	1.0
	<i>b</i> -HCH	1 $\alpha$ ,2 $\beta$ ,3 $\alpha$ ,4 $\beta$ ,5 $\alpha$ ,6 $\beta$ -hexachlorocyclohexane	0.01	0.10	1.0
	<i>g</i> -HCH	1 $\alpha$ ,2 $\alpha$ ,3 $\beta$ ,4 $\alpha$ ,5 $\alpha$ ,6 $\beta$ -hexachlorocyclohexane	0.02	0.10	1.0
	<i>p,p'</i> -DDT	<i>p,p'</i> -dichloro- $\alpha,\alpha$ -diphenyl- $\beta,\beta,\beta$ -trichloroethane	0.02	0.20	2.0
	<i>p,p'</i> -DDE	<i>p,p'</i> -dichloro-diphenyl-dichloroethylene	0.02	0.20	2.0
<i>p,p'</i> -DDD	<i>p,p'</i> -dichloro-diphenyl-dichloroethane	0.02	0.20	2.0	
Polychlorinated Biphenyls (PCBs)	CB 28	2,4,4'-trichlorobiphenyl	0.05	0.30	2.0
	CB 49	2,2',4,5'-tetrachlorobiphenyl	0.05	0.30	2.0
	CB 52	2,2',5,5'-tetrachlorobiphenyl	0.05	0.30	2.0
	CB 74	2,4,4',5-tetrachlorobiphenyl	0.05	0.30	2.0
	CB 95	2,2',3,5',6-pentachlorobiphenyl	0.02	0.20	1.0
	CB 99	2,2',4,4',5-pentachlorobiphenyl	0.02	0.20	1.0
	CB 101	2,2',4,5,5'-pentachlorobiphenyl	0.02	0.20	1.0
	CB 105	2,3,3',4,4'-pentachlorobiphenyl	0.01	0.10	1.0
	CB 110	2,3,3',4',6-pentachlorobiphenyl	0.01	0.10	1.0
	CB 118	2,3',4,4',5-pentachlorobiphenyl	0.01	0.10	1.0
	CB 138	2,2',3,4,4',5'-hexachlorobiphenyl	0.01	0.10	1.0
	CB 149	2,2',3,4',5',6-hexachlorobiphenyl	0.01	0.10	1.0
	CB 153	2,2',4,4',5,5'-hexachlorobiphenyl	0.01	0.10	1.0
	CB 156	2,3,3',4,4',5-hexachlorobiphenyl	0.01	0.10	1.0
	CB 170	2,2',3,3',4,4',5-heptachlorobiphenyl	0.01	0.10	1.0
	CB 171	2,2',3,3',4,4',6-heptachlorobiphenyl	0.01	0.10	1.0
	CB 177	2,2',3,3',4,5',6'-heptachlorobiphenyl	0.01	0.10	1.0
	CB 180	2,2',3,4,4',5,5'-heptachlorobiphenyl	0.01	0.10	1.0
	CB 183	2,2',3,4,4',5',6-heptachlorobiphenyl	0.01	0.10	1.0
	CB 187	2,2',3,4',5,5',6-heptachlorobiphenyl	0.01	0.10	1.0
	CB 194	2,2',3,3',4,4',5,5'-octachlorobiphenyl	0.01	0.10	1.0
	CB 206	2,2',3,3',4,4',5,5',6-nonachlorobiphenyl	0.01	0.10	1.0
	CB 209	Decachlorobiphenyl	0.01	0.10	1.0

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790 **Table S4:** Targeted flame retardants analysed in plasma, feather and preen oil samples from white-tailed eagle  
791 nestlings, sampled at Steigen and Smøla (Norway) in 2015 and 2016. PBDE congeners are numbered by the  
792 IUPAC system (International Union of Pure and Applied Chemistry). Only samples from 2016 were analysed for  
793 2'-MeO-BDE 68 and 6'-MeO-BDE 47. Limit of quantification (LOQ) for the compounds are the same for 2015  
794 and 2016 and are presented as ng/mL for plasma and ng/g for feathers and preen oil. Compounds not targeted  
795 (analysed) are marked with "n.a".

<b>Flame retardants</b>					
<b>Group</b>	<b>Abbreviations</b>	<b>Compounds</b>	<b>LOQ plasma</b>	<b>LOQ feathers</b>	<b>LOQ preen oil</b>
Polybrominated diphenyl ethers (PBDEs)	BDE 28	2',4,4'-tribromodiphenyl ether	0.002	0.10	0.4
	BDE 47	2,2',4,4'-tetrabromodiphenyl ether	0.002	0.10	0.4
	BDE 99	2,2',4,4',5'-pentabromodiphenyl ether	0.002	0.10	0.4
	BDE 100	2,2',4,4',6'-pentabromodiphenyl ether	0.002	0.10	0.4
	BDE 153	2,2',4,4',5,5'-hexabromobiphenyl ether	0.002	0.10	0.8
	BDE 154	2,2',4,4',5,6'-hexabromobiphenyl ether	0.004	0.10	0.8
	BDE 183	2,2',3',4,4',5',6'-heptabromodiphenyl ether	0.004	0.20	0.8
2016	2'-MeO-BDE 68	1,5-Dibromo-3-(2,4-dibromophenoxy)-2- methoxybenzene			n.a
2016	6'-MeO-BDE 47	1,5-Dibromo-2-(2,4-dibromophenoxy)-3- methoxybenzene			n.a
Dechlorane plus isomers (DPs)	Syn-DP	Syn-Dechlorane plus	0.002	0.10	0.4
	Anti-DP	Anti-Dechlorane plus	0.002	0.10	0.4
Novel brominated flame retardants (NBFRs)	TBB	2-ethylhexyl-2,3,4,5-tetrabromobenzoate	0.010	0.20	1.6
	TBPH	bis(2-ethylhexyl)-3,4,5,6- tetrabromophthalate	0.020	0.40	2.0
	BTBPE	1,2 Bis(2,4,6-tribromophenoxy)ethane	0.005	0.10	0.4
Phosphate flame retardants (PFRs)	TCEP	tris(chloroethyl) phosphate	0.100	1.00	n.a
	TBOEP	tris(2-butoxyethyl) phosphate	0.400	4.00	n.a
	EHDPP	2-ethylhexyl diphenyl phosphate	0.100	1.00	n.a
	TPhP	triphenyl phosphate	0.100	1.00	n.a
	TCPP	tris(2-chloroisopropyl) phosphate	0.100	1.00	n.a
	TCIPP	tris(1-chloro-2-propyl) phosphate	0.100	1.00	n.a
	TDCIPP	tris(1,3-dichloro-2-propyl)phosphate	0.200	1.00	n.a

796

797

798 **Table S5:** Compounds and their concentrations in internal standards used for extraction of targeted PCBs, PBDEs,  
 799 OCPs, NBFRs, PFRs, DPs and PFASs.

<b>Internal standard</b>	<b>Concentrations and compounds</b>
IS1 (POPs)	200 pg/μL PCB 143 25 pg/μL BDE 77 25 pg/μL ε-hexachlorocyclohexane (ε-HCH)
IS2 (ECs)	200 pg/μL <sup>13</sup> C-bis(2-ethylhexyl)-3,4,5,6-tetrabromophthalate (TBPH) 50 pg/μL <sup>13</sup> C-syn-dechlorane plus (DP) 50 pg/μL <sup>13</sup> C-anti-DP 1 ng/μL triphenyl phosphate (TPHP-d15) 1 ng/μL tris(chloroethyl) phosphate (TCEP-d12) 1 ng/μL tris-(1,3-dichloro-2-propyl) phosphate (TDCIPP-d15) 1 ng/μL triamyl phosphate (TAP) 2 ng/μL tri-(2-butoxyethyl) phosphate (TBOEP-d6)
IS3 (DPs)	200 pg/μL <sup>13</sup> C-DPs
IS4 (PFASs)	0.1 ng/μL <sup>13</sup> C-PFAS mix

800

801

802 **Table S6:** Targeted compounds for per-and polyfluorinated substance analysed in plasma and feather samples  
 803 from white-tailed eagle nestlings, sampled at Steigen and Smøla (Norway) in 2015 and 2016. Limit of  
 804 quantification (LOQ) is given as ng/mL for plasma and ng/g for feathers.

<b>Per- and polyfluorinated substances (PFASs)</b>						
	<b>Abbreviations</b>	<b>Compounds</b>	<b>LOQ plasma 2015</b>	<b>LOQ feathers 2015</b>	<b>LOQ plasma 2016</b>	<b>LOQ feathers 2016</b>
<b>Carboxylic acids</b>	PFBA	Perfluorobutanoic acid	169.11	0.498	0.05	
	PFPeA	Perfluoropentanoic acid	1.31	0.200	0.05	
	PFHxA	Perfluorohexanoic acid	1.31	0.002	0.10	0.193
	PFHpA	Perfluoroheptanoic acid	1.31	0.002	0.05	0.267
	PFOA	Perfluorooctanoic acid	0.10	0.029	0.05	0.210
	PFNA	Perfluorononanoic acid	0.10	0.029	0.08	0.259
	PFDCa	Perfluorodecanoic acid	0.10	0.029	0.05	0.262
	PFUnA	Perfluoroundecanoic acid	0.20	0.029	0.08	0.181
	PFDoA	Perfluorododecanoic acid	0.20	0.029	0.08	0.145
	PFTTrA	Perfluorotridecanoic acid	0.20	0.029	0.10	0.186
	PFTTeA	Perfluorotetradecanoic acid	0.20	0.029	0.10	0.200
<b>Sulfonamides</b>	PFOSA	Perfluorooctanesulfonamide	23.35	0.029	0.10	0.200
<b>Sulfonic acids</b>	PFBS	Perfluorobutane sulfonate	169.11	0.498	0.05	0.100
	PFPS	Perfluoropentane sulfonate		0.002		0.100
	PFHxS	Perfluorohexane sulfonate	0.10	0.002	0.05	0.393
	PFHpS	Perfluoroheptane sulfonate	0.01	0.002	0.08	0.127
	Lin-PFOS	Linear perfluorooctane sulfonate	0.20	0.029	0.10	0.050
	Br-PFOS	Branched perfluorooctane sulfonate	0.20	0.029	0.10	0.050
	PFNS	Perfluorononane sulfonate		0.030	0.10	0.050

805

806

807 **Additional analytical details**

808 **Clean-up of sample extracts for PCB, PBDE, OCP, NBFR, DP and PFR analyses**

809 Further procedures for clean-up, fractionation of the concentrated extracts and quantification  
810 were the same for plasma and feathers. Fractionation was performed on Supelclean™  
811 ENVI™18 Florisil cartridges (500 mg, 3mL, Supelco® Analytical). Anhydrous sodium sulfate  
812 (Na<sub>2</sub>SO<sub>4</sub>) was added to the cartridge before cleaning with 6 mL ethyl acetate, followed by 6  
813 mL *n*-hexane. The sample extract was transferred to the cartridge. The sample tube was washed  
814 twice with 0.5 mL of *n*-hexane and vortexed. The extracts were eluted in two fractions: the first  
815 fraction (F1), containing PCBs, PBDEs, DPs, NBFRs, and OCPs was eluted with 10 mL *n*-  
816 hexane:DCM (1:1, v/v). A second fraction (F2) was collected, only for plasma and feather  
817 extracts, in new tubes, containing the PFRs. F2 was eluted from the same columns as F1 by 10  
818 mL ethyl acetate. Both fractions were evaporated to near dryness by a gentle nitrogen steam.  
819 F1 was re-solubilised in 0.5 mL of *n*-hexane, followed by a second clean-up of F1 on acidified  
820 silica (5 %) in a 3 mL cartridge, pre-cleaned with 6 mL *n*-hexane. The tube was washed twice  
821 with 0.5 mL of *n*-hexane and vortexed.

822 The preen oil samples were only cleaned up on 6 mL columns, containing about 2.5 g of 44 %  
823 acidified silica. The cartridges were pre-cleaned with 6 mL *n*-hexane. Samples were added,  
824 and the tubes were washed twice with 1 mL of *n*-hexane and transferred to the cartridge. Preen  
825 oil samples were eluted with 10 mL of hexane:DCM (1:1, v/v).

826 All extracts were finally concentrated to near dryness under a gentle nitrogen stream and re-  
827 solubilized in 100 µL of iso octane. For each batch of 24 samples, 100 µL recovery standard  
828 (RS, CB 207, 50 pg/µL in iso-octane toluene 9:1, v/v) was added to five samples and vortexed  
829 for 30 s. Extracts were transferred to injection vials for gas chromatography with electron  
830 capture negative ionization and mass spectrometry (GC-ECNI/MS) analysis.

831

832 **Instrumental analysis of PCBs, PBDEs, OCPs, NBFRs, DPs and PFRs**

833 The analysis was performed with an Agilent 6890 GC (Palo Alto, CA, USA) coupled to an  
834 Agilent 5973 MS operated in electron capture negative ionization (ECNI) mode and equipped  
835 with a DB 5ms capillary column (30 m x 0.25 mm x 0.25 mm). The GC system was equipped  
836 with electronic pressure control and a programmable temperature vaporizer (PTV) inlet. The  
837 injection temperature was set at 92 °C, held 0.03 min, ramped at 700 °C/min to 300 °C, held  
838 30 min. Injection (1 µL) was performed under a pressure of 10.06 psi until 1.25 min and purge  
839 flow to split vent of 50 mL/min after 1.25 min. The GC temperature ramp started from 92 °C,

840 held 1.25 min, ramped at 10 °C/min to 300 °C, held 1 min, ramped at 40 °C/min to 310 °C and  
841 held 9.5 min. Helium was used as a carrier gas with a flow rate of 1.0 mL/min until 25 min,  
842 then increased to 1.5 mL/min. The ion source and quadrupole temperatures were set at 170 °C  
843 and 150 °C, respectively. The mass spectrometer was operated in selected ion monitoring  
844 (SIM) for the quantification of BDE 28, 47, 100, 99, 154, 153, 183, BTBPE, *s*-DP, *a*-DP, CB  
845 101, 99, 118, 153, 138, 187, 183, 180, 170, oxychlordan (OxC), trans nonachlor (TN), HCB,  
846 *p,p'*-DDE, *p,p'*-DDT,  $\alpha$ -HCH,  $\beta$ -HCH,  $\gamma$ -HCH. BDE 103 and BDE 128 were used as IS for all  
847 PBDE congeners and BTBPE; 13C *s*-DP and 13C *a*-DP, were used as IS for *s*-DP, and *a*-DP,  
848 respectively; CB 143 was used as IS for the targeted PCBs and OCPs.

849

850 PFRs were analysed using an Agilent 6890 GC coupled to an Agilent 5973 MS operated in EI  
851 mode. The GC system was equipped with an HT 8 column (25 m x 0.22 mm x 0.25 mm),  
852 electronic pressure control and a PTV inlet. The injection temperature was set at 80 °C, held  
853 0.03 min, ramped at 700 °C/min to 300 °C, held 40 min. Injection (1  $\mu$ L) was performed under  
854 a pressure of 13.65 psi until 1.25 min and purge flow to split vent of 50 mL/min after 1.25 min.  
855 The GC temperature ramp started from 80 °C, held 1.25 min, ramped at 15 °C/min to 200 °C,  
856 held 3 min, ramped at 5 °C/min to 270 °C, ramped at 20 °C/min to 310 °C and held 12 min.  
857 Helium was used as a carrier gas with a flow rate of 1.0 mL/min until 28 min, then increased  
858 to 1.5 mL/min. The mass spectrometer was run in SIM mode and TEHP, TCEP, TCIPP (2  
859 isomers), EHDPHP, TPHP, TDCIPP, TNBP and TBOEP were analysed. TAP was used as IS  
860 for TEHP, TNBP; TCEP-d12 was used for TCEP and TCIPP (2 isomers); TBOEP-d6 was used  
861 for TBOEP; TPHP-d15 was used for TPHP and EHDPHP; TDCIPP-d15 was used for TDCIPP.

862

863

864 **Table S7:** Recoveries of internal standards in plasma samples from white-tailed eagles from Smøla and Steigen  
 865 (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”.

Plasma	2015		2016	
	Mean	sd	Mean	sd
CB 143	69	39	108	8
e-HCH	53	12	79	11
BDE 77	83	9	87	10
<sup>13</sup> C-HCB	na	na	102	17
<sup>13</sup> C-TBPH	18	14	25	12
<sup>13</sup> C-s-DP	47	13	55	10
<sup>13</sup> C-a-DP	50	10	52	11
TAP	93	13	79	14
TCEP-d12	89	33	72	22
TBEP-d6	78	35	68	28
TPhP-d15	86	15	80	12
TDCPP-d15	92	15	80	11
<sup>13</sup> C-PFPA	101	16	87	12
<sup>13</sup> C-PFHxA	93	22	86	10
<sup>13</sup> C-PFHpA	90	16	93	12
<sup>13</sup> C-PFOA	88	15	87	12
<sup>13</sup> C-PFDcA	81	32	89	11
<sup>13</sup> C-PFUnA	77	13	84	12
<sup>13</sup> C-PFDoA	59	15	83	19
<sup>13</sup> C-PFHxS	90	17	87	12
<sup>13</sup> C-PFOS	79	17	84	12
<sup>13</sup> C-PFOSA	83	17	75	9

866

867



868 **Table S8:** Recoveries of internal standards in feathers and preen oil samples from white-tailed eagles from Smøla  
 869 and Steigen (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”.

Feathers	2015		2016	
	Mean	sd	Mean	sd
CB 143	79	10	92	5
e-HCH	75	10	82	4
BDE 77	91	9	86	10
<sup>13</sup> C-HCB	na	na	106	6
<sup>13</sup> C-TBPH	45	13	40	14
<sup>13</sup> C-s-DP	69	9	75	8
<sup>13</sup> C-a-DP	68	9	72	9
TAP	87	11	92	10
TCEP-d12	105	19	93	15
TBEP-d6	97	15	88	18
TPhP-d15	74	12	81	10
TDCPP-d15	82	13	85	11
<sup>13</sup> C-PFPA	43	5	101	22
<sup>13</sup> C-PFHxA	42	5	101	20
<sup>13</sup> C-PFHpA	40	4	118	29
<sup>13</sup> C-PFOA	37	15	114	22
<sup>13</sup> C-PFDcA	44	3	121	27
<sup>13</sup> C-PFUnA	43	7	121	29
<sup>13</sup> C-PFDoA	46	12	129	36
<sup>13</sup> C-PFHxS	39	3	150	55
<sup>13</sup> C-PFOS	38	6	89	16
<sup>13</sup> C-PFOSA	43	4	90	21
<b>Preen oil</b>	<b>Mean</b>	<b>sd</b>	<b>Mean</b>	<b>sd</b>
CB 143	93	9	84	13
e-HCH	88	14	76	10
BDE 77	120	9	104	11
<sup>13</sup> C-HCB	na	na	109	13

870

871 **Table S9:** Detection frequency, median, min and max concentrations of PCBs quantified in plasma samples from white-tailed eagles from Smøla and Steigen (Norway), from  
872 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Plasma concentrations are in ng/ml ww. Congeners not listed in this  
873 table were not detected in any of the samples.

Smøla									Steigen							
Plasma	2015 <i>n</i> = 13				2016 <i>n</i> = 22				2015 <i>n</i> = 14				2016 <i>n</i> = 21			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
CB 52	na				54	0.11	0.07	0.23	na				57	0.1	0.06	0.12
CB 74	na				9	0.18	0.12	0.25	na				24	0.15	0.06	0.25
CB 99	100	0.16	0.08	0.59	100	0.18	0.06	1.47	100	0.5	0.18	4.61	100	0.23	0.06	0.98
CB 101	100	0.21	0.09	0.31	100	0.14	0.01	0.56	100	0.16	0.07	0.56	86	0.14	0.02	0.25
CB 105	100	0.08	0.04	0.3	100	0.11	0.04	0.79	100	0.26	0.1	2.59	100	0.14	0.04	0.66
CB 118	100	0.23	0.11	0.81	100	0.41	0.17	2.92	100	0.7	0.28	7.3	100	0.50	0.14	2.30
CB 138	100	0.27	0.11	1.25	100	1.1	0.4	10.55	100	0.66	0.29	5.63	100	1.26	0.28	8.88
CB 149	na				14	0.38	0.09	0.47	na				na			
CB 153	100	0.74	0.21	3.06	100	1.44	0.55	9.48	100	2.05	1.12	26.27	100	1.75	0.43	10.16
CB 156	85	0.02	0.01	0.11	100	0.06	0.02	0.43	100	0.07	0.03	0.8	100	0.07	0.02	0.54
CB 170	100	0.07	0.02	0.36	100	0.22	0.07	1.3	100	0.18	0.06	2.16	100	0.23	0.07	1.98
CB 171	69	0.02	0.01	0.07	100	0.04	0.01	0.23	100	0.03	0.02	0.37	100	0.04	0.01	0.26
CB 177	77	0.02	0.01	0.07	100	0.04	0.02	0.43	100	0.03	0.01	0.18	100	0.05	0.01	0.17
CB 180	100	0.17	0.04	0.84	100	0.7	0.2	3.55	100	0.45	0.13	5.29	100	0.65	0.19	5.89
CB 183	85	0.04	0.01	0.19	100	0.12	0.04	0.76	100	0.1	0.04	1.15	100	0.13	0.03	1.03
CB 187	100	0.11	0.03	0.43	100	0.32	0.13	2.55	100	0.22	0.07	1.81	100	0.36	0.1	1.95
CB 194	69	0.03	0.02	0.08	100	0.08	0.02	0.3	93	0.05	0.02	0.38	100	0.07	0.02	0.79
CB 199	69	0.02	0.01	0.08	na				79	0.05	0.01	0.18	na			
CB 196/203	69	0.02	0.01	0.09	na				93	0.04	0.01	0.43	na			
CB 206	10	0.01	0.01	0.01	59	0.02	0.01	0.05	62	0.02	0.01	0.07	62	0.03	0.01	0.2
CB 209	nd				55	0.03	0.01	0.1	15	0.02	0.02	0.3	48	0.03	0.01	0.12

874

875 **Table S10:** Detection frequency, median, min and max concentrations of OCPs and PBDEs quantified in plasma samples from white-tailed eagles from Smøla and Steigen  
876 (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Plasma concentrations are in ng/ml ww. OCPs and  
877 PBDEs not listed in this table were not detected in any of the samples.

Plasma	Smøla								Steigen							
	2015 <i>n</i> = 13				2016 <i>n</i> = 22				2015 <i>n</i> = 14				2016 <i>n</i> = 21			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
OxC	100	0.04	0.02	0.14	100	0.08	0.02	0.53	100	0.24	0.05	2.16	100	0.13	0.04	0.6
TN	100	0.19	0.06	0.36	100	0.26	0.08	0.98	100	0.22	0.1	1.22	100	0.29	0.14	0.59
CN	100	0.07	0.04	0.12	100	0.13	0.05	0.39	100	0.07	0.04	0.44	100	0.14	0.08	0.25
<i>p,p'</i> -DDE	100	1.21	0.56	5.23	100	1.2	0.56	9.47	100	3.95	2.18	47.61	100	1.45	0.48	8.64
<i>p,p'</i> -DDD	na				77	0.09	0.03	0.32	na				67	0.08	0.05	0.19
<i>p,p'</i> -DDT	100	0.2	0.08	0.3	100	0.15	0.06	0.63	86	0.13	0.06	0.31	95	0.27	0.09	0.38
HCB	92	0.09	0.04	0.21	100	0.76	0.26	2.96	100	0.15	0.05	0.8	100	1.02	0.32	2.46
$\alpha$ -HCH	nd				nd				7	0.04	0.04	0.04	nd			
$\beta$ -HCH	100	0.04	0.03	0.06	55	0.02	0.01	0.08	100	0.06	0.03	0.32	86	0.02	0.01	0.07
$\gamma$ -HCH	nd				5	0.04	0.04	0.04	nd				nd			
BDE 28	85	0.003	0.002	0.005	23	0.003	0.002	0.005	86	0.01	0.004	0.02	nd			
BDE 47	100	0.06	0.03	0.28	100	0.08	0.01	0.81	100	0.19	0.06	1.82	100	0.09	0.01	0.36
BDE 99	85	0.01	0.003	0.03	86	0.02	0.003	0.16	100	0.04	0.01	0.23	95	0.03	0.002	0.08
BDE 100	100	0.03	0.01	0.11	96	0.03	0.004	0.35	100	0.08	0.02	0.5	100	0.03	0.002	0.14
BDE 153	62	0.01	0.003	0.02	96	0.01	0.004	0.08	100	0.02	0.003	0.07	95	0.01	0.002	0.08
BDE 154	92	0.01	0.004	0.03	100	0.03	0.01	0.17	93	0.01	0.01	0.04	100	0.02	0.003	0.99
BDE 183	nd				36	0.01	0.01	0.03	nd				33	0.01	0	0.01
2'-MeO-BDE68	na				45	0.02	0.01	0.08	na				81	0.01	0.01	0.04
6-MeO-BDE47	na				50	0.04	0.01	0.16	na				86	0.05	0.01	0.12

878

879

880 **Table S11:** Detection frequency, median, min and max concentrations of PCBs quantified in feather samples from white-tailed eagles from Smøla and Steigen (Norway), from  
 881 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Feather concentrations are in ng/g ww. Congeners not listed in this  
 882 table were not detected in any of the samples.

Feathers	Smøla								Steigen							
	2015 <i>n</i> = 13				2016 <i>n</i> = 22				2015 <i>n</i> = 14				2016 <i>n</i> = 21			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
CB 49	na				9	0.37	0.35	0.39	na				5	0.71	0.71	0.71
CB 52	na				59	0.70	0.31	1.14	na				81	0.79	0.37	1.33
CB 74	na				9	0.43	0.39	0.47	na				24	0.56	0.50	0.96
CB 95	na				9	0.22	0.21	0.23	na				na			
CB 99	92	1.18	0.40	3.85	100	0.92	0.41	7.89	100	7.78	2.71	31.05	100	1.08	0.21	3.61
CB 101	92	0.75	0.56	1.82	100	0.55	0.31	1.57	100	1.50	0.80	1.81	95	0.44	0.20	0.99
CB 105	85	0.30	0.11	0.57	100	0.24	0.11	0.94	100	1.37	0.52	3.85	100	0.31	0.12	1.05
CB 118	100	0.72	0.23	2.10	100	0.90	0.45	4.12	100	5.12	2.18	15.03	100	1.07	0.40	3.37
CB 138	100	0.42	0.17	1.51	100	1.86	0.75	11.78	100	1.34	0.60	5.89	100	2.64	0.62	6.26
CB 153	100	1.77	0.63	6.77	100	3.13	1.22	17.07	100	12.86	5.48	38.64	100	4.15	0.92	9.73
CB 156	8	0.10	0.10	0.10	36	0.12	0.10	0.20	71	0.18	0.10	0.45	52	0.16	0.11	0.26
CB 170	69	0.19	0.11	0.48	100	0.34	0.13	1.45	100	0.52	0.27	1.73	100	0.42	0.11	1.23
CB 171	15	0.12	0.10	0.14	14	0.12	0.11	0.34	57	0.20	0.16	0.47	48	0.15	0.10	0.19
CB 177	23	0.12	0.10	0.18	14	0.11	0.11	0.72	57	0.17	0.11	0.31	33	0.14	0.11	0.17
CB 180	100	0.23	0.10	1.03	100	0.74	0.26	3.06	100	0.96	0.56	3.30	100	0.92	0.24	2.73
CB 183	38	0.19	0.10	0.30	82	0.18	0.13	0.84	100	0.38	0.17	1.37	86	0.26	0.10	0.58
CB 187	85	0.19	0.10	0.78	100	0.48	0.17	3.21	100	0.76	0.26	1.89	100	0.45	0.14	1.08
CB 194	23	0.13	0.12	0.14	45	0.15	0.11	0.40	64	0.18	0.10	0.36	62	0.20	0.10	0.36
CB 199	8	0.10	0.10	0.10	na				14	0.15	0.12	0.18	na			
CB 196/203	8	0.10	0.10	0.10	na				57	0.17	0.11	0.47	na			
CB 206	nd				nd				7	0.11	0.11	0.11	nd			
CB 209	nd				23	0.14	0.12	0.22	7	0.11	0.11	0.11	24	0.14	0.1	0.16

883

884 **Table S12:** Detection frequency, median, min and max concentrations of OCPs and PBDEs quantified in feather samples from white-tailed eagles from Smøla and Steigen  
885 (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Feather concentrations are in ng/g ww. OCPs and  
886 PBDEs not listed in this table were not detected in any of the samples.

Feathers	Smøla								Steigen							
	2015 <i>n</i> = 13				2016 <i>n</i> = 22				2015 <i>n</i> = 14				2016 <i>n</i> = 21			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
OxC	69	0.13	0.10	0.32	100	0.26	0.13	1.28	100	1.36	0.36	6.89	100	0.48	0.08	1.47
TN	46	0.15	0.10	0.32	95	0.16	0.11	0.29	100	0.24	0.13	0.41	100	0.16	0.10	0.33
CN	15	0.11	0.10	0.12	59	0.11	0.10	0.16	64	0.12	0.07	0.16	67	0.15	0.10	0.17
<i>p,p'</i> -DDE	100	4.96	2.35	22.10	100	3.23	1.48	17.03	100	26.59	12.38	94.38	100	3.74	1.14	8.81
<i>p,p'</i> -DDD	na				50	0.66	0.28	0.97	na				57	0.55	0.37	1.30
<i>p,p'</i> -DDT	69	0.71	0.31	1.14	73	0.28	0.21	1.03	86	0.91	0.55	1.92	71	0.27	0.20	0.47
$\alpha$ -HCH	nd				5	0.13	0.13	0.13	nd				10	0.12	0.10	0.13
$\beta$ -HCH	77	0.30	0.17	0.44	36	0.12	0.10	0.29	100	0.46	0.18	0.71	48	0.12	0.10	0.18
$\gamma$ -HCH	92	0.15	0.10	0.24	14	0.11	0.11	0.11	100	0.14	0.05	0.18	5	0.17	0.17	0.17
HCB	23	0.14	0.11	0.17	100	0.32	0.21	0.67	100	0.18	0.09	0.33	100	0.40	0.25	0.99
BDE 28	nd				nd				7	0.10	0.10	0.10	nd			
BDE 47	100	0.56	0.28	1.58	100	0.30	0.16	2.43	100	1.73	0.63	3.59	100	0.45	0.14	1.06
BDE 99	69	0.16	0.1	0.37	23	0.11	0.10	0.49	100	0.32	0.14	0.69	48	0.15	0.10	0.2
BDE 100	92	0.40	0.11	3.76	27	0.15	0.12	0.18	93	0.16	0.12	5.08	14	0.11	0.10	0.27
BDE 153	31	0.16	0.15	0.21	nd				nd				nd			
BDE 154	nd				5	0.12	0.12	0.12	nd				nd			
6-MeO-BDE47	na				73	0.12	0.10	0.18	na				71	0.12	0.10	0.19

887

888

889 **Table S13:** Detection frequency, median, min and max concentrations of DPs, PFRs and NBFs quantified in plasma and feather samples from white-tailed eagles from Smøla  
890 and Steigen (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Plasma concentrations are in ng/ml ww,  
891 while feather concentrations are in ng/g ww. PFRs and NBFs not listed in this table were not detected in any of the samples.

Smøla					Steigen											
2015 <i>n</i> = 13					2016 <i>n</i> = 22				2015 <i>n</i> = 14				2016 <i>n</i> = 21			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
<b>Plasma</b>																
TPhP	15	0.14	0.11	0.17	nd				36	0.19	0.11	0.55	nd			
EHDPPH	15	0.22	0.16	0.27	nd				29	0.12	0.10	0.28	nd			
TDCIPP	nd				nd				71	0.55	0.14	1.41	nd			
TCPP	8	0.14	0.14	0.14	nd				29	0.18	0.10	0.34	nd			
BTBPE	nd				nd				nd				5	0.01	0.01	0.01
TBPH	nd				nd				nd				5	0.08	0.08	0.08
<i>s</i> -DP	31	0.01	0.0002	0.02	5	0.002	0.002	0.002	7	0.06	0.06	0.06	nd			
<i>α</i> -DP	38	0.01	0.003	0.03	5	0.01	0.01	0.01	7	0.01	0.01	0.01	nd			
<b>Feathers</b>																
TPhP	100	21.65	2.54	1229.25	95	3.59	1.29	10.89	100	7.98	3.26	64.34	81	3.08	1.29	10.75
EHDPPH	85	2.67	1.43	11.49	73	3.11	1.28	8.74	100	2.48	1.07	4.42	76	2.19	1.06	3.91
TDCIPP	62	2.61	1.07	8.32	41	2.30	1.30	8.20	71	1.87	1.27	4.28	19	2.70	1.20	4.00
TCPP1	100	18.43	9.22	52.34	100	24.40	6.08	139.41	100	13.27	7.84	28.18	100	20.89	13.13	55.98
TCPP2	54	4.56	1.65	7.28	68	4.94	1.97	29.19	57	2.44	1.38	4.88	33	4.82	2.88	12.38
ΣTCPP		21.21	10.30	59.62		26.95	6.10	168.60		13.96	7.84	32.64		20.90	13.10	68.40
TCEP	92	9.47	3.72	14.79	100	9.28	4.24	23.42	100	4.83	2.43	9.01	100	7.57	4.56	14.65
TBB	nd				nd				7	0.64	0.64	0.64	nd			
TBPH	31	0.62	0.22	0.70	8	0.44	0.44	0.44	36	0.51	0.40	1.03	nd			
<i>s</i> -DP	8	0.10	0.10	0.10	14	0.10	0.10	0.17	nd				33	0.17	0.11	0.43
<i>α</i> -DP	23	0.13	0.11	0.47	32	0.17	0.10	0.62	21	0.11	0.0001	0.76	33	0.43	0.14	1.14

892

893 **Table S14:** Detection frequency, median, min and max concentrations of PFASs quantified in plasma and feather samples from white-tailed eagles from Smøla and Steigen  
894 (Norway), from 2015 and 2016. Compounds not analysed are marked with “na”, while not detected are marked with “nd”. Plasma concentrations are in ng/ml ww, while feather  
895 concentrations are in ng/g ww. PFASs not listed in this table were not detected in any of the samples.

Smøla					Steigen											
2015					2016				2015				2016			
	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max	df (%)	median	min	max
<b>Plasma</b>	<i>n</i> = 13				<i>n</i> = 22				<i>n</i> = 14				<i>n</i> = 21			
PFOSA	nd				9	0.25	0.21	0.28	nd				nd			
PFHxS	15	1.64	1.49	1.79	86	0.09	0.04	0.48	21	0.91	0.90	1.25	81	0.28	0.06	0.57
PFHpS	54	0.11	0.02	0.26	nd				57	0.15	0.08	0.31	24	0.21	0.08	0.29
branched PFOS	100	2.23	0.55	4.20	100	0.65	0.29	1.49	100	5.38	1.85	11.70	100	2.12	0.72	6.83
linear PFOS	100	14.12	6.04	31.85	100	5.25	2.34	8.47	100	16.55	9.55	27.07	100	7.01	3.86	17.5
PFOA	92	0.35	0.12	0.57	86	0.12	0.03	0.29	93	0.53	0.14	1.27	100	0.40	0.11	0.95
PFNA	100	1.82	0.57	4.86	100	0.56	0.35	1.77	100	3.58	1.56	6.48	100	1.69	0.63	6.60
PFDCa	100	1.22	0.66	2.3	100	0.39	0.25	0.82	100	1.44	0.91	2.52	95	0.69	0.36	1.82
PFUnA	100	3.59	2.43	4.36	100	1.15	0.68	2.05	100	3.36	2.30	5.08	100	1.40	0.94	2.15
PFDoA	85	0.57	0.32	0.94	100	0.27	0.09	0.46	100	0.38	0.20	0.86	100	0.22	0.15	0.51
PFTriA	62	0.94	0.31	1.66	100	0.31	0.14	0.65	71	0.63	0.20	1.02	100	0.29	0.07	0.62
<b>Feathers</b>	<i>n</i> = 13				<i>n</i> = 22				<i>n</i> = 14				<i>n</i> = 19			
PFOSA	46	0.56	0.23	0.86	95	1.43	0.47	6.52	50	0.53	0.34	1.11	100	1.39	0.45	3.17
PFHxS	nd				59	2.7	1.16	5.50	29	2.92	0.62	5.17	68	3.25	1.03	7.70
PFHpS	nd				14	0.37	0.28	3.27	nd				58	0.35	0.24	0.92
linear PFOS	8	0.22	0.22	0.22	68	1.33	0.47	90.2	29	0.92	0.58	1.82	89	1.69	0.88	4.49
PFHxA	nd				5	0.27	0.27	0.27	21	0.04	0.02	0.07	32	0.28	0.20	0.33
PFOA	nd				36	0.34	0.24	0.91	7	0.11	0.11	0.11	5	0.23	0.23	0.23
PFNA	nd				45	0.4	0.27	0.45	21	0.18	0.05	0.28	79	0.37	0.29	0.45
PFDCa	8	0.07	0.07	0.07	14	0.29	0.23	0.30	nd				58	0.28	0.23	0.46
PFUnA	77	0.18	0.05	0.60	95	0.58	0.34	0.95	64	0.57	0.37	0.81	100	0.82	0.26	1.07
PFDoA	23	0.18	0.07	0.20	32	0.24	0.2	0.50	21	0.14	0.11	0.15	47	0.29	0.20	0.78
PFTriA	31	0.28	0.11	0.51	91	0.66	0.25	1.59	50	0.30	0.11	0.46	100	1.14	0.50	2.02
PFTeA	nd				9	0.23	0.21	0.32	nd				53	0.30	0.19	0.91

897 **Table S15:** Detection frequency, median, min and max concentrations of PCBs, OCPs, PBDEs, NBFs and DPs  
898 quantified in preen oil samples from white-tailed eagles from Smøla and Steigen (Norway), from 2016.  
899 Compounds not detected are marked with nd. Preen oil was not sampled in 2015. Preen oil concentrations are in  
900 ng/g ww. PCBs, OPCs, PBDEs and NBFs not listed here were not detected in any of the samples.

Preen oil		Smøla				Steigen			
		2016		<i>n</i> = 22		2016		<i>n</i> = 21	
		df (%)	median	min	max	df (%)	median	min	max
Polychlorinated biphenyls (PCBs)	CB 28	18	8.32	5.28	10.27	18	10.15	8.16	12.43
	CB 49	18	3.33	3.13	3.52	38	4.84	2.88	7.22
	CB 52	73	8.81	4.94	12.89	95	9.54	7.07	15.08
	CB 74	14	22.01	20.20	26.79	43	11.25	5.36	30.78
	CB 95	91	8.85	2.75	16.15	95	9.98	5.89	14.65
	CB 99	100	28.00	12.33	198.68	95	35.10	14.22	161.58
	CB 101	100	16.68	9.99	49.56	95	19.95	12.03	32.62
	CB 105	100	14.94	6.70	70.56	100	22.33	7.99	91.04
	CB 118	100	51.84	19.68	240.09	100	66.46	23.94	289.96
	CB 138	100	129.63	39.37	720.65	100	168.08	40.26	602.90
	CB 149	100	16.39	9.04	58.56	95	15.30	11.71	26.81
	CB 153	100	259.08	81.81	1420.57	100	327.32	94.03	1163.43
	CB 156	100	10.20	3.32	37.61	100	14.21	4.17	54.16
	CB 170	100	37.74	10.16	180.79	100	44.12	12.85	184.32
	CB 171	100	5.69	1.86	35.80	100	7.10	2.12	25.53
	CB 177	100	7.01	2.44	78.40	100	6.04	2.74	16.39
	CB 180	100	117.86	29.89	489.81	100	139.58	40.90	579.26
	CB 183	100	21.96	6.03	102.53	100	25.66	7.31	99.74
	CB 187	100	55.94	16.74	434.19	100	59.11	20.74	186.85
	CB 194	100	9.44	1.56	33.76	100	9.11	1.63	69.10
CB 206	100	3.87	1.31	17.45	90	3.57	1.15	19.34	
CB 209	100	4.50	1.01	25.02	76	4.56	1.16	14.56	
Organochlorinated pesticides (OCPs)	OxC	100	9.52	3.36	30.37	100	13.55	4.80	53.73
	TN	100	34.92	14.32	63.44	100	39.04	23.64	75.39
	CN	100	13.22	6.47	22.28	100	15.75	10.03	26.21
	<i>p,p'</i> -DDE	100	289.56	170.99	1447.32	100	364.88	185.38	932.96
	<i>p,p'</i> -DDD	91	13.39	7.89	29.52	95	13.40	9.59	23.43
	<i>p,p'</i> -DDT	100	7.07	3.85	15.09	100	7.65	5.06	46.63
	$\beta$ -HCH	82	1.67	1.13	7.07	100	2.32	1.05	4.41
Polybrominated diphenyl ethers (PBDEs)	BDE 28	18	0.60	0.36	1.66	33	0.43	0.37	1.51
	BDE 47	100	9.19	4.25	96.35	100	10.47	4.45	60.28
	BDE 99	100	1.68	0.81	10.34	100	1.67	0.83	6.70
	BDE 100	100	2.89	0.99	22.25	100	3.36	1.25	17.20
	BDE 153	82	1.26	0.76	5.77	67	1.44	0.83	3.83
	BDE 154	100	1.68	0.81	10.34	100	1.67	0.83	6.70
	BDE 183	nd				5	0.75	0.75	0.75
	6-MeO-BDE47	95	8.17	2.43	21.57	100	12.44	7.33	74.81
2-MeO-BDE68	95	1.84	0.70	3.66	48	2.52	1.21	8.42	
Dechlorane plus (DPs)	<i>s</i> -DP	nd				nd			
	<i>a</i> -DP	5	0.45	-	-	nd			



