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TITLE: Acute hydrogen peroxide exposure does not cause oxidative stress in late-copepodite stage of *Calanus finmarchicus*

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23 **Acute hydrogen peroxide exposure does not cause oxidative stress**
24 **in late-copepodite stage of *Calanus finmarchicus***

25
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34

35 **Abstract**

36 Use of hydrogen peroxide (H₂O₂) for removal of salmon lice in the aquaculture industry has been met
37 with a concern that non-target organisms may be affected during treatment scenarios. In the present
38 work, we evaluated the potential for H₂O₂ to cause reduced survival and oxidative stress in one of the
39 most abundant zooplankton species in Norwegian coastal areas, the copepod *Calanus finmarchicus*.
40 We subjected the copepod to two 96-hour tests: An acute toxicity test where mortality was
41 determined, and a second experiment where we treated copepods to an exposure concentration
42 below the No Observed Effect Concentration (0.75 mg L⁻¹) and analyzed for antioxidant enzymes,
43 glutathione (GSH) and malondialdehyde (MDA). Compared to available and comparable LC₅₀-values
44 from the literature, our results suggests that *C. finmarchicus* is highly sensitive to H₂O₂. However, 96

45 hour exposure of *C. finmarchicus* to 0.75 mg H₂O₂ L⁻¹ did not cause responses on the antioxidant
46 systems even though the concentration is just below the level where mortality is expected. This
47 suggests that aqueous H₂O₂ exposure does not cause cellular accumulation with associated cellular
48 oxidative stress, but rather cause acute effects on copepod surface (carapace). This certainly needs
49 further investigation in order to assure that aqueous exposure during H₂O₂ treatment in salmon fish
50 farms does not have adverse effects on local non-target crustacean species and populations.
51 Particularly, studies on copepod developmental stages with a more permeable carapace is warranted.

52

53 **Key words:** Superoxide dismutase; catalase; glutathione peroxidase; glutathione; glutathione S-
54 transferase; lipid peroxidation; H₂O₂

55 1. Introduction

56 In order to reduce the devastating impacts of salmon lice *Lepeophtheirus salmonis* (Copepoda:
57 Caligidae) on the marine salmon production and farming, several pesticides have been used over the
58 years, i.e. organophosphates, pyrethroids, avermectins, chitin synthesis inhibitors and topical
59 disinfectants (Denholm et al., 2002). The extensive use has resulted in genetic resistance in salmon lice
60 to some of these, mainly the organophosphates, carbamates and pyrethroids. Hydrogen peroxide
61 (H_2O_2), which was introduced in Norway as a treatment in the 90s, is by far the most widely used
62 salmon lice pesticide in Norway (31,577 tons used in 2014). H_2O_2 is a prooxidant and functions as a
63 pest control by being a topic disinfectant, meaning it is administered directly through water to remove
64 attached lice from the fish skin. Because it has an acute effect and is readily broken down to oxygen
65 and water through abiotic processes, it is considered less environmentally harmful than many of the
66 other pesticides which have much larger acute-to-chronic ratios (ACRs>300,000) due to their specific
67 (and delayed) modes of toxic action (Lillicrap et al., 2015). However, high H_2O_2 concentrations are
68 necessary for lice removal causing concern for acute effects on the treated fish as well as for organisms
69 in the environment surrounding the fish farms. Atlantic salmon (*Salmo salar*) post-smolts exposed to
70 $1230\text{ mg L}^{-1} H_2O_2$ for 20 min at 13.5°C suffered an acute toxicity resulting in a 35% mortality within 2
71 hours. Thirty-three per cent of adult and pre-adult sea lice (*L. salmonis*) were immobilized or killed
72 following exposure to $500\text{ mg L}^{-1} H_2O_2$ at 10°C , rising to 98% at 2000 mg L^{-1} (Bruno&Raynard, 1994).

73 Due to the large quantities used, and the manner in which application is performed, there is a concern
74 about acute effects of H_2O_2 on non-target species like zooplankton. Acute toxicity levels of H_2O_2 to
75 crustaceans have been shown to be in the concentration range used for application in fish farms with
76 1 hour- LC_{50} s of 973, 1637 and 3182 mg L^{-1} for *Mysis* sp., American lobster (*Homarus americanus*, stage
77 1) and sand shrimp (*Crangon septemspinosa*), respectively (Burrige et al., 2014). Other types of
78 effects have also been reported at lower exposure concentrations, e.g. inhibited feeding activity in the
79 copepod *Acartia hudsonica* (Van Geest et al., 2014), reduced metabolic rate and muscle intracellular

80 pH in shrimp (*Crangon crangon*), and oxidative stress in the copepod *Tigriopus japonicus*
81 (Lee&Raisuddin, 2008).

82 Oxidative stress, i.e. the production of reactive oxygen species (ROS) like superoxide anion ($O_2^{\bullet-}$),
83 hydrogen peroxide (H_2O_2) and the highly reactive hydroxyl radical (OH^{\bullet}) may occur naturally during
84 oxidative phosphorylation in mitochondria or induced by redox cycling chemicals, transition metals
85 and many other compounds (Stohs&Bagchi, 1995; Livingstone, 2001). To reduce oxidative damage by
86 ROS (e.g. lipid peroxidation, DNA damage), organisms have antioxidant defense systems in the form of
87 enzymes, vitamins and pigments. Important antioxidant enzymes include superoxide dismutase (SOD,
88 EC 1.15.1.1), catalase (CAT, EC 1.11.1.6) and glutathione peroxidase (GPX, EC 1.11.1.9), and the latter
89 two detoxifies H_2O_2 and organic hydroperoxides resulting from oxidative stress. The tripeptide
90 glutathione (GSH) also functions as an antioxidant through conjugation to electrophilic substances
91 often catalyzed by the enzyme glutathione S-transferase (GST) (Di Giulio et al., 1995;
92 Halliwell&Gutteridge, 2015).

93 The main aim of this work was to evaluate the potential of acute aqueous H_2O_2 exposure to cause
94 oxidative stress in the copepod *Calanus finmarchicus* (Copepoda: Calanoida). This copepod is the most
95 abundant zooplankton species in the Norwegian Sea and serves a crucial role in the marine food web,
96 as it is an important transfer route for energy between primary producers and fish. Late developmental
97 stages consist of large lipid reservoirs, which are utilized for seasonal periods of diapause and
98 reproduction. In this work, we performed an acute toxicity test for H_2O_2 to determine a no-effect
99 concentration based on copepod mortality. Thereafter, copepods were treated with in a sub-lethal
100 H_2O_2 concentration where analyses of a battery of antioxidant enzymes was performed.

101

102 **2. Materials and Methods**

103 **2.1. Copepod culture**

104 Copepods (*C. finmarchicus*) from the continuous lab culture at SINTEF/NTNU Sealab were used for the
105 experiments on this species. They are routinely kept at 10 °C, and the details regarding the culturing
106 have previously been described (Hansen et al., 2007).

107

108 **2.2. Preparation of exposure medium**

109 The substance tested in this study was H₂O₂ (30% pro analysis Perhydrol®, Merck, Darmstadt,
110 Germany). Although we did not verify the H₂O₂ concentrations in the different treatment, the
111 substance was considered fully soluble in seawater at the respective desired concentrations. As
112 aeration and high temperatures increases the degradation of H₂O₂ (Bruno&Raynard, 1994; Burrige et
113 al., 2010), relatively low temperatures as well as capped bottles were used during treatment to limit
114 degradation during treatment. Stock solution was prepared by diluting a test substance in a deionized
115 water, and the stock solution was then dissolved in seawater in a 2L borosilicate glass bottle by
116 automatic pipettes to give the desired exposure concentrations.

117

118 **2.2. Acute toxicity test**

119 The acute toxicity test design applied for *C. finmarchicus* was adopted from standard tests on *Acartia*
120 *tonsa* (ISO) with adaptations to temperature (10±2°C), exposure volume (500 mL) and exposure time
121 (96 hours). Briefly, 7 copepods (CV or adults) were transferred to exposure bottles. A total of seven
122 exposure concentrations were used based on nominal dilutions of stock solution covering the range of
123 expected no effect to full immobilization after 96 hours based on a preliminary assay. At each
124 concentration, four replicates (N=4) were set up. Negative controls with seawater only was also
125 prepared (N=8). To reduce the loss of the active compound during exposure the seawater was filtered
126 to 0.22 µm (Sterivex™, Millipore) to reduce the level of organic particles. The animals were not fed

127 during exposure, and the test was performed in darkness except for when the survival was scored at
128 24, 48, 72 and 96 hours exposure.

129

130 **2.3. Sub-lethal experiment**

131 Two treatment groups were used in this experiment; Negative controls containing only filtered
132 seawater, and a 'treated group' exposed to a sub-lethal concentration of 0.75 mg L⁻¹ H₂O₂ based on the
133 96 hours-NOEC from the acute toxicity test. Both groups with five bottles (borosilicate, 2L) were
134 stocked with 70 copepods in each and a total volume of 2 L exposure solution. As for the acute test
135 the water used was filtered to 0.22 µm (Sterivex™, Millipore), the animals were not fed during
136 exposure, and the exposure was performed in darkness at 10±2°C.

137

138 **2.4. Determination of enzyme activities, concentrations of GSH and levels of MDA.**

139 The samples (50 individuals per sample) were homogenized using a MPW-309 universal laboratory aid
140 (Mechanika Precyzyjna, Warszawa) for 30 s. in ice-cold buffer (to obtain 20% homogenate) containing
141 50 mM Tris-H₂SO₄, pH 7.6 with 0.1 mM EDTA, 1 mM PMSF, 2 mM DTT and 0.2% Triton® X. The
142 homogenates were centrifuged at 14 000 RCF for 30 minutes at 4°C using Centrifuge SIGMA 3K18. The
143 supernatants were transferred into fresh tubes and used for analysis. Cytosolic protein was
144 determined by Lowry method (Lowry et al., 1951) with modification of Peterson (1977).

145 The enzymatic methods used for analyses of all oxidative stress markers described below were all
146 based on established methods utilized in a copepod species (*Limnocalanus macrurus*) from the
147 northern Baltic Sea (Vuori et al., 2015), and are therefore expected to work well for *C. finmarchicus* as
148 well.

149 Samples for the GSH determination were made from the protein supernatant via deproteinization by
150 adding a 1:1 volume of 5% sulfosalicylic acid (SSA). The sample was incubated with SSA on ice for 5 min

151 and centrifuged for 2 min at $10\,000 \times g$. The total glutathione (totGSH) concentration, which includes
152 reduced (GSH) and oxidized glutathione (GSSG) species was measured with Glutathione Assay Kit
153 (CS0260, Sigma-Aldrich) using GSH as the standard (Sigma-Aldrich). Samples were pipetted on a 96-
154 well microplate (Thermo Scientific), and working mixture reagent was added. The plate was then
155 incubated for 5 min at room temperature and kinetic read of absorbance was measured at 412 nm in
156 Synergy 2 Multi-Mode Reader (BioTek) to determine the level of totGSH. The totGSH were normalized to
157 the protein content of the samples.

158 Glutathione S-transferase (GST; EC 2.5.1.18) activity was determined spectrophotometrically (UV-VIS
159 Spectrophotometer, Beckman Coulter) by the method of Habig et al. (1974). The reaction mixture
160 contained 100 mM phosphate-buffered saline buffer (pH 6.5), 100 mM 1-chloro-2, 4-dinitrobenzene
161 (CDNB) and the reaction was started by the addition of supernatant and 100 mM glutathione as
162 substrate. The absorbance was followed for 5 min at 340 nm. Total GST activity was expressed in
163 nmol/min/mg of total protein concentration.

164 Catalase (CAT; EC 1.11.1.6) activity was measured following the method described by Kankofer (2001).
165 To supernatant cold 6 mM H_2O_2 was added and vortexed. After incubation in ice, the reaction was
166 stopped by 3 mM H_2SO_4 . Then 2 mM $KMnO_4$ was added, vortexed and the absorbance was read at 480
167 nm (UV-VIS Spectrophotometer, Beckman Coulter). The enzyme activity was determined by
168 measurement of H_2O_2 reacting with a standard excess of $KMnO_4$ and the detection of the residual
169 $KMnO_4$ spectrophotometrically. Catalase activity was expressed in $U\ mg^{-1}$ of total protein
170 concentration.

171 Superoxide Dismutase (SOD; EC 1.15.1.1) activity was measured using the modified method of Sun et
172 al. (Sun et al., 1988). There are several isoforms of SOD. Two main forms occurring in cells are CuZnSOD,
173 which is present primarily in the cytoplasm and the MnSOD form, which can be found in mitochondria.
174 In this experiment the sum of two isoforms of SOD were detected. SODs activity involved inhibition of
175 nitroblue tetrazolium reduction, with xanthine-xanthine oxidase used as a superoxide generator. The

176 reaction mixture contained 50 mM Na₂CO₃, 3 mM xanthine, 3 mM EDTA, 0.75 mM NBT (nitro blue
177 tetrazolium), 15% BSA (bovine serum albumin) and 0.05 mU/ml xanthine oxidase. The absorbance was
178 measured at a wavelength of 560 nm. The total activity was expressed in units per mg protein where
179 1 U of SOD was defined as the amount of protein that inhibits the rate of NBT reduction by 50%.

180 MDA level was detected using the Lipid Peroxidation (MDA) Assay Kit (MAK085, Sigma-Aldrich). In this
181 kit, lipid peroxidation was determined by the reaction of malondialdehyde (MDA) with thiobarbituric
182 acid (TBA) to form a colorimetric product, proportional to the MDA present. Samples were pipetted
183 on a 96 - well microplate and absorbance was measured at 532 nm in Synergy 2 Multi-Mode Reader
184 (BioTek). The amount of MDA was normalized to the protein content of the samples and expressed in
185 nmol/mg of total protein concentration.

186 Glutathione peroxidase (GPx; EC 1.11.1.9) activity was measured using the Glutathione Peroxidase
187 Cellular Activity Assay Kit (CGP1, Sigma - Aldrich) with 30 mM *tert*-butyl hydroperoxide as the
188 substrate. The assay system consisted of: 50 mM Tris-HCl, pH 8.0 with 0.5 mM EDTA, 5 mM NADPH,
189 42 mM GSH, 10 U/ml of glutathione reductase. The product of GPx catalyzed reaction: GSSG
190 (glutathione disulfide) was recycled to GSH using glutathione reductase and NADPH. Oxidation of
191 NADPH to NADP⁺ was monitored spectrophotometrically at 340 nm wavelength in Synergy 2 Multi-
192 Mode Reader (BioTek) and was proportional to GPx activity in the sample. Total GPx activity was
193 expressed in nmol/min/mg of total protein concentration. For this assay, limited sample was available,
194 so two pooled samples from each of the treatments were analyzed. Although not enough analytical
195 data for statistical analyses were obtained, the data is still included for reporting enzyme activity range
196 values for *C. finmarchicus*.

197

198 **2.5. Statistical analyses**

199 For the acute toxicity test, the number of immobilized copepods in replicate bottles in each
200 concentration of the exposure media was pooled before calculation, and normalized numbers of

201 immobilized copepods were calculated as percentage of control. The LC-values were calculated by a
202 non-linear model based on a sigmoidal dose-response model with variable slope (four-parameter
203 logistic equation) using GraphPad Prism version 6.00 for Macintosh (GraphPad Software, La Jolla,
204 California). When performing the calculations constraints on the exposure concentration-response
205 curve was placed both on top and bottom forcing the effect to be calculated within the interval 0-100%
206 effects. To compare the enzyme activities between the two groups in the sub-lethal experiment, the
207 non-parametric Mann-Whitney test was also performed using GraphPad Prism.

208

209

210 **3. Results and Discussion**

211 **3.1. Acute toxicity**

212 Despite being considered a relatively environmentally friendly pesticide to remove salmon lice, very
213 little information exists on the ecotoxicity of H₂O₂ in the literature. Acute toxicity tests performed on
214 H₂O₂ have been done almost exclusively on salmon lice, and the exposures have utilized high
215 concentrations and very short exposure times, typically 20 min to 5 hours (Burridge et al., 2014; Abele-
216 Oeschger et al., 1997; Bruno&Raynard, 1994). This because rapid dilution and degradation is expected
217 during real applications. High concentrations are used because acute mortality of salmon lice is the
218 desired outcome of the application, and fast dilution is expected when H₂O₂-containing water is
219 dumped into the sea after bath treatment. When concerned with exposures of non-target species,
220 more diffuse exposure regimes characterized by lower concentrations and longer exposure times may
221 be expected. Our acute toxicity tests on *C. finmarchicus* was performed over a period of 96 hours and
222 the highest concentration being 6 mg L⁻¹. Survival is plotted as a function of exposure concentrations
223 in Fig. 1. The figure displays the curves for 24, 48, 72 and 96 hours exposure. Our acute test included
224 readings from 24 hours and up to 96 hours because exposures of *C. finmarchicus* in the environment
225 will be expected to occur for longer periods and at much lower concentrations than typically found

226 inside aquaculture cages during application. This, however, makes our data less comparable to
227 available data from other species. LC₅₀ values decline exponentially with time as observed in our acute
228 experiment (Table 1). At 24 hours, the calculated LC₅₀ value was approx. 6 mg L⁻¹ declining to 2.5 mg L⁻¹
229 ¹ after 96 hours of exposure. Comparing our results to the few available and comparable acute toxicity
230 data from other species, it seems that *C. finmarchicus* is a much more sensitive species than a number
231 of other marine species. Substantially higher LC₅₀-values have been reported for brine shrimp (*Artemia*
232 *salina*, 24 hour-LC₅₀ of 800 mg L⁻¹) (Matthews, 1995), amphipod (*Corophium volutator*, 96 hour LC₅₀ of
233 46 mg L⁻¹) (Smit et al., 2008), zebra mussel (*Dreissena polymorpha*, 72 hour-LC₅₀ of 30 mg L⁻¹) (Martin
234 et al., 1993), rabbitfish (*Siganus fuscescens*, 24 hour-LC₅₀ of 224 mg L⁻¹), striped goby (*Tridentiger*
235 *trigonocephalus*, 24 hour-LC₅₀ of 155 mg L⁻¹) and jack mackerel (*Trachurus japonicus*, 24 hour-LC₅₀ of
236 89 mg L⁻¹) (Kanda et al., 1989). For mysids, lobster and shrimps, LC₅₀s in the range 1000 - 3000 mg L⁻¹
237 has been reported following 1 hour of exposure with a subsequent 95 hours recovery (Burrige et al.,
238 2014). Five hours exposure to 680 mg L⁻¹ caused altered metabolic rate and decreased intracellular pH
239 in the sand shrimp (Abele-Oeschger et al., 1997).

240

241 **3.2. Sub-lethal experiment – design**

242 The final experiment was designed to cause no mortality, as the concentration used (0.75 mg L⁻¹) did
243 not result in any mortality in the 96 hours acute toxicity test. The concentration may also be considered
244 more environmentally relevant as it is well below normal treatment concentrations for salmon lice
245 (1200-1500 mg L⁻¹; (Kierner&Black, 1997)) which is expected to be rapidly diluted in the sea. The
246 exposure time used (96 hours) was chosen in order to simulate an environmental exposure scenario,
247 but also to allow time for enzyme activation through gene expression. In *C. finmarchicus*, increased
248 stress gene expression has previously been reported already after 24 hours exposure to ionic mercury
249 (Øverjordet et al., 2014), polycyclic aromatic hydrocarbons (Hansen et al., 2008) and oil (Hansen et al.,
250 2011). No mortality was observed during this treatment.

251

252 **3.3. Antioxidant enzymes**

253 Microcrustaceans do possess antioxidant enzymes and the ability to detoxify prooxidants and deal with
254 environmentally stressful situations through enzymatic activity (Barata et al., 2005; Cailleaud et al.,
255 2007; Cailleaud et al., 2009). However, in the present experiment no significant responses on
256 antioxidant enzymatic activities were observed as a function of H₂O₂ exposure (Fig. 2).

257 Although the antioxidant systems are not well described for *C. finmarchicus*, it is expected to be
258 somewhat similar to vertebrates. In addition, other copepods, like *Tigriopus japonicus*, express SOD
259 (Kim et al., 2011), CAT (Han et al., 2015), GPx (Kim et al., 2015) and GSTs (Lee et al., 2008; Lee et al.,
260 2007) in response to oxidative stress-inducing pollutants. Importantly however, most studies on
261 antioxidants in microcrustaceans have involved gene expression analyses alone and not being
262 complemented with enzyme activity measurements, but there are exceptions, e.g. in the study on
263 *Daphnia magna* transplanted to polluted areas (Rivetti et al., 2015).

264 No significant differences between treatments were observed for SOD and CAT, however, a weak
265 negative relationship ($p=0.548$, $R^2=0.39$, linear regression, slope= -1.6) between their activities in the
266 samples was observed. This may be explained by the fact that the substrate of SOD inhibits CAT activity
267 (Kono&Fridovich, 1982). SOD converts the superoxide anion $\bullet O_2^-$ to H₂O₂, and as such provides the
268 substrate for CAT, which reduces H₂O₂ to water. GPX also detoxifies H₂O₂ and organic hydroperoxides
269 produced during oxidative stress, for example, by lipid peroxidation (Di Giulio et al., 1995;
270 Halliwell&Gutteridge, 2015). Lipid peroxidation is considered a major toxic mechanism by which
271 radicals cause tissue damage resulting in altered cell membranes and thereby disrupts vital cell
272 functions (Rikans&Hornbrook, 1997). Lipid peroxides are known to cause production of a variety of
273 substances, the most important of which is malondialdehyde (MDA) (Leibovitz&Siegel, 1980).
274 Particularly for lipid-storing copepods like *C. finmarchicus*, lipid peroxidation may have adverse
275 outcomes, as their discrete lipid storage is the basis for longer periods of starvation during winter

276 (diapause), and subsequently utilized during reproduction to produce eggs (Marshall&Orr, 1972). It
277 has been shown that the expression of ferritin was greater in diapausing copepodite V (CV) with large
278 oil sacs, consistent with a role of ferritin in chelating metals to protect cells from oxidative stress
279 suggesting that the integrity of lipids is of importance (Tarrant et al., 2008). However, evidence of lipid
280 peroxidation caused by exposure to 0.75 mg H₂O₂ L⁻¹ in the present exposure of CV copepods was not
281 provided, since MDA levels were unaltered by the exposure (Fig. 3B).

282 In *C. finmarchicus*, GST gene expression has shown to be responsive to alkanolamines (Hansen et al.,
283 2010), PAHs (Hansen et al., 2008), oil (Hansen et al., 2009; Hansen et al., 2011), marine fuels (Hansen
284 et al., 2013) and mercury (Øverjordet et al., 2014). However, the gene expression data have never
285 been complemented by enzyme activity analyses. In the present experiment, increased GST enzyme
286 activity was observed as a function of exposure, however not significant (p=0.07). In the copepod
287 *Limnocalanus macrurus*, enzymatic GST activities have been shown between different polluted sites,
288 and the enzyme activity range (0.56-4.58 µmol/min/mg protein) were comparable to our study (Vuori
289 et al., 2015). In addition, the copepod *Eurytemora affinis* has displayed GST enzyme responses to
290 environmental stress, however, for this species, enzyme activity range was much higher (600-800
291 mol/min/mg protein) (Cailleaud et al., 2009; Cailleaud et al., 2007). GST catalyzes the conjugation of
292 glutathione (GSH) with various electrophilic substances and as such plays an important role preventing
293 oxidative damage. Analyses of GSH in H₂O₂-treated *C. finmarchicus* were also in line with the responses
294 on GST enzyme activity, displaying no significant response (Fig. 3A). Measurements of total GSH may
295 be less sensitive to stress than using the ratio between reduced (GSH) and oxidized glutathione (GSSG),
296 however, our study only included the totGSH. The totGSH concentration range observed for *C.*
297 *finmarchicus* in our study was in the low end of what was observed for *Limnocalanus macrurus* (Vuori
298 et al., 2015), supporting low level of oxidative stress during our sublethal H₂O₂ experiment.

299 **4. Conclusion**

300 Exposure of *C. finmarchicus* to hydrogen peroxide (H₂O₂) does not appear to cause cellular oxidative
301 stress during the sub-lethal exposure concentration used in the present experiment (0.75 mg L⁻¹) even
302 though the concentration is just below concentrations causing mortality after 96-hours exposure. A
303 lack of antioxidant enzymatic response and no alterations in GSH and MDA levels to a 96-hour
304 treatment to waterborne H₂O₂ suggests that cellular oxidative stress is not a primary mode of action
305 for this chemical. This does not necessarily provides evidence for reduced concern for H₂O₂ exposure
306 on this species; as the sensitivity measured as 24 hour-LC₅₀ is 10 - 100-fold lower than most other
307 reported sensitivity levels in other species. As H₂O₂ is a very reactive compound, there is reason to
308 believe aqueous exposure does not cause cellular accumulation and associated cellular toxicity of this
309 chemical, but rather cause acute effects on copepods directly through effects on copepod surface
310 (carapace). This certainly needs further investigation in order to assure that aqueous exposure during
311 H₂O₂ application in salmon fish farms does not have health effects on non-targeted local crustacean
312 species and populations. Currently, H₂O₂ applications involves using bath treatments and subsequent
313 dumping of H₂O₂-containing water in more open water. Future studies should also involve
314 investigations of earlier life stages of copepods, which are less mobile as well as having a less protective
315 exoskeleton. Studies on commercially important species such as lobster support this suggestion as
316 early developmental stages have displayed higher sensitivity than older (Burrige et al., 2014). Also in
317 the few cases where *L. salmonis* has displayed resistance to H₂O₂, a mechanism related to reduced
318 carapace permeability has been proposed (Treasurer et al., 2000).

319

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324

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460 **Tables**

461

462 Table 1: Calculated LC₅₀-values for 24, 48, 72 and 96 hours of exposure of *Calanus finmarchicus* to
463 hydrogen peroxide. The values are listed as the nominal concentrations in mg H₂O₂ L⁻¹ causing 50%
464 effect with 95 % confidence intervals (CI).

Effect concentration	24 hour	48 hour	72 hour	96 hour
H₂O₂ mg/L	5.992*	3.912	3.824	2.540
95 % CI	NC**	3.187-4.801	3.536-4-136	2.356-2.738

465 *Calculated value ambiguous. **Not calculated

466

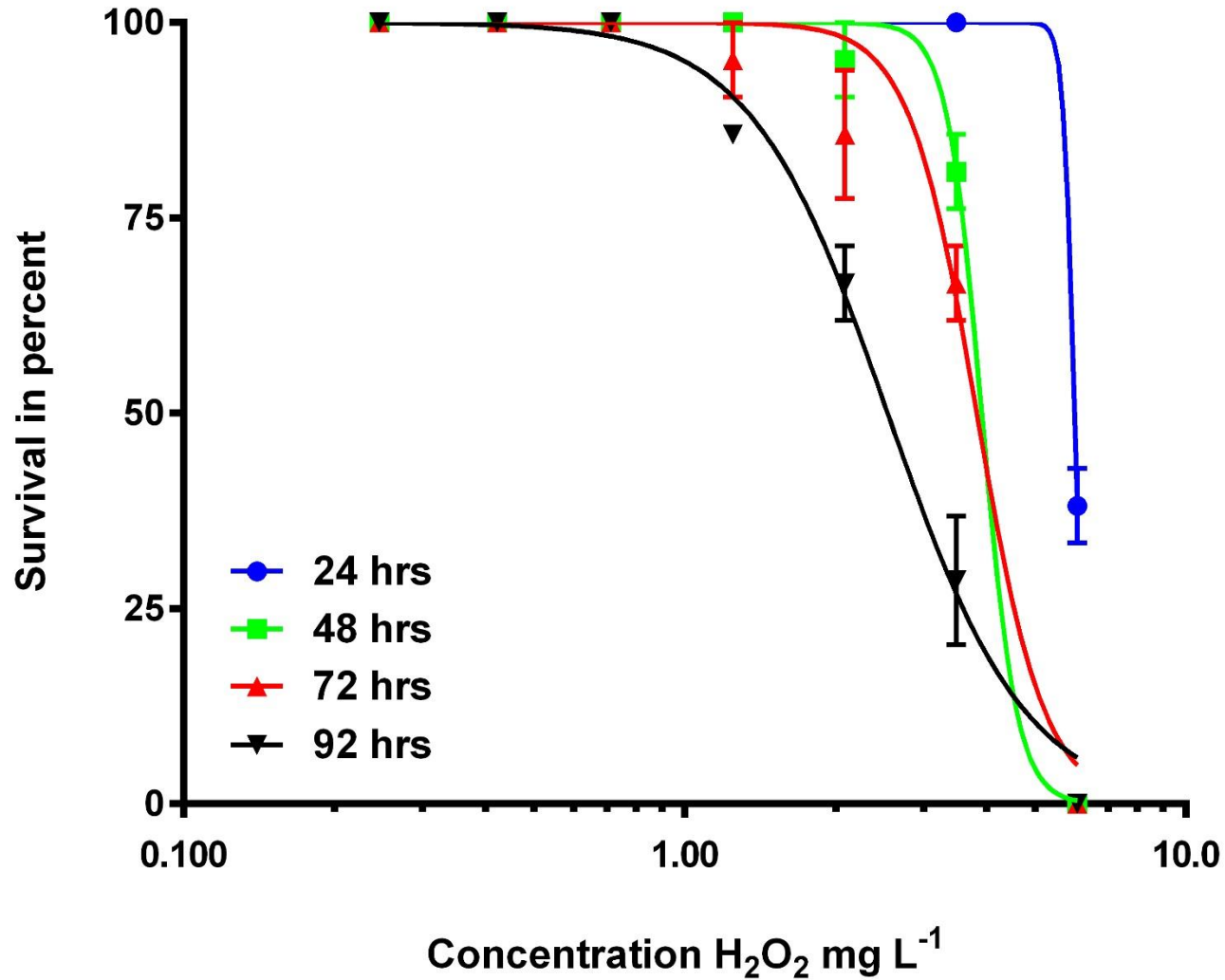


Figure 1: Survival in percent of *Calanus finmarchicus* plotted as a function of exposure concentration of H_2O_2 . Solid lines represent the data fitted to the non-linear sigmoidal model and scored data for survival at 24, 48, 72 and 96 hours of exposure are plotted as average (\pm SE).

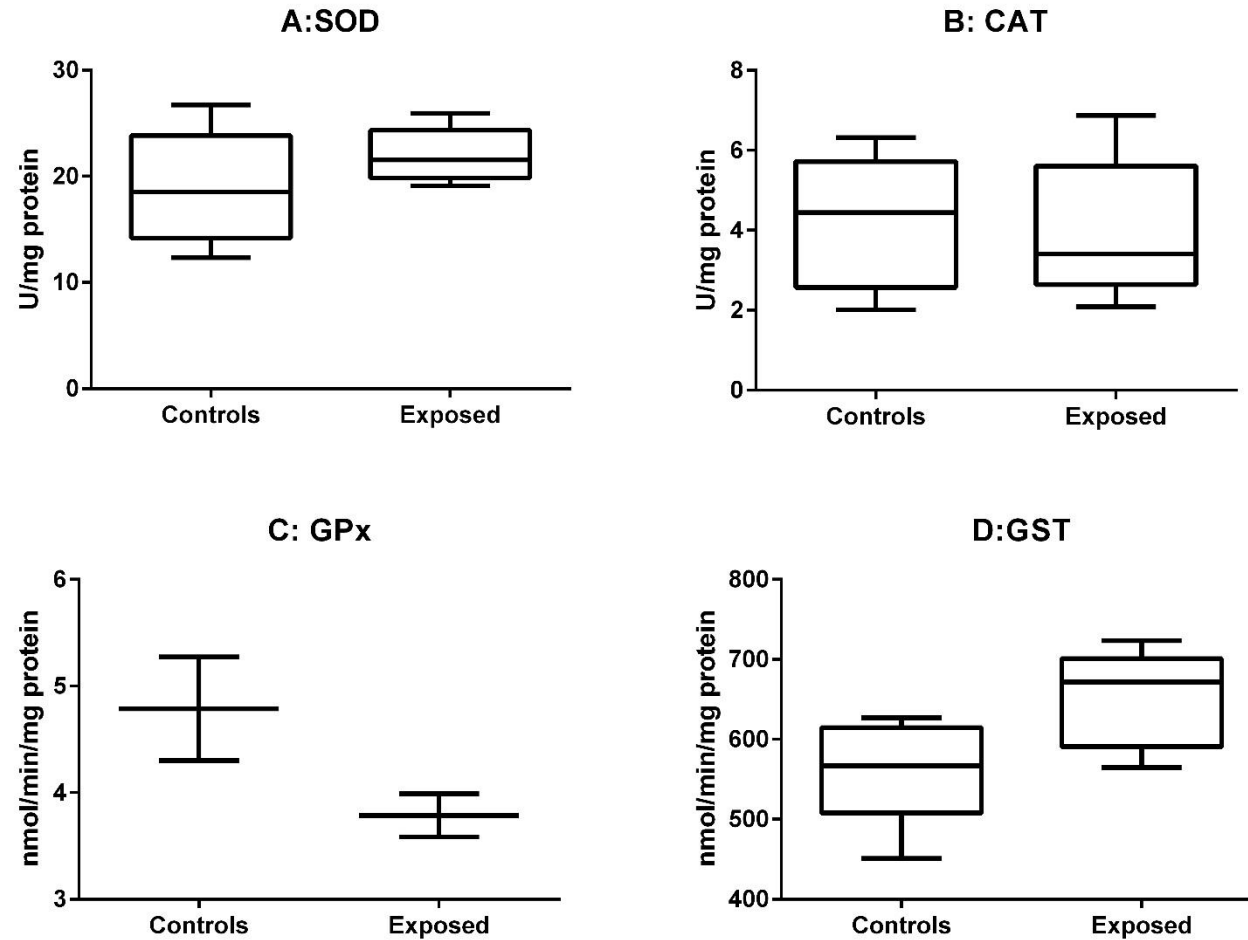


Figure 2: Enzyme activities in *Calanus finmarchicus* treated to clean seawater (negative control) and 0.75 mg H₂O₂ L⁻¹ for 96 hours. The data are displayed as whiskers (N=5 throughout except for GPx, where N=2). A: Superoxide dismutase activity (in U mg⁻¹). B: Catalase activity (in U mg⁻¹ protein). C: Glutathione peroxidase activity (in nmol/min/mg protein). Activity was only analyzed for two controls and two exposed samples for this enzyme. D: Glutathione S-transferase activity (in nmol/min/mg protein).

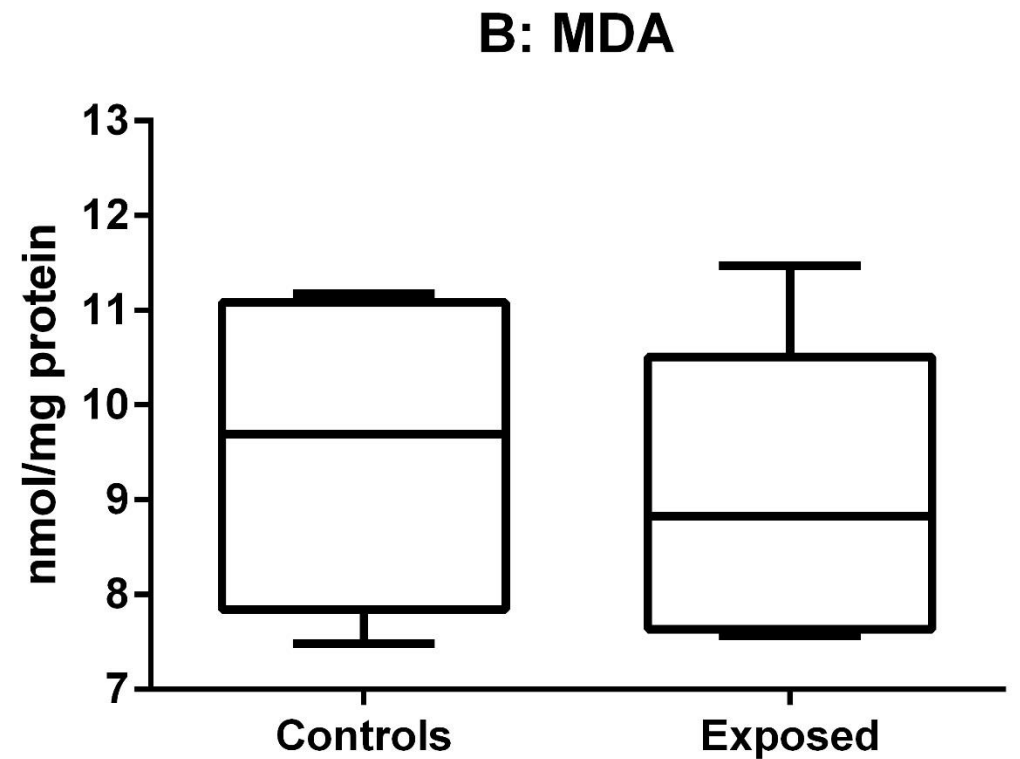
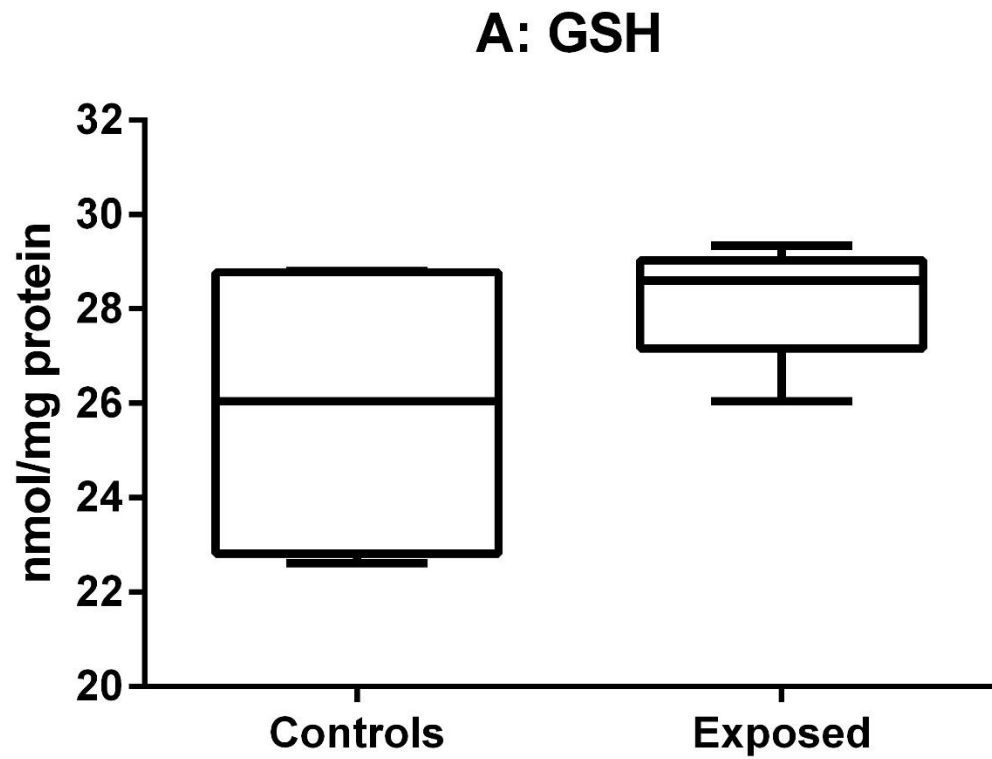


Figure 3: A: Concentrations of glutathione and MDA in *Calanus finmarchicus* treated with clean seawater (negative control group) and with 0.75 mg H₂O₂ L⁻¹ (exposed group) for 96 hours. The data are displayed as whiskers (N=5 throughout).

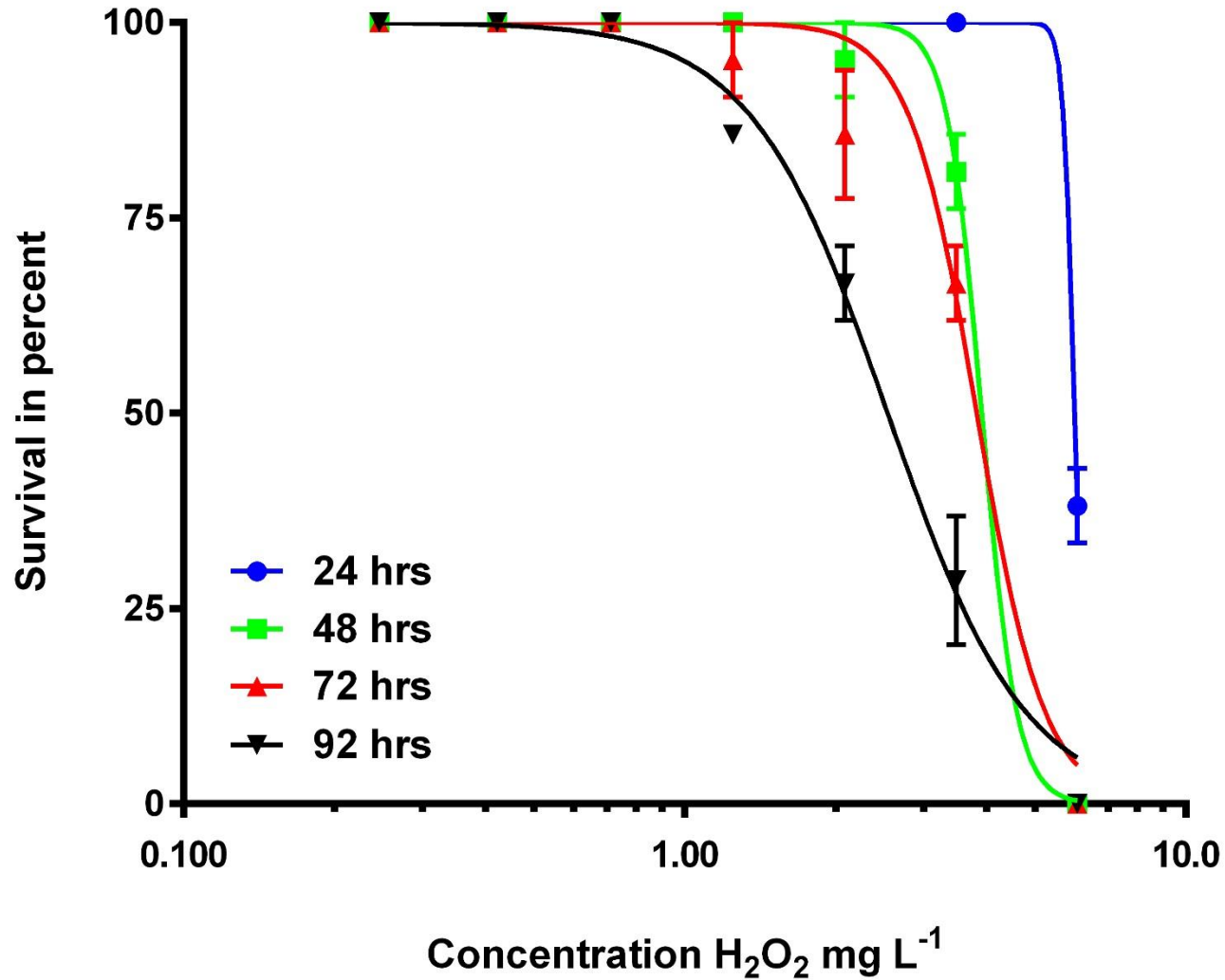


Figure 1: Survival in percent of *Calanus finmarchicus* plotted as a function of exposure concentration of H_2O_2 . Solid lines represent the data fitted to the non-linear sigmoidal model and scored data for survival at 24, 48, 72 and 96 hours of exposure are plotted as average (\pm SE).

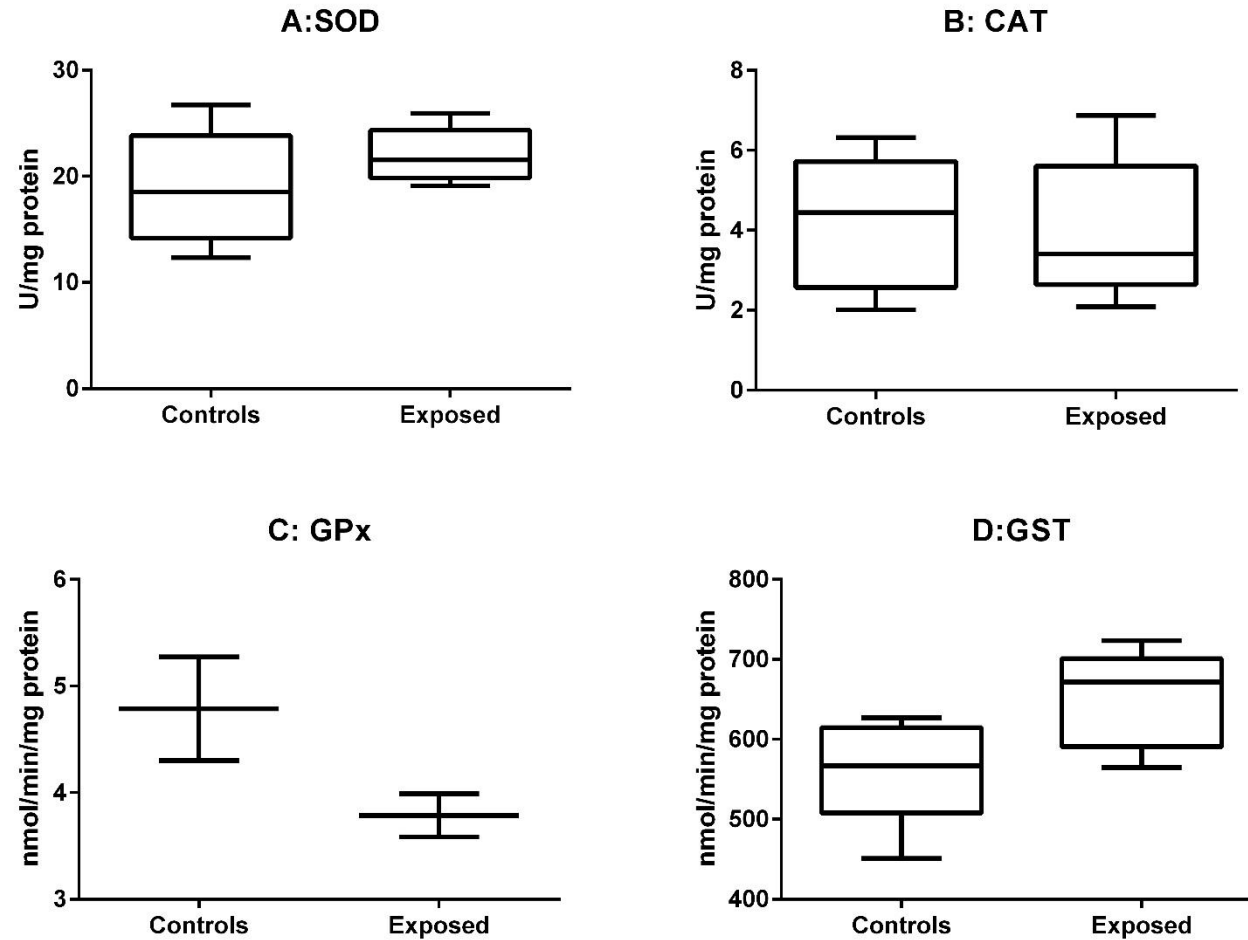


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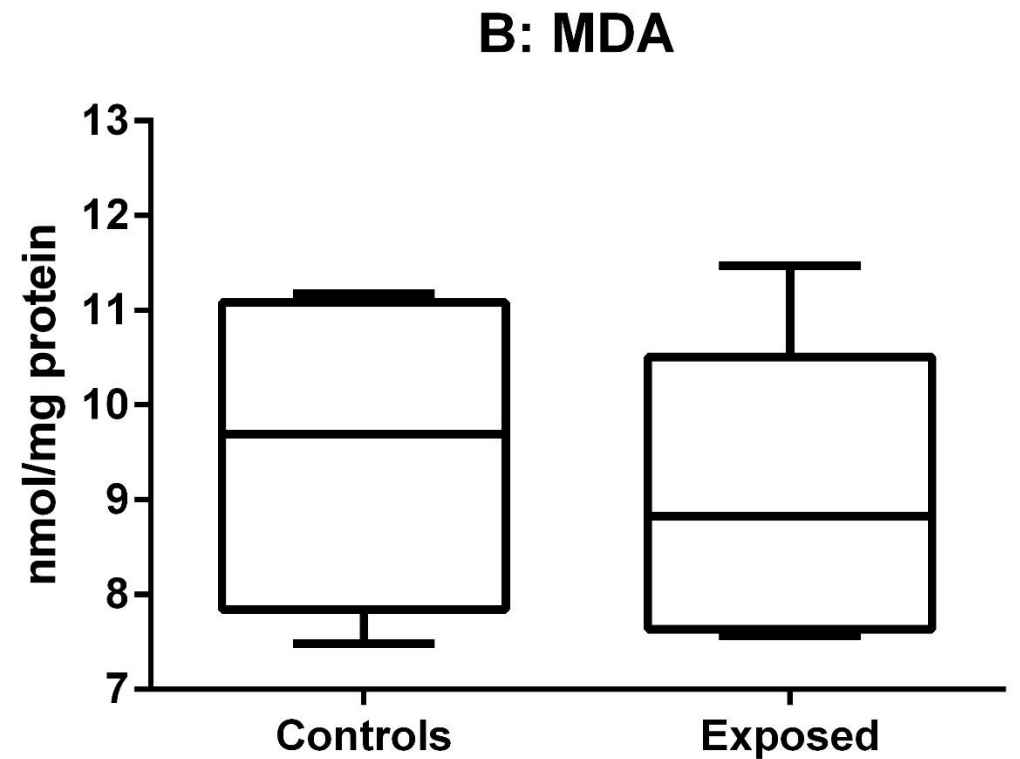
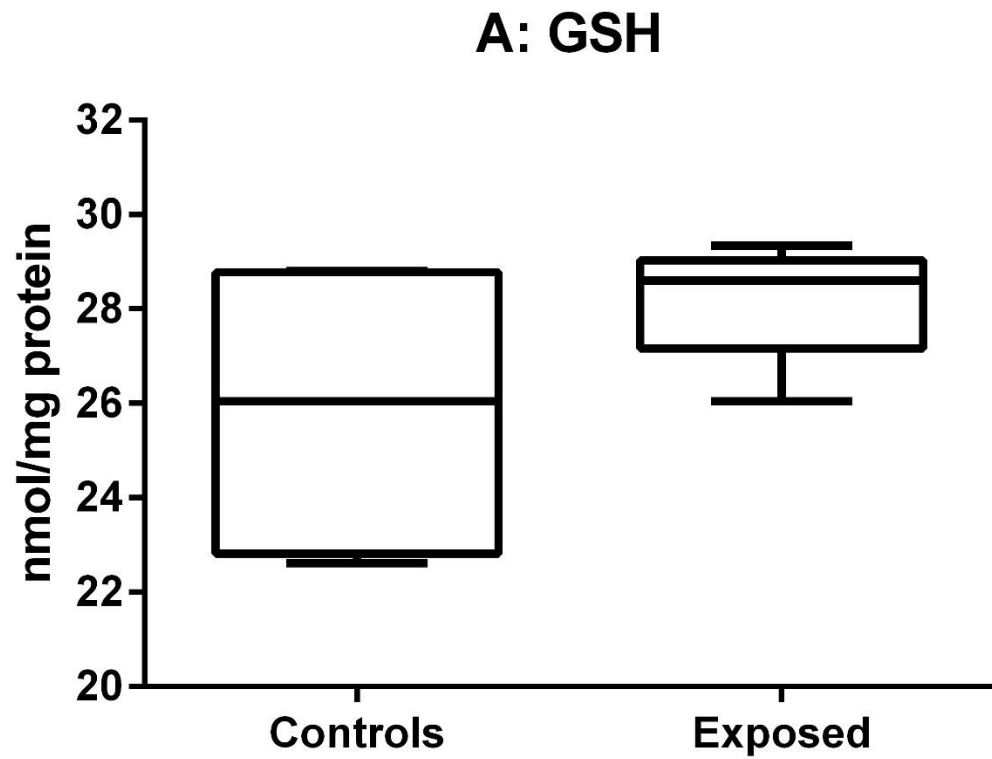


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